

Mesozooplankton Grazing on Picocyanobacteria in the Baltic Sea as Inferred from Molecular Diet Analysis

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Abstract

Our current knowledge on the microbial component of zooplankton diet is limited, and it is generally assumed that bacteria-sized prey is not directly consumed by most mesozooplankton grazers in the marine food webs. We questioned this assumption and conducted field and laboratory studies to examine picocyanobacteria contribution to the diets of Baltic Sea zooplankton, including copepods. First, qPCR targeting ITS-1 rDNA sequence of the picocyanobacteria *Synechococcus* spp. was used to examine picocyanobacterial DNA occurrence in the guts of Baltic zooplankton (copepods, cladocerans and rotifers). All field-collected zooplankton were found to consume picocyanobacteria in substantial quantities. In terms of *Synechococcus* quantity, the individual gut content was highest in cladocerans, whereas biomass-specific gut content was highest in rotifers and copepod nauplii. Moreover, the gut content in copepods was positively related to the picocyanobacteria abundance and negatively to the total phytoplankton abundance in the water column at the time of sampling. This indicates that increased availability of picocyanobacteria resulted in the increased intake of this prey and that copepods may rely more on picoplankton when food in the preferred size range declines. Second, a feeding experiment with a laboratory reared copepod *Acartia tonsa* fed a mixture of the picocyanobacterium *Synechococcus bacillaris* and microalga *Rhodomonas salina* confirmed that copepods ingested *Synechococcus*, even when the alternative food was plentiful. Finally, palatability of the picocyanobacteria for *A. tonsa* was demonstrated using uptake of ¹³C by the copepods as a proxy for carbon uptake in feeding experiment with ¹³C-labeled *S. bacillaris*. These findings suggest that, if abundant, picoplankton may become an important component of mesozooplankton diet, which needs to be accounted for in food web models and productivity assessments.

Citation: Motwani NH, Gorokhova E (2013) Mesozooplankton Grazing on Picocyanobacteria in the Baltic Sea as Inferred from Molecular Diet Analysis. PLoS ONE 8(11): e79230. doi:10.1371/journal.pone.0079230

Editor: Erik Sotka, College of Charleston, United States of America

Received: July 9, 2013; **Accepted:** September 27, 2013; **Published:** November 18, 2013

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Funding: Financial support was received from The Swedish Research Council for Environment, Agricultural Sciences and Spatial Planning (FORMAS) and Stockholm University's strategic marine environmental research program "Baltic Ecosystem Adaptive Management". The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

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Introduction

The smallest photosynthetic organisms include autotrophic picoplankton, a diverse group united by size $<2\ \mu\text{m}$. This group contributes as much as 40% of global ocean primary productivity and is mainly composed by picocyanobacteria [1,2]. In marine environments, picocyanobacteria encompassing diverse strains are represented by the genera *Synechococcus* and *Prochlorococcus* [1–3], with the former being the major contributor to the total photosynthetic biomass in the temperate oceans [1,4]. Similar to other marine areas, *Synechococcus*-type strains dominate Baltic Sea picocyanobacteria [5] that contribute up to 50% of total phytoplankton biomass [6] and up to ~70% of total chl *a* during summer [7,8] in offshore Baltic Sea waters.

Although much of biomass and primary production, particularly in low productive systems is due to the picoplankton, this phytoplankton fraction is considered largely unavailable for most metazooplankton, with heterotrophic nanoflagellates and ciliates being the major grazers on pico-sized prey [2,9]. Among metazooplankton, appendicularians [10], cladocerans [11,12], rotifers [12], and bivalve larvae [13] are known to substantially feed on picoplankton, but not copepods. These most important

grazers in marine systems do not feed efficiently on particles of this size as shown by feeding experiments with algal cultures [14] and natural phytoplankton assemblages [15]. The size of the smallest algae that a filtrator can capture is a function of the distance between the setules on the filtering appendages, whereas the maximum size of ingestible particles is generally determined by the grazer body size [16]. Colony-building picoplankton can easily be grazed by crustacean zooplankton [2,9], while single-celled species $<2\ \mu\text{m}$ would be too small to be efficiently retained by most of mesozooplankton filtrators [14–17]. Therefore, mesozooplankton grazing on picoplankton is generally considered to be non-efficient or intermittent [16,17]. Nevertheless, it has been suggested that "picocyanobacteria are in a size range suitable for utilization by nauplii and early copepodite stages as well as rotifers" [2], and some field and experimental studies indicate that ingestion of unicellular picoautotrophs by copepod species does occur [17,18]. In many systems, under food limiting conditions, feeding on picoplankton would be an advantage for grazers. It has been hypothesized that at low concentrations of phytoplankton, zooplankton grazers reduce their energy expenditure or even stop feeding [19]. Also, at low concentrations of preferred phytoplank-

ton species, zooplankton may switch to more abundant suboptimal prey [20].

Most of what we know about prey size selectivity in zooplankton is derived from feeding experiments that involve bottle incubations and analysis of the prey disappearing from the media, gut fluorescence measurements of grazers, and radioactive labeling [16]. Of these approaches, only gut fluorescence method is applicable for field studies and has been used to detect and quantify picocyanobacteria in copepod gut contents [18]. Grazing on picoplankton has also been studied using other techniques, such as fluorescent labeled cells, metabolic inhibitors, dilution technique, flow cytometry and radioisotope-labeled prey [9]. General pitfalls associated with these methodologies are the “bottle effects” and improper controls which do not correct for nutrient regeneration by zooplankton, resulting in underestimation of grazing rate and misinterpretation of selective feeding [21,22]. Moreover, many of the early isotope studies were not accurate because of recycling of the isotopes as a result of the excretion and respiration by phytoplankton and zooplankton [23]. There are also sources of error involved with gut fluorescence measurements and calculations of both ingestion and filtration rates that, particularly with fast growing picoplankton, can lead to underestimation of the grazing impact due to breakdown of pigment during digestive activity [24].

Currently, studies on trophic relationships are rapidly turning to DNA-based techniques [25]. Molecular methods based on quantitative PCR (qPCR) that can both identify prey of interest and quantify its contribution to the stomach content have been recently applied for diet analysis in zooplankton, including copepods [26–28]. The approach is particularly relevant for detecting prey groups with variable morphological characters and pigment composition, such as picocyanobacteria [29]. In qPCR-based diet analysis, there are a few molecules that are particularly informative for target identification and quantification (e.g., 16S rDNA, cytochrome *c*, and nuclear ribosomal genes and their spacers) [25]. The high abundance of these genes makes them an attractive target in molecular diet analysis [26–28,30], but also adds difficulty to account for copy number variability per cell in response to environmental conditions and strain composition in wild populations [31]. Another technical difficulty inherent to molecular diet studies on microscopic aquatic animals is to control for non-ingestion contamination by the target prey [27] that can be present in carry over water and/or adhere to body surfaces of the animal.

In line with the current views on mesozooplankton ability to graze on picoplankton, we hypothesized that in the Baltic plankton communities, picocyanobacteria are consumed mostly by nauplii, rotifers and cladocerans, but not by larger copepodites [2]. To test this hypothesis, we used molecular diet analysis based on qPCR targeting ITS-1 sequence of the picocyanobacteria *Synechococcus*. Using this technique, we quantified picocyanobacterial DNA in the guts of different zooplankters (rotifers, cladocerans and dominant copepod species at various developmental stages) collected during the growth season in a coastal area of the northern Baltic proper. Furthermore, the amount of picocyanobacteria in the guts was related to the ambient *Synechococcus* spp. and phytoplankton abundances. We also conducted feeding experiments with laboratory reared copepod *Acartia tonsa* fed picocyanobacterium *Synechococcus bacillaris* and cryptophyte *Rhodomonas salina* to (1) test whether copepods ingest picocyanobacteria in the absence of protozoan grazers; (2) determine the non-consumptive contribution of picocyanobacteria to zooplankton samples, due to adherence to body surfaces and other sources of contamination; and (3) quantify carbon uptake from the

picocyanobacteria in the copepods using ^{13}C -labeled *S. bacillaris* as prey.

Materials and Methods

Ethics Statement

The sampling was conducted within national Swedish monitoring in the coastal waters of Sweden and no specific permissions were required for the sampling locations of this study. Also, we did not require ethical approval to conduct this study as we did not handle or collect animals considered in any animal welfare regulations and no endangered or protected species were involved in the samplings or the experiments.

Field zooplankton collections

Zooplankton samples were collected in the Himmerfjärden Bay, a coastal area of the northern Baltic proper (59°00' N; 17°43' E, bottom depth ~28 m). Samples were collected around noon, bi-weekly, July to September 2008, by vertical bottom to surface tows using a 90 μm WP-2 net (diameter 57 cm). From each tow, randomly selected zooplankton were preserved in bulk using RNAlater and stored at -20°C for ~2 years [32]. From these samples, different species and developmental stages of mesozooplankton were picked under a dissecting microscope with forceps, rinsed in artificial seawater, and transferred in groups (7–10 ind sample $^{-1}$ for crustacean zooplankton and 12–25 ind sample $^{-1}$ for rotifers) into 1.5 ml Eppendorf tubes for DNA extraction. The following species/groups were selected for the analysis: (1) copepodites (CII–VI) of *Acartia* spp. and *Eurytemora affinis*, (2) cladoceran *Bosmina maritima*, (3) podonids (mixed samples for *Podon intermedius* and *P. leuckartii*), (4) copepod nauplii (stages N1–N6; mixed samples for *Acartia* spp. and *E. affinis*), and (5) rotifers (mixed samples for *Synchaeta* spp., *Keratella quadrata*, and *K. cochlearis*). To prepare reference samples (contamination control), freshly hatched *Artemia* spp. nauplii (San Francisco Bay Brand; 10 ind sample $^{-1}$) were used and treated in the same way as the zooplankton samples.

Synechococcus in the water column

Phytoplankton were collected with a plastic hose (inner diameter 19 mm) as integrated water samples (0–14 m) on the same occasions as zooplankton. The samples were immediately pre-filtered with a 35 μm sieve to remove large plankton and 100–250 ml of the filtrate were concentrated onto a 0.2 μm nylon membrane (47 mm diameter; MilliporeTM). The filters were folded, transferred in the 1.5 ml Eppendorf tubes, and stored at -80°C until further analysis.

DNA extraction

Zooplankton samples were incubated in 40 μl of 10% Instagene Chelex (Bio-Rad) for 30 min at 105°C [33]. After centrifugation (12 000 $\times g$, 5 min), the supernatant (30 μl) was transferred to a clean Eppendorf tube and stored at 4°C for 1–2 days. To extract DNA from the filters with phytoplankton assemblages, four sections from each filter were excised with 7 mm diameter punch and disrupted using Fast Prep[®] instrument and glass beads (<106 μm , Sigma-Aldrich) for 40 s. DNA was subsequently extracted with 400 μl of 10% Instagene Chelex-100 as described above, intermittently mixing the tube manually. The DNA measurement and quantification of *Synechococcus* spp. by qPCR were conducted using the same protocol as for the zooplankton samples. Concentrations of DNA (7.5–110 ng μl^{-1}) and purity were determined with a NanophotometerTM (Implen); A_{260}/A_{280} varied from 1.8 to 1.9.

Quantification of *Synechococcus* in zooplankton guts and water samples

To quantify *Synechococcus* spp. in zooplankton samples and plankton assemblages collected on the filters, a qPCR assay was applied using universal primers specific for *Synechococcus* (P100A: 5' ggt tta gct cag ttg gta gag cgc 3'; P3: 5' ttg gat gga ggt tag cgg act 3') and hydrolysis probe (S100A: 5' FAM- ctt tgc aag cag gat gtc agc ggt t- TAMRA 3') targeting the ITS-1 sequence spanning between 16S rDNA and 23S rDNA genes in the ribosomal operon of *Synechococcus* spp. [34]. These primers and probe have been broadly tested for their ability to amplify different *Synechococcus* strains from five different lineages using a 16S rRNA inferred phylogenetic tree; the strains were isolated from various fresh and brackish waters, including the Baltic Sea [34–36]. A synthetic DNA oligonucleotide (Invitrogen Ltd.) comprising 75 bp of the target sequence (*Synechococcus* sp. BS20: positions 1868–1942; 5' ggt tta gct gta ttg gta gag cgc ctt tgc aag cag gat gtc agc ggt tgc agt ccg cta acc tcc atc caa 3') was used as a standard [37]. Standard curves were generated using a five step 10-fold dilution series, 1.1×10^3 – 1.1×10^7 target amplicons per reaction; triplicate no template controls (NTC) were included in all runs. Reactions were performed in triplicate using the TaqMan Gene Expression Master Mix (Applied Biosystems) and a StepOne real-time cyler (Applied Biosystems). Amplifications were performed in a 20 μ l reaction mixture as follows: 2 minutes at 50°C, 10 minutes at 95°C, and 40 cycles of 15 seconds at 95°C and 1 minute at 60°C. Threshold cycle (C_t) was set automatically by StepOne Software 2.0. For each standard curve, the r^2 value, the amplification efficiency ($E\%$) and the y -intercept value were recorded. Coefficient of variation was used to estimate intra- and inter-assay variability [38]; see Table S1 for statistical evaluation of efficiency and repeatability of the qPCR assays. PCR products were also visualized in the GelDoc after electrophoresis at 90 V for 1 h on 2.5% agarose gel prepared in 1 \times TAE buffer containing 0.5 mg ml⁻¹ ethidium bromide.

In the test samples, *Synechococcus* spp. ITS-1 copy number was estimated from the standard curves. The molecular weight of the standard was used to calculate ITS-1 gene copy number per reaction:

$$N_{ITS-1} = \frac{A \times DNA}{MW},$$

where; N_{ITS-1} is number of copies (μ l⁻¹), A is 6×10^{23} is the gene copies mol⁻¹ (the Avogadro constant), DNA is DNA concentration ($g \mu$ l⁻¹), and MW —molecular weight of the amplicon, 46228 g mol⁻¹ [39].

Total phytoplankton

Sampling and analysis of phytoplankton were conducted as a part of Swedish National Monitoring Programme, following HELCOM guidelines [40]. Briefly, samples were settled in Utermöhl chamber and examined using a NIKON inverted microscope with phase contrast. Phytoplankton (>2 μ m; ≥ 500 cells) were counted in diagonals or on the half/whole chamber bottom at 150 \times and 400 \times magnification; cell volume was calculated from size measurements (≥ 25 cells species⁻¹).

Non-ingestion contamination by picocyanobacteria (Experiments I and II)

To determine the amount of *Synechococcus* that might have been attached to external body parts of copepods, but not ingested, feeding experiments were conducted with the copepod *Acartia tonsa*

reared in the laboratory and axenic cultures of *Synechococcus bacillaris* (CCAP 1479; cell size: 2 μ m) and *Rhodomonas salina* (strain CCAP 978/24; cell size: 8 μ m) as food; the latter alga is a high quality food commonly used in experiments with *Acartia* [27]. The picocyanobacterial and algal concentrations (cells ml⁻¹) were determined using a haemocytometer and converted to carbon mass [41]. As *Synechococcus* has been reported to build colonies and aggregates [42], the cultures were pre-filtered using 20 μ m sieve and the number of cells associated in microcolonies was noted (mean \pm SD: $1.8 \pm 0.3\%$; $n = 5$) when determining cell concentrations. To relate cell number to copy number of ITS-1 in the standard, DNA was extracted from 200 μ l of *S. bacillaris* culture with known cell density using Chelex method and analyzed by qPCR in the same way as the copepod samples.

Older copepodites (CIV–V; thereafter referred to as adults) and nauplii were used as test animals. The adults were picked using a wide mouth pipette and incubated in artificial seawater (7 PSU, 18°C) for 8 h without food. To obtain nauplii, eggs were collected from the batch cultures and incubated in 96-well microplate in the sterile artificial seawater. Starved adults (experiment I) and nauplii (experiment II) were randomly assigned to two treatments: (1) dead controls (22–30 and 10 ind sample⁻¹ for adults and nauplii, respectively); newly hatched nauplii (non-feeding stage) were used in this treatment to ensure empty guts; and (2) fed adults (13–15 ind. sample⁻¹) and nauplii at the first feeding stage NIII (10 ind. sample⁻¹). To prepare dead controls, all animals were killed by immersing in 95% ethanol prior to exposure to *Synechococcus* to prevent ingestion. In all experiments and treatments, copepod groups were placed in 50 ml chambers with false bottoms (mesh size 60 and 20 μ m for adults and nauplii, respectively) to prevent ingestion of fecal pellets by live copepods, and exposed to a mixture (1:6 by carbon content; 0.25 mg C l⁻¹) of *Synechococcus* (1.5×10^5 cells ml⁻¹) and *Rhodomonas* sp. (7.4×10^5 cells ml⁻¹); these prey densities were selected to approximate summer phytoplankton community in terms of the proportion between the picoplankton and larger phytoplankton fractions in the Baltic Sea [6]. The exposure lasted 3 h; this time was considered sufficient for the copepods to recover from handling, start feeding normally and fully fill their guts. Upon termination of the experiment, the copepods were collected on the 20 μ m sieve, washed twice in 7 PSU artificial sea water and transferred with forceps to Eppendorf tubes with RNA $later$. The samples were then processed in the same way as the field samples.

Carbon uptake by copepods fed picocyanobacteria (Experiment III)

The uptake of carbon from picocyanobacteria by copepods was measured using *S. bacillaris* labeled with ¹³C and fed to *A. tonsa* copepodites. The ¹³C-labeled *Synechococcus* was prepared by replacing NaH¹²CO₃ with NaH¹³CO₃ in the f/2 medium that was done by adding 2 ml of a NaH¹³CO₃ stock solution (336 mg NaH¹³CO₃ in 100 ml H₂O sodium bicarbonate, ¹³C, 99%, Cambridge Isotope Laboratories) per 100 ml of the medium and growing the culture for 4 days in a climate room at 18–20°C at constant illumination. The labeling resulted in isotope signatures ($\delta^{13}C$) of -13.6% and 6683.4‰ for untreated and ¹³C enriched cultures, respectively. To measure the ¹³C uptake by the copepods, 25–30 copepodites (CIII–CIV) of *A. tonsa* were assigned to two treatments, each in three replicates: (1) dead controls that were prepared as described in experiments I and II and incubated with the ¹³C-labeled picocyanobacteria for 4 h, and (2) fed copepods that were incubated for 96 h. In addition, animals sampled at time 0 were used to measure the carbon signature before feeding on the enriched material. At the end of the incubation, the copepods were

rinsed with excess of Milli-Q water, and live copepods were incubated with unlabelled picocyanobacteria for another 4 h to ensure replacing of the ^{13}C -labeled food in the guts. All copepods were transferred into pre-weighed tin capsules (25 ind. sample $^{-1}$) and dried at 60°C for 24 h. The $\delta^{13}\text{C}$ values in the copepods were used as a proxy for carbon uptake; these values were measured with a continuous flow isotope ratio mass spectrometer (Europa Integra) at the UC Davis Stable Isotope Facility (University of California, USA).

Statistical analysis

Gut content (GC) in terms of *Synechococcus* ITS-1 copies ind. $^{-1}$ and size specific GC (ssGC; copies μgWW^{-1} , where WW is zooplankton wet weight [43]) in the field-collected animals were compared among zooplankton groups and species by unpaired *t*-test with Welch's correction for unequal variances (GraphPad Prism 5.0®, GraphPad Software). The $\delta^{13}\text{C}$ values of the copepods in the experiment III were compared among the treatments (i.e., start animals, dead controls and fed copepods) using one-way ANOVA followed by *a posteriori* comparisons with the Tukey HSD test. To evaluate effects of *Synechococcus* and total phytoplankton abundance on GC in the copepods, generalized linear model (GLM) with normal distribution and log-link (Statistica v. 10, StatSoft Inc.) and pooled data for *Acartia* spp. and *E. affinis* were used. The regression analysis was limited to the copepods, because the GC data for this group were available on most sampling occasions. Data were Box-Cox transformed and the residuals were linear, homogenous, normally distributed and not correlated.

Results

Presence and abundance of *Synechococcus* DNA in mesozooplankton

All field-collected zooplankton samples tested positively for *Synechococcus* DNA (Table 1), whereas no amplification was observed in the reference samples (newly hatched *Artemia*). The amount of ITS-1 copies varied about 7-fold (8×10^3 to 53.8×10^3 per zooplankton), with substantial differences between the species and groups (Table 1, Figure 1). The differences in GC between main zooplankton groups were significant: copepods vs. microzooplankton (unpaired *t*-test; $t_{17} = 2.150$, $p < 0.04$), cladocerans vs. microzooplankton ($t_7 = 3.891$, $p < 0.006$), and copepods vs. cladocerans ($t_6 = 2.403$, $p < 0.05$). For ssGC, a different pattern was observed, with the values decreasing with the body size and differing significantly between the zooplankton groups: copepods vs. microzooplankton ($t_{17} = 12.507$, $p < 0.0001$), copepods vs. cladocerans ($t_6 = 5.60$, $p < 0.0014$), and cladocerans vs. microzooplankton ($t_7 = 3.402$, $p < 0.0145$; Figure 1).

Changes in total phytoplankton and picocyanobacteria during the season

During July–September 2008, the ambient *Synechococcus* spp. abundance in terms of the number of ITS-1 copies varied from 2.2×10^5 – 7.6×10^5 copies ml^{-1} , with the highest values observed in August (monthly average 4.9×10^5 copies ml^{-1}). Total phytoplankton biovolume ranged from 0.4 to 1.5 $\text{mm}^3 \text{ml}^{-1}$, with the peak observed at the end of the August ($1.5 \text{mm}^3 \text{ml}^{-1}$).

Relationship between picocyanobacteria intake and their abundance

There were no significant differences in the GC between the copepodites of *E. affinis* and *Acartia* spp. on each sampling occasion ($t_4 = 1.70$, $p > 0.05$). Hence, these species were pooled for GLM

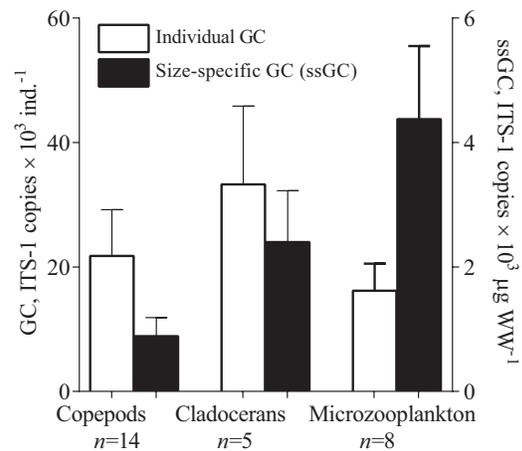


Figure 1. Occurrence of *Synechococcus* spp. in field-collected zooplankton. Individual gut content (GC; prey ITS-1 copies $\times 10^3 \text{ ind}^{-1}$) and size-specific gut content (ssGC; prey ITS-1 copies $\times 10^3 \mu\text{gWW}^{-1}$) in main zooplankton groups: copepods (adults and older copepodites of *Acartia* spp. and *Eurytemora affinis*), cladocerans (*Bosmina maritima* and *Podon* spp.) and microzooplankton (rotifers *Synchaeta* spp., *Keratella quadrata*, and *K. cochlearis*, and copepod nauplii). Data are shown as mean \pm SD, number of samples is given below the group name.

doi:10.1371/journal.pone.0079230.g001

relating individual GC to *Synechococcus* abundance (ITS-1 copies ml^{-1}) and total phytoplankton (biovolume, $\text{mm}^3 \text{ml}^{-1}$). In this model, the amount of *Synechococcus* DNA in the copepod gut was positively related to the picocyanobacteria abundance and negatively to the total phytoplankton stocks at the time of sampling (Table 2).

Experiments I and II

Synechococcus DNA were detected in both killed and live copepods exposed to the experimental feeding media (Figure 2), with values being ~ 5 fold higher in the live copepods (adults: $t_4 = 32.61$, $p < 0.0009$; nauplii: $t_4 = 32.73$, $p < 0.0001$). The percentage of *Synechococcus* measured in the dead individuals compared to the live animals of the same developmental stage was similar between the adults and nauplii, 21.8% and 21.2%, respectively. The copy number of ITS-1 per cell in the *Synechococcus* culture was 2.04 ± 0.03 as estimated by qPCR analysis of samples with known cell abundance (7.2×10^7 cells ml^{-1}). Thus, non-ingestion background corresponded to about 2200 and 965 *Synechococcus* cells ind^{-1} for adults and nauplii of *Acartia tonsa*, respectively.

Experiment III

There was a significant carbon uptake by *A. tonsa* copepodites fed *Synechococcus* (ANOVA; $F = 556$, $p < 0.0001$; Figure 3). Moreover, measurable increase was also found in dead controls, albeit this increase was ~ 10 -fold lower than in the fed copepods (Tukey HSD, $q = 4.305$, $p < 0.05$).

Discussion

Contrary to the generally accepted view that mesozooplankton are inefficient at capturing pico-sized particles, such as autotrophic picoplankton [9], all tested species of Baltic zooplankton, including copepods of all stages, were found to directly consume substantial quantities of picocyanobacteria. Two crucial results support this conclusion. First, qPCR-based diet analysis revealed presence of

Table 1. *Synechococcus* abundance (ITS-1 copies $\times 10^3$ ind $^{-1}$) detected in different mesozooplankton species/groups.

| Species/group | Category | <i>Synechococcus</i> mean (min–max) | n |
|---|------------------|-------------------------------------|---|
| <i>E. affinis</i> , adults | copepods | 21 (16–31) | 7 |
| <i>Acartia</i> spp., adults | copepods | 23 (11–36) | 7 |
| <i>Acartia</i> spp., <i>E. affinis</i> , copepodites (CII–VI) | copepods | 21 (17–24) | 2 |
| <i>Acartia</i> spp., <i>E. affinis</i> , nauplii (N1–N6) | microzooplankton | 18 (15–24) | 5 |
| <i>B. maritima</i> (0.4–0.6 mm) | cladocerans | 23 (20–27) | 2 |
| podonids | cladocerans | 35 (24–54) | 4 |
| rotifers | microzooplankton | 10 (8–12) | 3 |
| Nauplii and rotifers | microzooplankton | 8 | 1 |

Field-collected samples were used for the analysis. *E. affinis* – *Eurytemora affinis*, *Acartia* spp. – *Acartia bifilosa* and *A. longiremis*, *B. maritima* – *Bosmina maritima*, podonids – *P. intermedius* and *P. leuckartii*, rotifers – *Synchaeta* spp., *Keratella cochlearis* and *K. quadrata*; nauplii – *Acartia* spp. and *E. affinis*; n – number of samples analyzed.

doi:10.1371/journal.pone.0079230.t001

Synechococcus, the dominant picocyanobacteria in the Baltic Sea [5], in the guts of all major zooplankton groups, including rotifers, cladocerans and copepods. Second, the feeding experiments confirmed that both nauplii and adults of *Acartia tonsa* were ingesting and assimilating *Synechococcus bacillaris* even when alternative food was plentiful and no protozoan grazers were present. The latter implies that the picocyanobacteria occurrence in the field-collected zooplankton are primarily the result of the direct grazing on picocyanobacteria and not the secondary consumption, i.e., consumption of prey that had been feeding on the picocyanobacteria. Also, the amount of *Synechococcus* adhering to the outside of zooplankters and/or caused by possible contamination during the sorting procedure was ~20% of the total as indicated by the comparison of the picocyanobacteria abundance in the live and dead copepods exposed to *Synechococcus* in the feeding experiments. Although this indicates that gut content was the main source of the PCR-based estimates of picocyanobacteria abundance in the zooplankton samples, the non-ingestion background should be taken into consideration when analyzing zooplankton samples. The percentage of contamination was not affected by the size of the animals (adult and nauplii copepod stages), which allows applying the 0.2 correction factor for background contamination in the field samples. However, this value may depend on the ambient picocyanobacteria abundance in the water, which should be further investigated in similarly designed experiments with varying picocyanobacteria densities in the media.

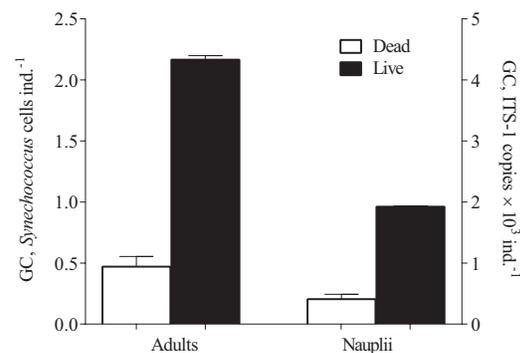
Table 2. Statistical summary of the generalized linear model examining effects of *Synechococcus* abundance (ITS-1 copies $\times 10^3$ ml $^{-1}$) and total phytoplankton (>2 μ m) biovolume (mm 3 ml $^{-1}$) in the water column (0–14 m) on the abundance of *Synechococcus* DNA in copepod stomachs (ITS-1 copies $\times 10^3$ ind $^{-1}$).

| | Estimate | Standard error | Wald statistic | p-value |
|----------------------|----------|----------------|----------------|--------------|
| Intercept | 0.243 | 0.332 | 0.535 | 0.464 |
| <i>Synechococcus</i> | 0.238 | 0.065 | 13.53 | 0.000 |
| Total phytoplankton | -0.474 | 0.136 | 12.03 | 0.001 |

Data are Box-Cox transformed, significant effects are in bold face.

doi:10.1371/journal.pone.0079230.t002

As hypothesized, cladocerans, rotifers and nauplii were found to have the highest biomass-specific amounts of picocyanobacteria in their guts. These zooplankters have been reported to feed efficiently on bacteria-sized particles, including picocyanobacteria (cladocerans: [44,45], rotifers [45] and nauplii [46]). However, contrary to our expectations and various feeding studies showing that older copepodites do not feed on picoplankton [14,15], guts of *Acartia* spp. and *Eurytemora affinis* copepodites exposed to picocyanobacteria contained $\sim 2 \times 10^4$ *Synechococcus* ITS-1 copies ind $^{-1}$ (Table 1). According to the filtration theory, single-celled organisms <2 μ m are outside of the size range of particles that copepods can retain on their feeding appendages [16,47]. In *Acartia tonsa*, for example, the retention efficiency drops dramatically below ~ 5 μ m particles [48,49]. What are then the mechanisms by which large copepods ingest relatively large quantities of *Synechococcus*? First, as mentioned above, although *Synechococcus* cells are basically solitary, they can build microcolonies with 2–50 cells colony $^{-1}$ [42] and/or occur in loose agglomerates, particularly in summer [50]. The presence of these colonies and agglomerates in the picocyanobacteria populations would greatly increase retention efficiency for *Synechococcus*. Second, autotrophic picoplankton occurs in aggregates with detrital particles and heterotrophic bacteria, which enhances their availability for mesozooplankton [18]. Finally, copepods feeding

**Figure 2.** Quantities of *Synechococcus bacillaris* (ITS-1 copies $\times 10^3$ ind $^{-1}$) and cells $\times 10^3$ ind $^{-1}$) detected in the live and dead individuals of the copepod *Acartia tonsa* (adults and nauplii) exposed to the picocyanobacterium in the feeding experiments (Experiments I and II). Data are shown as mean \pm SD, n = 3 in all cases.

doi:10.1371/journal.pone.0079230.g002

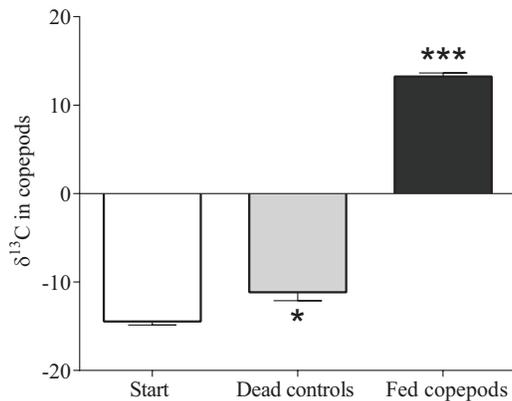


Figure 3. Carbon uptake from ¹³C-labeled *Synechococcus bacillaris* by the copepod *Acartia tonsa* (live and dead individuals) exposed to the picocyanobacterium (Experiment III). Carbon uptake is expressed as change in δ¹³C of the copepods from the start values. Differences between the start and each treatment group are shown by asterisks (*: $p < 0.05$; ***: $p < 0.0001$). Data are shown as mean \pm SD, $n = 3$ in all cases. doi:10.1371/journal.pone.0079230.g003

on small particles [51] can adjust the posture of their filtering appendages and beat their maxillae at a faster rate than the same animals feeding on large particles to obtain the same ration [47].

In our feeding experiments, adult *A. tonsa* exposed to *Synechococcus bacillaris* with a very few (<2% of the total number of cells) microcolonies, consumed the picocyanobacterium, albeit at lower quantities than *Acartia* spp. (*A. bifilosa* and *A. longiremis*) in the field ($\sim 5 \times 10^3$ and 21×10^3 *Synechococcus* ITS-1 copies ind^{-1} in the laboratory fed and field collected copepods, respectively). The higher abundance of picocyanobacteria cells in the guts of field-collected copepods could be related to (1) larger body size of *A. tonsa* compared to the other two *Acartia* species [43] and lower capacity for picoplankton ingestion, (2) greater picocyanobacteria aggregate formation in the field increasing retention rate, and (3) secondary consumption in the field, where feeding on protozooplankton feeding on picoplankton was likely to occur. The latter mechanism requires additional experimental studies to establish detection efficiency for DNA in the mesozooplankton guts for *Synechococcus* that underwent secondary consumption. These experimental data and the difference indicate that both solitary and grouped picoplankton cells can be ingested by the copepods and that all mechanisms outlined above may contribute to the observed variation in consumption of *Synechococcus* in the field-collected zooplankton.

Cell counts and qPCR analysis of *Synechococcus bacillaris* culture used in the feeding experiment revealed that *S. bacillaris* has two copies of ITS-1 gene per cell, which concurs with earlier observed two ribosomal operons per cell in four strains of *Synechococcus* from the Baltic Sea [36]. Assuming that in the study area picocyanobacteria populations consisted of several *Synechococcus* strains [52] and their average ITS-1 cell copy number equaled 2, we attempted to calculate the pigment-based contribution of *Synechococcus* to the gut content of the zooplankton. Using the GC data for *Synechococcus* spp. in terms of cells ind.^{-1} and conversion factors reported in the literature: cell carbon content of picocyanobacteria of $0.25 \text{ pg cell}^{-1}$ [41] and C:Chl *a* ratio for picocyanobacteria of 32:1 [53], we arrived at the pigment-based equivalent of *Synechococcus* in the zooplankton guts being in the range of 0.03–0.15 ng Chl *a* ind^{-1} . These values are comparable to the measured gut fluorescence 0.06–0.77 and 0.15–0.4 ng Chl *a*

ind^{-1} , reported for estuarine *E. affinis* [54] and *A. bifilosa* [54,55], respectively. Thus, the contribution of picocyanobacteria in the diet of these copepods may account for 8–35% and 10–47% of copepod total gut content, respectively. Whereas the highest values are probably an overestimation related to less than full guts in the copepods, the lower end of the range, i.e., 8–10% of the total gut content would represent a conservative estimate. These values are very close to the contribution of picoplankton ($\sim 10\%$) to the total carbon-based ingestion rate that has been observed in the copepods *Acartia clausi* and *Temora stylifera* in the feeding experiments using size-fractionated plankton assemblages [56]. Also, prokaryotes were found to contribute >50% to the gut content of the Baltic copepod *Limnocalanus macrurus* [57], which emphasized the possible importance of bacteria-sized particles to zooplankton diets. The comparison of *Synechococcus*-based GC with the gut pigment content measured in the Baltic cladocerans and rotifers, 0.05–1.10 and 0.05–0.37 ng Chl *a* ind^{-1} , respectively [54], implies possible contribution of the picocyanobacteria being as high as 15–33%.

In our study, no significant difference in grazing on picocyanobacteria was observed between the copepods *Acartia* spp. and *E. affinis*, although the feeding behavior of these two species has been reported to be substantially different. For example, *Acartia* spp. often feed on specific food rather than most available food [58]. By contrast, *E. affinis* have opportunistic feeding mechanism; consuming smaller prey to compensate for food limitation when its preferred food is less abundant [59]. The most probable explanation for the lack of the observed differences in *Synechococcus* amounts in the guts between these copepod species is that feeding strategies in copepods may vary with food abundance in the environment [58] and that both copepods can use picoplankton as an alternative food to compensate for low phytoplankton availability. Indeed, the observed variations in *Synechococcus* GC of copepods were negatively related to availability of phytoplankton (>2 μm) and positively to the picocyanobacteria abundance (Table 2). This suggests that both copepods may increase picocyanobacteria consumption when this prey is highly abundant and when there is a food limitation. In particular, occurrence of dense blooms of filamentous cyanobacteria during summer in the Baltic Sea with concomitant decrease of edible phytoplankton [6] can substantially worsen food availability for mesozooplankton and thus contribute to the increased consumption of *Synechococcus* spp. It is also possible that zooplankton would prefer picocyanobacteria to less edible food, such as toxic filamentous cyanobacteria, particularly in the light of our findings that *Synechococcus* is digested and assimilated by the copepods. Also, the evidence is accumulating that picocyanobacteria may be a valuable nitrogen source for grazers [60]. In the Baltic Sea, the microbial food web receives substantial amounts of fixed nitrogen by diazotrophic cyanobacteria and leaking out, thus fueling microbial production including nitrogen-limited picocyanobacteria during summer [61]. Therefore, ready access to a nitrogen source may be an important adaptive trade-off for zooplankton during periods of nitrogen limitation.

The observed feeding by mesozooplankton on picocyanobacteria has several important implications for our understanding of the marine planktonic food webs. A direct pathway of carbon transfer from picoautotrophs to metazooplankton implies a higher transfer efficiency from primary producers to primary consumers. It is commonly accepted that metazooplankton utilize bacterial carbon by preying on protozoans feeding on bacteria or ingesting detritus to which bacteria adhere [18]. However, most studies agree that picoplankton production is not efficiently transferred to metazooplankton because of the multiple trophic steps in the

microbial loop [62,63] and, consequently, energy and nutrient dynamics models [64] and budget calculations [65] have no direct bacteria → zooplankton pathway when copepods (with any demographic population structure) dominate the community. While the microbial loop pathway is by no doubt the major energy route from picoplankton to metazooplankton in most pelagic food webs, the direct grazing by metazooplankton on picoplankton and filamentous cyanobacteria [66] may contribute measurably to zooplankton growth, particularly during periods of high picoplankton abundance and poor availability of larger phytoplankton. Although digestibility of picocyanobacteria by metazooplankton has been questioned [2], our experiments showed that *Synechococcus* was not only ingested but also assimilated by the copepods. Therefore, grazing on picoplankton by crustacean zooplankton should be more appreciated in food web models and productivity assessments. This is further supported by studies on metabolic budgets for herbivorous zooplankton showing that their daily ingestion rates on phytoplankton are insufficient to balance their respiration needs, and consumption of bacteria-sized particles may be necessary to satisfy zooplankton energy requirements [67]. Finally, ecosystem response to environmental change and cyanobacterial blooms have been suggested to increase energy flow through the microbial loop, which would decrease energy transfer efficiency to the higher trophic levels. Therefore, in the systems, where zooplankton grazers are capable to directly utilize picocyanobacterial biomass, the energy transfer from the microbial loop to the top consumers might be close to that in the classical food chain. To conclude, our findings demonstrate an important trophic link between mesozooplankton, including copepods and picocyanobacteria represented by the globally important primary producer *Synechococcus* spp. The grazing on picocyanobacteria may be a common year-round phenomenon in

the Baltic Sea and, perhaps, in other aquatic environments, particularly during periods of a low food abundance. If metazooplankton grazers, particularly copepods, are capable to directly and efficiently utilize picoplankton, this would facilitate a direct energy transfer from microbial producers to metazooplankton, surpassing the microbial loop. Our results warrant a revision of current pelagic food web models linking phytoplankton to secondary production in the Baltic Sea as well as other systems where picoplankton contributes substantially to primary production.

Supporting Information

Table S1 Regression coefficient (r^2), amplification efficiency (E), y-intercept values of the standard curves and no template controls (NTC) generated on five analytical occasions using the synthetic oligonucleotide as a standard for ITS-1 of *Synechococcus* spp. (DOC)

Acknowledgments

We thank Josefin Sefbom (Göteborg University) and Helena Höglander (Stockholm University), for collecting samples that were used for molecular analysis of picoplankton in the water, Susanna Hajdu (Stockholm University) for providing the phytoplankton data, and Anna Edlund, Jacob Walve, Ulf Larsson and Ragnar Elmgren (Stockholm University) for fruitful discussions.

Author Contributions

Conceived and designed the experiments: EG. Performed the experiments: NM EG. Analyzed the data: NM EG. Contributed reagents/materials/analysis tools: EG. Wrote the paper: NM EG.

References

- Agawin NSR, Duarte CM, Agustí S (2000) Nutrient and temperature control of the contribution of picoplankton to total phytoplankton biomass and production. *Limnol Oceanogr* 45: 591–600.
- Stockner JG, Antia NJ (1986) Algal picoplankton from marine and freshwater ecosystems: A multidisciplinary perspective. *Can J Fish Aquat Sci* 43: 2472–2503.
- Scanlan DJ, Ostrowski M, Mazard S, Dufresne A, Garczarek L, et al. (2009) Ecological genomics of marine picocyanobacteria. *Microbiol Mol Biol Rev* 73: 249–299.
- Glover HE, Campbell L, Prezelin BB (1986) Contribution of *Synechococcus* spp. to size-fractionated primary productivity in the three water masses in the Northwest Atlantic. *Mar Biol* 91: 193–203.
- Sánchez-Baracaldo P, Handley BA, Hayes PK (2008) Picocyanobacterial community structure of freshwater lakes and the Baltic Sea revealed by phylogenetic analyses and clade-specific quantitative PCR. *Microbiology* 154: 3347–3357.
- Hajdu S, Höglander H, Larsson U (2007) Phytoplankton vertical distributions and composition in Baltic Sea cyanobacterial blooms. *Harmful Algae* 6: 189–205.
- Albertano P, Di Somma D, Capucci E (1997) Cyanobacterial picoplankton from the Central Baltic Sea: cell size classification by image-analyzed fluorescence microscopy. *J Plankton Res* 19: 1405–1416.
- Stal LJ, Staal M, Villbrandt M (1999) Nutrient control of cyanobacterial blooms in the Baltic Sea. *Aquat Microb Ecol* 18: 165–173.
- Callieri C, Stockner JG (2002) Freshwater autotrophic picoplankton: a review. *J Limnol* 61: 1–14.
- Gorsky G, Chrétiennot-Dinet MJ, Blanchot J, Palazzoli I (1999) Picoplankton and nanoplankton aggregation by appendicularians: Fecal pellet contents of *Megalocercus huxleyi* in the equatorial Pacific. *J Geophys Res: Oceans* 104: 3381–3390.
- Lipej L, Mozeti P, Turk V, Malej A (1997) The trophic role of the marine cladoceran *Penilia avirostris* in the Gulf of Trieste. *Hydrobiologia* 360: 197–203.
- Sanders RW, Porter KG, Bennet SJ, DeBiase AE (1989) Seasonal patterns of bacterivory by flagellates, ciliates, rotifers and cladocerans in a freshwater plankton community. *Limnol Oceanogr* 34: 673–687.
- Tomaru Y, Kawabata Z, Nakano S (2000) Consumption of picoplankton by the bivalve larvae of Japanese pearl oyster *Pinctada fucata martensii*. *Mar Ecol Prog Ser* 192: 195–202.
- Frost BW (1972) Effect of size and concentration of food particles on the feeding behavior of the marine planktonic copepod *Calanus finmarchicus*. *Limnol Oceanogr* 17: 805–815.
- O'Connors HB, Biggs DC Jr, Ninivaggi DV (1980) Particle-size dependent maximum grazing rates for *Temora longicornis* fed natural particle assemblages. *Mar Biol* 56: 65–70.
- Kjørboe T (2011) How zooplankton feed: mechanisms, traits and trade-offs. *Biol. Rev.* 86: 311–339.
- Pace ML, McManus GB, Findlay SEG (1990) Planktonic community structure determines the fate of bacterial production in a temperate lake. *Limnol Oceanogr* 35: 795–808.
- Wilson SE, Steinberg DK (2010) Autotrophic picoplankton in mesozooplankton guts: evidence of aggregate feeding in the mesopelagic zone and export of small phytoplankton. *Mar Ecol Prog Ser* 412: 11–27.
- Pahlow M, Prowe AEF (2010) Model of optimal current feeding in zooplankton. *Mar Ecol Prog Ser* 403: 129–144.
- Landry MR (1981) Switching between herbivory and carnivory by the planktonic marine copepod *Calanus pacificus*. *Mar Biol* 65: 77–82.
- Harbison GR, McAlister VL (1980) Fact and artifact in copepod feeding experiments. *Limnol Oceanogr* 25: 971–981.
- Roman MR, Rublee PA (1980) Containment effects in copepod grazing experiments: A plea to end the black box approach. *Limnol Oceanogr* 25: 982–990.
- Conover RJ, Francis V (1973) The use of radioactive isotopes to measure the transfer of materials in aquatic food chains. *Mar Biol* 18: 272–283.
- Pandolfini E, Thys I, Leporcq B, Descy JP (2000) Grazing experiments with two freshwater zooplankters: fate of chlorophyll and carotenoid pigments. *J Plankton Res* 22: 305–319.
- Sheppard SK, Harwood JD (2005) Advances in molecular ecology: tracking trophic links through predator-prey food webs. *Funct Ecol* 19: 751–762.
- Nejstgaard JC, Frischer ME, Simonelli P, Troedsson C, Brakel M, et al. (2008) Quantitative PCR to estimate copepod feeding. *Mar Biol* 153: 565–577.
- Engström-Ost J, Hogfors H, El-Shehawey R, De Stasio B, Vehmaa A, et al. (2011) Toxin-producing cyanobacterium *Nodularia spumigena*, potential competitors and grazers: testing mechanisms of reciprocal interactions. *Aquat Microb Ecol* 62: 39–48.
- Durbin EG, Casas MC, Ryncarson TA (2012) Copepod feeding and digestion rates using prey DNA and qPCR. *J Plankton Res* 34: 72–82.

29. Stomp M, Huisman J, Voros L, Pick FR, Laamanen M, et al. (2007) Colourful coexistence of red and green picocyanobacteria in lakes and seas. *Ecol Lett* 10: 290–298.
30. Hoogendoorn M, Heimpel GE (2001) PCR-based gut content analysis of insect predators: using ribosomal ITS-1 fragments from prey to estimate predation frequency. *Mol Ecol* 10: 2059–2067.
31. Galluzzi L, Bertozzini E, Penna A, Perini F, Garcés E, et al. (2010) Analysis of rRNA gene content in the Mediterranean dinoflagellate *Alexandrium catenella* and *Alexandrium taylori*: implications for the quantitative real-time PCR-based monitoring methods. *J Appl Phycol* 22: 1–9.
32. Gorokhova E (2005) Effects of preservation and storage of microcrustaceans in RNA later on RNA and DNA degradation. *Limnol Oceanogr Methods* 3: 143–148.
33. Giraffa G, Rossetti L, Neviani E (2000) An evaluation of Chelex-based DNA purification protocols for the typing of lactic acid bacteria. *J Microbiol Methods* 42: 175–184.
34. Becker S, Fahrbach M, Böger P, Ernst A (2002) Quantitative tracing, by Taq nuclease assays, of a *Synechococcus* ecotype in a highly diversified natural population. *Appl Environ Microbiol* 68: 4486–4494.
35. Becker S, Richl P, Ernst A (2007) Seasonal and habitat-related distribution pattern of *Synechococcus* genotypes in Lake Constance. *FEMS Microbiol Ecol* 62: 64–67.
36. Ernst A, Becker S, Wollenzien UA, Postius C (2003) Ecosystem-dependent adaptive radiations of picocyanobacteria inferred from 16S rRNA and ITS-1 sequence analysis. *Microbiology* 149: 217–228.
37. Vermeulen J, Pattyn F, De Preter K, Verduyck L, Derveaux S, et al. (2009) External oligonucleotide standards enable cross laboratory comparison and exchange of real-time quantitative PCR data. *Nucl Acids Res* 37: e138.
38. Smith CJ, Nedwell DB, Dong LF, Osborn AM (2006) Evaluation of quantitative polymerase chain reaction-based approaches for determining gene copy and gene transcript numbers in environmental samples. *Environ Microbiol* 8: 804–815.
39. Vaitomaa J, Rantala A, Halinen K, Rouhiainen L, Tallberg P, et al. (2003) Quantitative real-time PCR for determination of microcystin synthetase *E* copy numbers for *Microcystis* and *Anabaena* in lakes. *Appl Environ Microbiol* 69: 7289–7297.
40. Helcom (2008). Manual for Marine Monitoring in the COMBINE Programme of HELCOM, Annex 6: Guidelines concerning phytoplankton species composition, abundance and biomass. Available: <http://www.helcom.fi/action-areas/monitoring-and-assessment/manuals-and-guidelines/combine-manual> Accessed 2013 Oct 17.
41. Mullin MM, Sloan PR, Eppley RW (1966) Relationship between carbon content, cell volume, and area in phytoplankton. *Limnol Oceanogr* 11: 307–331.
42. Callieri C, Lami A, Bertoni R (2011) Microcolony formation by single-cell *Synechococcus* strains as a fast response to UV radiation. *Appl Environ Microbiol* 77: 7533–7540.
43. Hernroth L (1985) Recommendations on methods for marine biological studies in the Baltic Sea. Mesozooplankton biomass assessment. The Baltic Marine Biologists. Publication No. 10, ISSN: 0282–8839.
44. Geller W, Müller H (1981) The filtration apparatus of Cladocera: filter mesh-sizes and their importance on food selectivity. *Oecologia* 49: 316–321.
45. Stockner JG, Shortreed KS (1989) Algal picoplankton and contribution to food webs in oligotrophic British Columbia Lakes. *Hydrobiologia* 173: 151–166.
46. Roff JC, Turner JT, Weber MK, Hopcroft RR (1995) Bacterivory by tropical copepod nauplii: extent and possible significance. *Aquat Microb Ecol* 9: 165–175.
47. Boyd CM (1976). Selection of particle sizes by filter-feeding copepods: A plea for reason. *Limnol. Oceanogr.* 21: 175–180.
48. Berggreen U, Hansen B, Kiorboe T (1988) Food size spectra, ingestion and growth of the copepod *Acartia tonsa* during development: Implications for determination of copepod production. *Mar Biol* 99: 341–352.
49. Bartram WC (1981) Experimental development of a model for the feeding of neritic copepods on phytoplankton. *J Plankton Res* 3: 25–51.
50. Kuosa H (1988) Occurrence of autotrophic picoplankton along an open sea-inner archipelago gradient in the Gulf of Finland, Baltic Sea. *Ophelia* 28: 85–93.
51. Richman S, Heinle DR, Huff R (1977) Grazing by adult estuarine calanoid copepods of the Chesapeake Bay. *Mar Biol* 42: 69–84.
52. Haverkamp T, Acinas SG, Doeleman M, Stomp M, Huisman J, et al. (2008) Diversity and phylogeny of Baltic Sea picocyanobacteria inferred from their ITS and phycobiliprotein operons. *Environ Microbiol* 10: 174–188.
53. Takahashi M, Kikuchi K, Hara Y (1985) Importance of picocyanobacteria biomass (unicellular, blue-green algae) in the phytoplankton population of the coastal waters off Japan. *Mar Biol* 89: 63–69.
54. Uitto A (2000) Diurnal and vertical grazing activity of mesozooplankton during summer on the SW coast of Finland. *Boreal Environ Res* 5: 137–146.
55. Burdloff D, Gasparini S, Villate F, Uriarte I, Cotano U, et al. (2002) Egg production of the copepod *Acartia bifilosa* in two contrasting European estuaries in relation to seston composition. *J Exp Mar Biol Ecol* 274: 1–17.
56. Bouvy M, Pagano M, M'Boup M, Got P, Troussellier M (2006) Functional structure of microbial food web in the Senegal River Estuary (West Africa): impact of metazooplankton. *J Plankton Res* 28: 195–207.
57. Olsen BR, Dahlgren K, Schander C, Båmstedt U, Torerapp H, et al. (2012) PCR-DHPLC assay for the identification of predator-prey interactions. *J Plankton Res* 34: 277–285.
58. Cowles TJ, Olson RJ, Chisolm SW (1988) Food selection by copepods: discrimination on the basis of food quality. *Mar Biol* 100: 41–49.
59. Gasparini S, Castel J (1997) Autotrophic and heterotrophic nanoplankton in the diet of the estuarine copepods *Eurytemora affinis* and *Acartia bifilosa*. *J Plankton Res* 19: 877–890.
60. Ohlendieck U, Stühr A, Siegmund H (2000) Nitrogen fixation by diazotrophic cyanobacteria in the Baltic Sea and transfer of the newly fixed nitrogen to picoplankton organisms. *J Mar Syst* 25: 213–219.
61. Ploug H, Adam B, Musat N, Kalvelage T, Lavik G, et al. (2011). Carbon, nitrogen and O₂ fluxes associated with the cyanobacterium *Nodularia spumigena* in the Baltic Sea. *ISME J* 5: 1549–1558.
62. Azam F, Fenchel T, Field JG, Gray JS, Meyer-Reil LA, et al. (1983) The ecological role of water column microbes in the sea. *Mar Ecol Prog Ser* 10: 257–263.
63. Berglund J, Muren U, Båmstedt U, Andersson A (2007) Efficiency of a phytoplankton-based and a bacteria-based food web in a pelagic marine system. *Limnol Oceanogr* 52: 121–131.
64. Wylie JL, Currie DJ (1991) The relative importance of bacteria and algae as food sources for crustacean zooplankton. *Limnol Oceanogr* 36: 708–728.
65. Johansson M, Gorokhova E, Larsson U (2004) Annual variability in ciliate community structure, potential prey and predators in the open northern Baltic Sea proper. *J Plankton Res* 26: 67–80.
66. Gorokhova E (2009) Toxic cyanobacteria *Nodularia spumigena* in the diet of Baltic mysids: evidence from molecular diet analysis. *Harmful Algae* 8: 264–272.
67. Pagano M, Champalbert G, Aka M, Kouassi E, Arfi R, et al. (2006) Herbivorous and microbial grazing pathways of metazooplankton in the Senegal River Estuary (West Africa). *Estuar Coast Shelf Sci* 67: 369–381.