

RESEARCH ARTICLE

Clade-Specific Quantitative Analysis of Photosynthetic Gene Expression in *Prochlorococcus*

María-Carmen Fernández-Pinos¹, Marta Casado¹, Gemma Caballero¹, Erik R. Zinser², Jordi Dachs¹, Benjamin Piña^{1*}

1 Department of Environmental Chemistry, IDAEA-CSIC, Barcelona, Catalonia, Spain, **2** Department of Microbiology, University of Tennessee, Knoxville, Tennessee, United States of America

* bpcbmc@cid.csic.es



OPEN ACCESS

Citation: Fernández-Pinos M-C, Casado M, Caballero G, Zinser ER, Dachs J, Piña B (2015) Clade-Specific Quantitative Analysis of Photosynthetic Gene Expression in *Prochlorococcus*. PLoS ONE 10(8): e0133207. doi:10.1371/journal.pone.0133207

Editor: Amanda M Cockshutt, Mount Allison University, CANADA

Received: August 6, 2014

Accepted: June 24, 2015

Published: August 5, 2015

Copyright: © 2015 Fernández-Pinos et al. This is an open access article distributed under the terms of the [Creative Commons Attribution License](https://creativecommons.org/licenses/by/4.0/), which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the paper and its Supporting Information files.

Funding: This work was supported by the Spanish Government through the Consolider-Ingenio 2010 program (CSD2008-00077) and a MICINN grant (CTM2011-439 30471-C02-01), and CTM2014-51985-R (EMRISK). M.C. Fernández-Pinos acknowledges a predoctoral fellowship from the Spanish National Research Council (CSIC), and E.R. Zinser acknowledges grants (OCE0526072 and OCE1030518) from the National Science Foundation. Additional funding was received from the Catalan

Abstract

Newly designed primers targeting *rbcL* (CO₂ fixation), *psbA* (photosystem II) and *rnpB* (reference) genes were used in qRT-PCR assays to assess the photosynthetic capability of natural communities of *Prochlorococcus*, the most abundant photosynthetic organism on Earth and a major contributor to primary production in oligotrophic oceans. After optimizing sample collection methodology, we analyzed a total of 62 stations from the Malaspina 2010 circumnavigation (including Atlantic, Pacific and Indian Oceans) at three different depths. Sequence and quantitative analyses of the corresponding amplicons showed the presence of high-light (HL) and low-light (LL) *Prochlorococcus* clades in essentially all 182 samples, with a largely uniform stratification of LL and HL sequences. *Synechococcus* cross-amplifications were detected by the taxon-specific melting temperatures of the amplicons. Laboratory exposure of *Prochlorococcus* MED4 (HL) and MIT9313 (LL) strains to organic pollutants (PAHs and organochlorine compounds) showed a decrease of *rbcL* transcript abundances, and of the *rbcL* to *psbA* ratios for both strains. We propose this technique as a convenient assay to evaluate effects of environmental stressors, including pollution, on the oceanic *Prochlorococcus* photosynthetic function.

Introduction

Oceanic phytoplankton are responsible for almost a half of the global net primary production (NPP) [1], whereas marine picocyanobacteria account for 32 to 80% of primary production in the oligotrophic oceans [2–5]. *Prochlorococcus* is the smallest and most abundant photosynthetic organism known on Earth, ubiquitously found throughout the euphotic zone in tropical and subtropical oligotrophic oceans from 40° S to 40° N [6], with population abundances of about 10⁵ cells/mL [7–13]. Therefore, *Prochlorococcus* contributes significantly to the NPP in this latitudinal band [8, 14, 15], playing a relevant role in the global carbon cycle. There are two clades of *Prochlorococcus*, adapted to either high-light (HL) and low-light (LL) conditions [16]. They differ in a number of genetic and ecophysiological characteristics, including their divinyl-

Government through the support to the group on Geochemistry of Climatic and Global Change (2009SGR01178).

Competing Interests: The authors have declared that no competing interests exist.

chlorophyll *chl*_{b2} / *chl*_{a2} ratios, ribosomal 16S rDNA sequences, and distributions in the water column [16–18]. Multiple genetic and physiologically different *Prochlorococcus* lineages have been so far characterized [17, 19, 20], showing distinct adaptations to solar irradiance, nutrients availability, and, consequently, different horizontal and vertical distributions [11, 13, 21, 22]. Preliminary data indicate that there are many *Prochlorococcus* lineages still to be characterized in the world oceans [23–29].

Although genetic variability of *Prochlorococcus* has been extensively studied in the ocean, these studies were mainly based on the distinct sequences of the 16S and 23S rDNAs and the 16S/23S rRNA internal transcribed spacer (ITS) [7, 20, 22, 28, 30, 31]. In contrast, analyses of functional gene expression are usually performed in laboratory conditions using pure cultures and specific probes for the target strains [32–37]. When functional genes have been studied in the field, the assessment has been restricted to a limited group of strains [38, 39].

In this paper, we focus on the photosynthetic activity of *Prochlorococcus* given its relevance as a primary producer in the global oceans. With this purpose, we designed new primers for two functional photosynthetic protein-coding genes, *rbcL* and *psbA*, and measured their mRNA abundances by quantitative PCR assays along a global circumnavigation sampling campaign. The gene *rbcL* encodes the large protein subunit of the RuBisCO enzyme, responsible of catalyzing the rate-limiting step of Calvin cycle. RbcL protein abundance has been considered a good representative for the activity of the entire Calvin cycle [40] and hence quantification of the *rbcL* mRNA levels is considered a useful proxy for the carbon fixation activity [38, 39, 41, 42]. While its transcription is mainly regulated by light intensity [41, 43, 44], it also correlates with other variables like nitrogen concentration [45]. The *psbA* gene encodes the photosystem II (PSII) core protein D1, which is the primary target of photo-inactivation and protect the cell from photo-oxidative stress [46]. The damage of D1 protein results in photo-inhibition, decrease of the photosystem II efficiency, and a drop of photosynthetic carbon fixation [47, 48]. Levels of the PsbA protein subunit have been shown to reflect the cellular PSII content [40]. We therefore consider the *psbA* transcript concentration as a proxy of the functionality of the PSII in the cell. Although *psbA* expression is mainly affected by light intensity [32, 44, 49] and UV radiation [48], it is also sensitive to variables such as iron starvation and glucose availability [50]. We selected the *rnpB* gene as endogenous standard of the quantification of the two target genes. This gene encodes the RNA component of RNaseP, a ubiquitous enzyme required for tRNA 5' end maturation in prokaryotes. It has been previously reported as a suitable reference gene for *Prochlorococcus* in qRT-PCR analyses [44, 51–53], since its levels of mRNA remain stable at different conditions of irradiation, iron, phosphate, glucose, UV radiation, nitrogen, or light quality [32–37, 43, 52, 54, 55]. Aiming to integrate the high genetic variability of *Prochlorococcus* strains by a simple biomarker, we designed for each of the three tested genes, *rnpB*, *rbcL* and *psbA*, two separate sets of primers for HL and LL *Prochlorococcus*. A fundamental reason for the selection of these genes (and of the regions within them to be amplified) is their high sequence conservation among the different isolates of the genus, and their relative divergence from homologous sequences from other cyanobacteria. However, giving the close phylogenetic relationship between some LL *Prochlorococcus* and strains from the genus *Synechococcus*, at least some cross-amplification seemed *a priori* unavoidable [20, 28].

There are multiple environmental variables (light, nutrients, etc.) known to affect photosynthesis, and hence prone to alter the expression of both target genes [32, 41, 43, 44, 48, 49]. In this work, we selected persistent organic pollutants (POPs) as anthropogenic stressors of *Prochlorococcus* photosynthesis capacity, since they are broadly distributed over the globe and reach remote oceanic waters [56–58]. They accumulate in marine phytoplankton [59–64] decreasing cyanobacterial growth rate and biomass, inhibiting the PSII, and causing cellular bleaching and death for *Prochlorococcus* [65, 66].

The objectives of the present work were i) to develop a simple and high-throughput amenable methodology to quantify and detect changes in the expression of *rbcL* and *psbA* genes in *Prochlorococcus*; ii) to test the applicability and specificity of this methodology to both laboratory axenic cultures and field samples collected during an oceanographic cruise; and iii) to analyze the potential effects of POPs, under controlled conditions on *Prochlorococcus* pure cultures. The main goal is to generate a molecular tool applicable to natural communities to study the effects of the influence of diverse environmental stressors, and particularly anthropogenic ones, on the photosynthetic capacity of *Prochlorococcus* natural communities and, ultimately, on the oceanic carbon cycle.

Material and Methods

Optimization of sample collection and RNA extraction

We performed preliminary filtration tests using water samples from the North Western Mediterranean Sea (41° 39.7' N 02° 54.6' E). Based on previous estimation of *Prochlorococcus* mRNA half-life rates [67], we set a maximum operation time of 10 minutes from sampling to RNA stabilization, and all parameters were adjusted to this limit. We tested three different 47-mm-diameter, 0.2- μm -pore-size filters: Nucleopore polycarbonate (Whatman, Freiburg, GE), Durapore PVDF (Millipore, Billerica, MA) and Omnipore PTFE (Millipore). Finally, for each filter type we evaluated the saturation capacity, cell retention and ease of use for nucleic acids isolation.

Design of primers and specificity checks using axenic cultures

Primers for *Prochlorococcus* *rbcL*, *psbA* and *rnpB* genes were designed *de novo* for both HL and LL clades (Table 1), starting from different known *Prochlorococcus* sequences ([68], listed in S1 Table) using Geneious 5.6.6, Biomatters (available from <http://www.geneious.com/>). We tested the specificity and efficiency of the designed primers by qRT-PCR analyses using total RNA from axenic cultures of *Prochlorococcus* MED4 and MIT9515 HL strains, MIT9313 and NATL2A LL strains, and *Synechococcus* WH7803 strain. Cultures were grown in artificial media for *Prochlorococcus* (AMP-I) [69], at 22°C in a natural diel light cycle incubator [69] with a maximum irradiance of 100 $\mu\text{mol quanta m}^{-2} \text{s}^{-1}$. In addition, a culture of *Prochlorococcus* EQPAC1 strain [31] was obtained from the Roscoff Culture Collection and processed for nucleic acid analyses without further manipulation. All the strains were collected at approximately 4×10^5 cells/mL, similar to the mean abundance of natural *Prochlorococcus* communities [8, 9, 15, 25]. For consistency, the *Synechococcus* WH7803 strain was collected at the same cell density, despite the fact that this genus is typically one or two orders of magnitude less abundant than *Prochlorococcus* [6, 70, 71]. The collection and storage procedure was as used for field samples, which is explained in detail below.

Field sample collection and storage

A total of 182 samples of oceanic water were collected at 62 stations in the Atlantic, Indian and Pacific Oceans during the Malaspina 2010 circumnavigation, from 14 December 2010 to 14 July 2011 aboard the R/V *BioHesperides* (Fig 1 and S2 Table). We sampled three depths at each station: 3 m depth, (Niskin bottle), deep chlorophyll maximum (DCM) depth, and DCM+40 m depth (both in Niskin bottles attached to a rosette—CTD system). Samples were collected between 8 h and 12 h am local time (10 h 45 min \pm 46 min, S2 Table). This time period coincides with the peak of carbon fixation by *Prochlorococcus*, which occurs between dawn and midday [43, 72], as well as with the maximal expression of *rbcL* and *psbA* genes [44]. One liter

Table 1. Designed primers used for qRT-PCR analysis, and expected PCR fragment sizes.

Gene	Primer sequences (5'- 3')		Amplicon size (bp)
<i>mpB</i> -HL	Fp	GTGTTGGCTAGGTAAACCCCG	81
	Rp	ATCTACTTTTAAGCGCCGCTTG	
<i>rbcL</i> -HL	Fp	ATGGTCATCCATGGGGTTCAGC	104
	Rp	GGTCGCGAAATCGAAAAAGAGAGT	
<i>psbA</i> -HL	Fp	ACCAGTTTCAGCAGCTTTCGCA	128
	Rp	TGTTTTCCAGGCAGAGCACAACA	
<i>mpB</i> -LL	Fp	TGCCACAGAAAMACACCGC	106
	Rp	GCATCGAGAGGTGCTGGC	
<i>rbcL</i> -LL	Fp	GAAGATATCCGCTTCCCGATGGC	140
	Rp	AAGCCAAAGCTTGGCCTTTCTGG	
<i>psbA</i> -LL	Fp	TCTGGTGTGTGTTCTTCCAG	200
	Rp	GTATGCGCCCTTGGATCTGTGT	

Fp: Forward primer.

Rp: Reverse primer.

doi:10.1371/journal.pone.0133207.t001

of seawater from each sample was transferred from the Niskin bottle to a glass bottle, previously rinsed with milli-Q water and a small volume of the sampled seawater, prefiltered onto a 20- μ m-pore-size net, and finally filtered onto 47-mm-diameter, 0.2- μ m-pore-size PTFE filter at 80 mbar vacuum pressure. Filters were split into two halves. One half was preserved in lysis buffer (50 mM Tris-HCl, 40 mM EDTA, 0.75 M Sucrose) and stored at -20°C for genomic DNA extraction. The second half was preserved in RNAlater (Sigma-Aldrich, Saint Louis, MO)

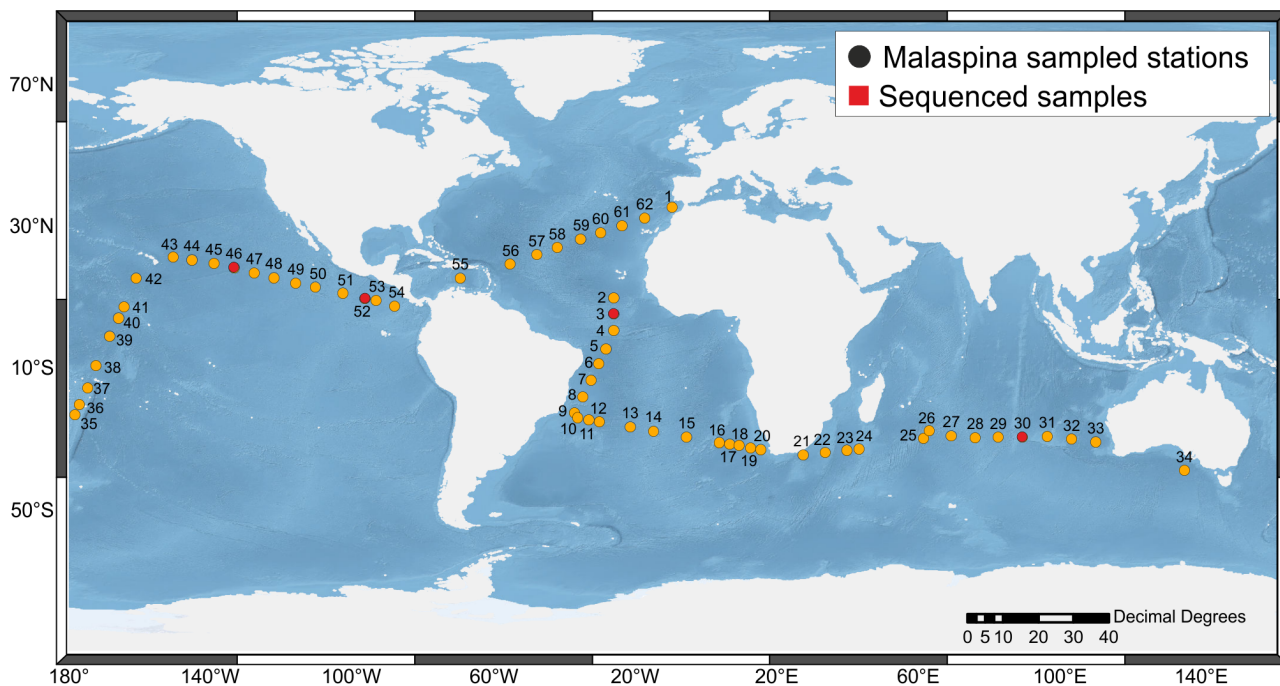


Fig 1. Oceanic stations from Malaspina 2010 circumnavigation selected for RNA analysis. Stations labeled in red correspond to those from which amplicon sequences were obtained.

doi:10.1371/journal.pone.0133207.g001

for RNA isolation analysis and preserved at -80°C . In all cases, we kept rigorously the time limit of 10 min between sample collection and the storage of the filter.

Experimental exposure of axenic cultures to pollutant mixtures

To test the sensitivity of the designed primers to the potential effects of environmental stressors of photosynthesis presumably inducing a small variation on the gene expression, we decided to perform experiments under controlled conditions using axenic cultures and commercial mixtures of organic pollutants. We selected the well-characterized MED4 and MIT9313 *Prochlorococcus* strains as representatives of HL and LL clades, respectively, since they differ in size, nutritional requirements, preferences in light intensity, and genome size [73]. We chose as stressors a mixture of the US Environmental Protection Agency 16 priority polycyclic aromatic hydrocarbons (PAHs) [74] (Dr. Ehrenstorfer, Augsburg, GE), and a mixture of five organochlorine pesticides (OCIP), hexachlorobenzene (HCB), and hexachlorocyclohexane (HCH, α , β , δ , and γ isomers), given their toxicity, persistence, potential to bioaccumulate, and environmental relevance [57, 58, 75, 76]. Growing MED4 and MIT9313 cultures were challenged with either pollutant mixtures. The PAH mixture was added to a final concentration of 700 ng/L, the approximated concentration estimated to reduce growth of natural populations of *Prochlorococcus* by 10% (LC10) [66]. No equivalent toxicity data exist for the OCIP mixture; thus, we used a final concentration of 500 ng/L that is known to have no effect on growth rate in the experimental conditions (not shown).

We performed 4 experiments, one for each combination of strain and pollutant mixture. For each experiment, we used a culture adapted previously to the experimental conditions, which was diluted to an approximate concentration of 10^5 cells/mL. After two days, once it had reached a concentration of $2\text{--}4 \times 10^5$ cells/mL, we split it into twelve 0.5 L flasks. 6 flasks were challenged with the pollutant mixture (treatments) and the other 6 were grown in regular conditions (controls). Three pairs of treatment and control flasks were collected after 30 minutes and the rest after 24 hours of exposure. Every experiment was started at 11 h local time, at the same point of the incubator diel cycle, when the radiation was still increasing before reaching the maximum, to avoid differences between experiments due to the effect of the diel cycle on the mRNA abundances of the target genes [44]. Cell concentration was measured at the beginning and at the end of the experiment with a flow cytometer (Guava easyCyte, Millipore), using the characteristic red auto-fluorescence of *Prochlorococcus* [77]. To analyze the content of chlorophyll *a*, a 50 mL aliquot of each culture was filtered through a GF/F filter (Whatman) under vacuum pressure and stored at -20°C until analysis. Samples for molecular analyses were collected and preserved as described above for field samples.

DNA and RNA isolation

For DNA extraction, the corresponding half filters, once thawed, were incubated at 37°C for 45 minutes with a 5 mg/mL lysozyme solution in lysis buffer. Then, 0.5 mg/mL proteinase K and 100 μL of 10% sodium dodecyl sulphate were added and further incubated at 55°C for 1 h. DNA was extracted twice with phenol-chloroform-isoamyl alcohol (25:24:1; pH 8), and once with chloroform-isoamyl alcohol (24:1). Genomic DNA from the aqueous phase was then precipitated adding ammonium acetate to 0.4 M and one volume of isopropyl alcohol. After 20 minutes of incubation at -20°C , the precipitate was centrifuged and the supernatant was removed. The obtained DNA pellet was washed with 70% ethanol and finally suspended in 30 μL of TE (10 mM Tris, 1 mM EDTA; pH 8).

Total RNA was isolated from the RNA-stabilized filter using the mirVana kit (Ambion, Austin, TX) [78], after removing the excess of RNAlater. The final elution volume (about

100 μ L) was concentrated by partial lyophilization to approximately 40 μ L, and total RNA concentration was measured by spectrophotometric absorption at 260 nm with a NanoDrop ND-8000 spectrophotometer (NanoDrop Technologies, Delaware, DE). RNA quality was checked with an Agilent 2100 Bioanalyzer (Agilent Technologies, Santa Clara, CA). Total RNA was treated with DNase I (Ambion) to remove genomic DNA contamination and reverse transcribed to cDNA using First Strand cDNA Synthesis Kit (Roche, Mannheim, GE). Lastly, the resulting cDNA was stored at -20°C for further analyses.

qRT-PCR analysis, cloning and sequencing of amplicons from cultures and field samples

Aliquots of genomic DNA (2.2 ng for culture samples or 5 ng for field samples), or total RNA (2.2 ng for culture samples or 3.75 ng for field samples), were used to quantify specific transcripts by qRT-PCR in a LightCycler 480 (Roche Diagnostics, Indianapolis, IN) thermocycler using SYBR Green Mix (Takara Bio Inc., Siga, Japan). After thermal activation at 95°C for 10 s, forty-five amplification cycles (95°C for 5 s, 60°C for 35 s), followed by a melting curve program (65–95°C with a heating rate of 0.11°C/s) and a final extension step at 60°C for 30 s. PCR efficiency values for the tested genes were calculated as described elsewhere [79]. When necessary, we redesigned the primers until reaching efficiencies between 95% and 105%. Relative mRNA or genomic DNA abundances of the different genes were calculated using the second derivative maximum of their respective amplification curves (C_p). All samples were run by duplicate, ensuring that the difference between the replicates was less than 0.25 cycles. C_p values are inversely correlated with the logarithm of the initial number of copies of the amplified DNA sequence, N_0 , following the equation

$$N_0 = k(1 + E)^{-C_p}$$

in which E is the efficiency of the reaction (equal to 1 if the primer efficiency is 100%), and k a coefficient related to the number of amplified molecules needed to detect the amplification product by the instrument. Therefore, high C_p values implicate low initial concentrations, and vice versa.

For relative gene expression analyses, we normalized the C_p values of the *rbcL* and *psbA* target genes (tg) to the corresponding values of the *rnpB* reference gene (ref) to obtain ΔC_p values, $\Delta C_p = C_{p_{ref}} - C_{p_{tg}}$, to account for differential cell concentrations and sample processing when quantifying gene expression. The ratios between treatments and controls mRNA/DNA levels were calculated from these ΔC_p values, as

$$\frac{\text{Copies}_{\text{Treatment}}}{\text{Copies}_{\text{Control}}} = 2^{\Delta C_{p_{\text{Treatment}}} - \Delta C_{p_{\text{Control}}}}$$

Similarly, the ratios between *rbcL* and *psbA* mRNA/DNA levels were calculated as

$$\frac{\text{Copies}_{rbcL}}{\text{Copies}_{psbA}} = 2^{\Delta C_{p_{rbcL}} - \Delta C_{p_{psbA}}}$$

Melting temperature (T_m) values for each amplified product from each sample was calculated as the negative first derivation curve of the fluorescence intensity curve over temperature set in the PCR protocol (-dF/dT curve) [80]. Only amplifications with single peaks, denoting a single amplification product, were considered for quantification. PCR products of MED4 and MIT9313 cultures and field samples from 4 oceanic stations (Table 2) were cloned into the vector pTZ57R/T (InsTAclone PCR clone kit, Thermo Scientific, Waltham, MA) and propagated using XL-Blue competent cells. DNA sequencing was performed on 3730 DNA Analyzer

Table 2. Coordinates and collection data from four different oceanic stations sampled during Malaspina circumnavigation.

Sample Name	Station number	Location	Date (dd-mm-yy)	DCM depth (m)	Mixed layer depth (m)	Collection local time (h)		
						3m	DCM	DCM+40
Atlantic	3	05 0.40 N. 26 1.59 W	30-12-10	120	76	8:16	10:35	10:35
Indian	30	29 40.27 S. 89 26.46 E	05-03-11	135	36	8:05	10:12	10:15
Pacific 1	46	18 4.30 N. 133 19.27 W	22-05-11	125	94	7:16	10:32	10:34
Pacific 2	52	9 26.42 N. 96 20.17 W	04-06-11	19	18	7:20	10:09	10:08

doi:10.1371/journal.pone.0133207.t002

(Applied Biosystems), and we compared the results to existing DNA sequences by the BLAST algorithm at NCBI server (<http://www.ncbi.nlm.nih.gov/blast/Blast.cgi>) [81]. We performed phylogenetic trees including the closest sequences identified in the BLAST analyses and the newly amplified sequences using Geneious 5.6.6, Biomatters (available from <http://www.geneious.com>) for sequence alignment and phylogenetic tree design. Final unrooted tree diagrams were drawn using the FigTree software (<http://tree.bio.ed.ac.uk>), and edited using the Corel Draw program (Corel Corporation, Ottawa, Ontario, Canada).

Chlorophyll analyses

Frozen filters were extracted in variable volumes (3–5 ml) of 95% acetone during 24 hours followed by sonication during 5 min at low temperature (4°C). Extracts were then centrifuged at 4000 rpm for 10 minutes to remove cell and filter debris. A 1.5-mL aliquot of the acetone extract was measured for chlorophyll using a UV-spectrophotometer. We assumed that the absorbance at 665 nm was linearly correlated to the total amount of divinyl-chlorophyll *a* [82].

Statistical tests

Statistical tests were performed using the SPSS 19 (SPSS Inc., Chicago, IL) package, with additional calculations performed using the R package (<http://CRAN.R-project.org/>). Normality of data distributions was checked by the Kormogorov-Smirnov test. Differences on mRNA abundances at different sampling depths were tested by the ANOVA plus Tukey HSD post-hoc test. Effects on the *rbcl* and *psbA* gene expression or DNA abundance due to organic pollutant mixtures were tested by paired T-tests between treated and untreated cultures. A 3-way-ANOVA general linear model (GLM) was used to evaluate the differential effects of pollutants on cultures by comparison of the mean fold changes of the target genes of each treatment with respect to its control (mRNA Copies_{Treatment}/Copies_{Control}). Three categorical predictors were used in the GLM to assess the variability on gene expression: Treatment (PAH mixture vs. OCIP mixture), Strain (MED4 vs. MIT9313) and Time of exposure (0.5 h vs. 24 h). Significance levels were set to 0.05.

Results

Methodological optimization

Sampling procedure. Comparison of filtration rates of PVDF, PTFE and PC filters under vacuum showed that only PVDF and PTFE filters maintained a filtration rate fast enough to filter 1 L of seawater in less than ten minutes, a time considered optimal to minimize changes on mRNA levels during sampling (graphs in [S1 Fig](#)). In addition, PTFE filters were compatible with both nucleic acid stabilization reagents used in this work (RNAlater for RNA and DNA lysis buffer for DNA), and did not interfere with the corresponding RNA and DNA extraction

methods. A typical RNA extraction from 1 L of natural sea water yielded 100 to 3000 ng of DNA/RNA (average 900 ng of DNA/RNA), an amount considered sufficient for the intended analyses. Whereas PVDF and PTFE showed similar (and fast) filtration rates (S1 Fig), PTFE filters showed better RNA extraction efficiency and cell retention (S2 and S3 Figs).

Specificity of designed primers. qRT-analysis of cDNA samples from axenic cultures of HL *Prochlorococcus* (MED4, MIT9515 and EQPAC1) and LL strains (MIT9313 and NATL2A) showed a strong specificity of the designed primers (S3 Table). HL and LL strains amplified with the appropriate HL or LL primers showed Cp values at least 10 cycles lower than the values obtained with the non-appropriate LL or HL primers, which represents a 1000 fold difference in amplification efficiency. Amplicons from MED4 and MIT9313 amplified with HL and LL primer pairs, respectively, were sequenced and the results corresponded to the expected *rbcL*, *psbA* and *rnpB* sequences when analyzed by BLAST (Table 3). LL *rnpB* primers showed detectable amplification products when challenged with *Synechococcus* templates with similar efficiency than LL *Prochlorococcus* templates (S3 Table), whereas both LL *rbcL* and LL *psbA* primers gave Cp values more than 10 cycles higher for *Synechococcus* than for *Prochlorococcus*—again, a 1000-fold lower efficiency. The LL *rbcL* amplification products of *Prochlorococcus* and *Synechococcus* differed in their melting temperature (T_m), which ranged from 81°C to 86°C for LL *Prochlorococcus*, and from 88 to 89°C for *Synechococcus* strains (see S4 Fig and S3 Table), suggesting that this last amplicon corresponded to *Synechococcus*, although presenting a low-efficiency amplification. This characteristic can be used in field samples to detect the presence of *Synechococcus* or to evaluate sequence heterogeneity within a given sample. S4 Fig illustrates this possibility, as it shows the T_m profile from a cDNA preparation that combined cDNA of *Prochlorococcus* strain MIT9313 and *Synechococcus* strain WH7803 at equal cell concentrations.

Applicability of the developed methodology to *Prochlorococcus* natural communities

Analysis of mRNA abundance in oceanic populations of *Prochlorococcus*. Analysis of *rbcL*, *psbA* and *rnpB* Cp values in samples from the vertical profiles of 62 oceanic stations showed the expected distribution of HL and LL sequences at the three sampled depths (Table 4) [11, 18, 22]. In 3 m depth samples, amplification with HL primers occurred 5–10 cycles earlier than with LL primers. This indicates a concentration of the corresponding mRNAs 30 to 1000 times higher, according to the reverse correlation between Cp values and the mRNA concentrations. The reverse occurs at both DCM and DCM+40, in which LL primers showed lower Cp values than HL primers (Table 4). If we consider *rnpB* level as a direct indicator of the concentration of metabolically active cells, the data suggest that HL strains represent some 97% of the total *Prochlorococcus* cells at 3 m depth, a mere 6% at DCM and from 1 to 2% at DCM+40 (Table 4). A characteristic of these data is their relatively low variation coefficient, less than 10% in most cases. If we consider that they correspond to samples from three oceans and collected in a period of seven months, this represents a remarkable homogeneity of *Prochlorococcus* populations across the globe.

Fig 2 shows the global distributions of relative abundances of *rbcL* and *psbA*, normalized by the reference gene *rnpB*, for the two clades and at the three sampled depths. The relative mRNA abundances of both HL genes decreased with the depth, being maximal at 3 m depth and minimal at DCM+40. On average, 3 m depth samples showed five times more HL *rbcL*, and three times more HL *psbA* mRNA copies than DCM+40 samples (Fig 2). LL genes showed the opposite tendency, although the effect was less clear. LL *rbcL* mRNA was about three times more abundant at DCM, or DCM+40, than at 3 m depth samples, whereas no significant

Table 3. Characterization by BLAST of sequenced amplicons obtained using the designed primers on total RNA samples from *Prochlorococcus* MED4 and MIT9313 pure cultures.

Primers	Specie/ Strain	Query coverage	Identity	Accession no.
MED4 (HL)				
<i>mpB</i> -HL	<i>Pro.</i> MED4	98%	100%	BX548174.1
	<i>Pro.</i> PCC9511	98%	100%	AJ272223.1
<i>rbcL</i> -HL	<i>Pro.</i> MED4	100%	100%	BX548174.1
	<i>Pro.</i> MIT9215	100%	93%	CP000825.1
<i>psbA</i> -HL	<i>Pro.</i> MED4	100%	100%	BX548174.1
	<i>Pro.</i> MIT9515	100%	99%	AY599030.1
	<i>Pro.</i> MIT9312	100%	99%	AY599028.1
	<i>Pro.</i> MIT9116	100%	99%	AY599031.1
	<i>Pro.</i> MIT9301	100%	98%	CP000576.1
	<i>Pro.</i> MIT9302	100%	98%	AY599029.1
MIT9313 (LL)				
<i>mpB</i> -LL	<i>Pro.</i> MIT9313	100%	99%	BX548175.1
	<i>Pro.</i> MIT9303	100%	98%	CP000554.1
	<i>Syn.</i> WH8102	100%	93%	BX569689.1
	<i>Syn.</i> WH7803	100%	91%	CT971583.1
	<i>Syn.</i> CC9311.	100%	91%	CP000435.1
<i>rbcL</i> -LL	<i>Pro.</i> MIT9313	100%	100%	BX548175.1
	<i>Pro.</i> MIT9303	100%	96%	CP000554.1
	<i>Syn.</i> PCC7009	95%	81%	AM701777.1
	<i>Syn.</i> CB0209	95%	81%	AY452729.1
<i>psbA</i> -LL	<i>Pro.</i> MIT9313	100%	99%	BX548175.1
	<i>Pro.</i> MIT9303	94%	98%	AY599035.1
	<i>Syn.</i> CC9902	100%	88%	CP000097.1
	<i>Syn.</i> WH8102	97%	88%	BX569693.1

Pro. *Prochlorococcus*; *Syn.*: *Synechococcus*.

doi:10.1371/journal.pone.0133207.t003

differences were observed for LL *psbA*, in part due to the high dispersion of the values at 3 m depth (Table 4).

Analysis of mRNA sequence heterogeneity in oceanic samples. A combined analysis of *T_m* values of the different amplicons and their corresponding *C_p* values was used to estimate the sequence heterogeneity (and hence, genetic variability) within each sample (Fig 3). *mpB*

Table 4. Descriptive statistics of *C_p* values corresponding to the different primer pairs, stations, and depths sampled during Malaspina Circumnavigation.

	3 m		DCM		DCM+40	
	mean ± sd	n	mean ± sd	n	mean ± sd	n
<i>mpB</i> -HL	20.98±2.34	61	22.65±2.73	61	26.17±2.57	60
<i>rbcL</i> -HL	21.09±2.35	61	24.05±2.8	61	28.71±2.33	60
<i>psbA</i> -HL	16.87±2.38	61	19.85±2.71	61	23.64±2.31	60
<i>mpB</i> -LL	26±3.09	61	18.8±1.58	61	20.67±2.25	60
<i>rbcL</i> -LL	31.83±3.14	61	23.22±2.63	61	24.84±2.72	60
<i>psbA</i> -LL	25.07±3.85	61	17.38±2.51	61	20.02±2.85	60

doi:10.1371/journal.pone.0133207.t004

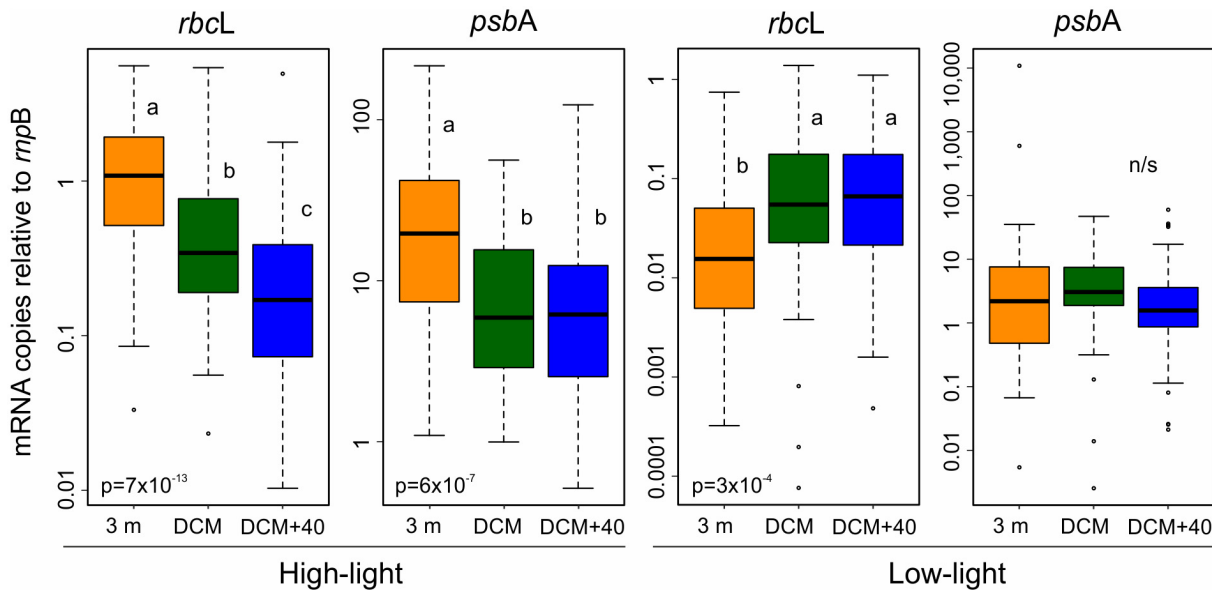


Fig 2. Relative expression of *rbcL* and *psbA* in surface (orange), DCM (green) and DCM+40 m (blue) samples for HL (left) and LL (right) clades. The boxes indicate data distribution parameters: range (whiskers), 25 to 75 percentiles (boxes) and median (thick horizontal bar). Circles represent outliers. Letters indicate statistically different distribution (ANOVA plus SD Tukey post hoc test); the corresponding p values are indicated at the bottom.

doi:10.1371/journal.pone.0133207.g002

and *psbA* HL amplicons showed an extreme homogeneity of T_m values for all 182 samples, irrespective of the depth or of their relative abundance (Fig 3). In fact, all samples showed essentially the same T_m value, 81 and 82°C, respectively, within a margin of few tenths of°C (Table 5). The situation was similar for HL *rbcL* amplicons, although some samples, particularly at DCM+40, showed clearly differentiated T_m values (Fig 3, central left panel). In this case, 13% of DCM+40 samples showed enough HL *rbcL* amplicon heterogeneity to be resolved in two T_m peaks, a phenomenon that was only episodic in samples from 3 m or DCM and unobserved with the other two HL amplicons at any depth (Table 5). These atypical amplicons corresponded to samples with very low abundance in HL sequences—i.e., high Cp values (Fig 3). However, we consider that this T_m variability is still consistent with the natural variation of *Prochlorococcus* strains, as HL *rbcL* amplicons with a low T_m value were also observed for the cultured EQPAC1-C strain (marked as "EQ" in Fig 3, S3 Table in supplementary material).

LL amplicons showed much higher sequence heterogeneity than their HL counterparts (Table 5, Fig 3, right panel). The corresponding graphs in Fig 3 show up to three distinct amplicon populations, particularly at 3 m depth. Also in this case, the highest variability corresponded to samples with low amplicon abundance (i.e., higher Cp values). We attributed a substantial fraction of this amplicon variability, at least at 3 m depth, to the presence of *Synechococcus*, as illustrated by the distribution of the LL *rbcL* amplicon in the central right panel of Fig 3. This particular panel shows a cluster of surface samples at high Cp values and high T_m values, similar to the ones observed for the cultured *Synechococcus* strain WH7803 (marked as WH in Fig 3, S3 Table). However, the data also suggest the presence of some LL *Prochlorococcus* cells. Up to 26% of surface samples showed double T_m peaks for the LL *rbcL* amplicon, one of the peaks coinciding with the typical T_m values for *Prochlorococcus* (Table 5). The other two LL genes do not allow distinction between the two genera, LL *Prochlorococcus* MIT9313 strain (M3 in Fig 3) shows amplicons with essentially the same T_m as their *Synechococcus* counterparts (Fig 3, S3 Table).

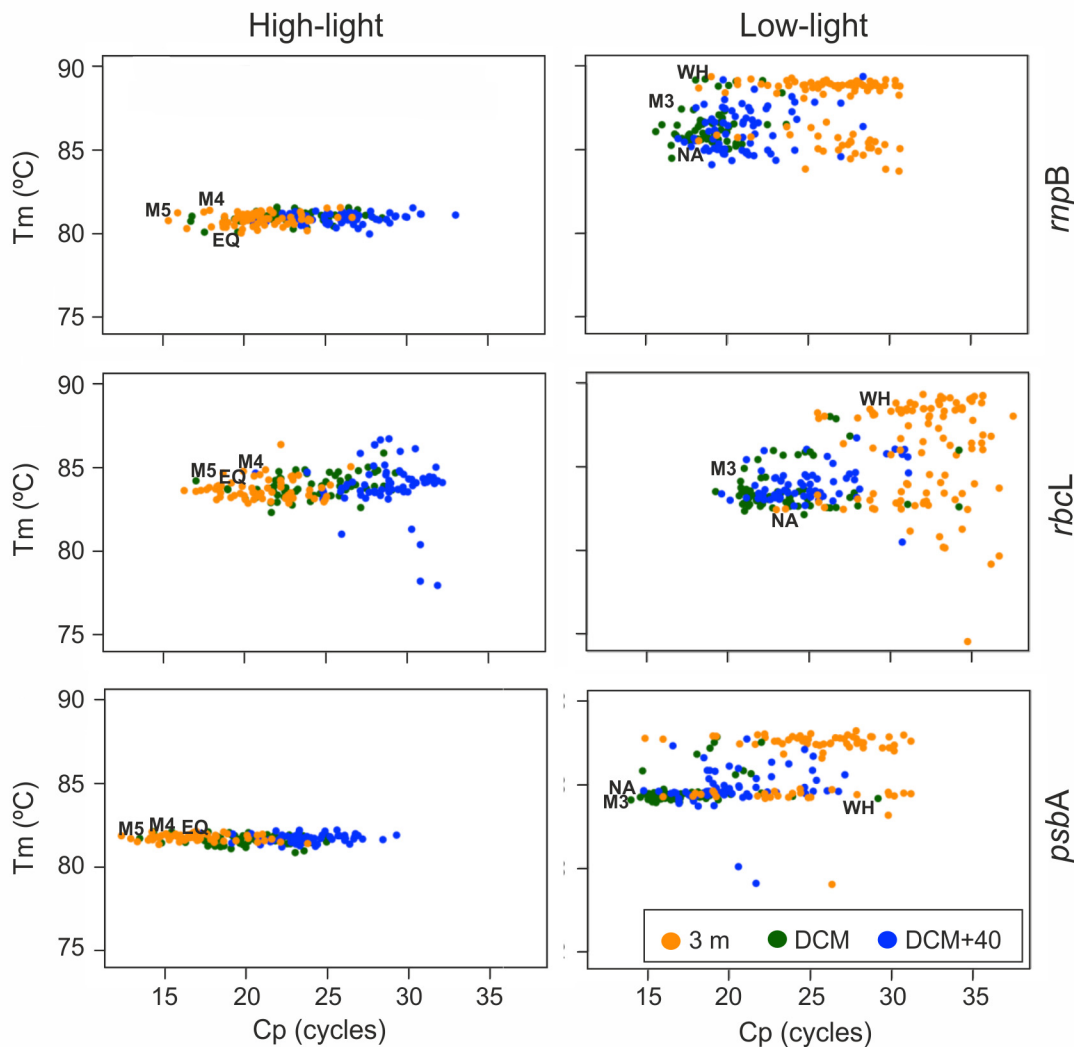


Fig 3. Representation of amplicon melting temperature (T_m) values, in $^{\circ}\text{C}$, plotted against C_p values (in cycles) obtained for field samples at the three sampled depths: Surface (orange), DCM (green) and DCM+40 (blue). For comparison, the graphs include the corresponding values from supplementary S3 Table. HL amplicons: *Prochlorococcus* strains MED4 (M4), MIT9515 (M5) and EQPAC1-C (EQ). LL amplicons: *Prochlorococcus* strains MIT9313 (M3), NATL2A (NA), and the *Synechococcus* strain WH7803 (WH).

doi:10.1371/journal.pone.0133207.g003

Sequence analysis of amplicons from four different stations at the three sampled depths demonstrated that they all encoded the targeted genes (Fig 4 and supplementary S4 Table). Sequences from HL amplicons were very similar in all cases to reported sequences from HL *Prochlorococcus* strains belonging to the eMED4 and eMIT9312 ecotypes (S4 Table). Similarly, amplicons from LL primers showed strong sequence similarities to LL *Prochlorococcus* strains included into the eMIT9313, eMIT9211, eNATL2A, and eSS120 ecotypes [20]. However, ten out of the 21 sequenced LL amplicons from surface samples were identified as *Synechococcus* sequences (S4 Table). This is keeping with the predominance of *Synechococcus* strains in conditions of high solar irradiance [10–12, 16–18, 22], and corroborates our interpretation of the amplicon T_m variability observed in Table 5 and Fig 3.

Table 5. Descriptive statistics of Tm values of amplicons from field samples.

Primers	Cultured strains		Field samples									
	Tm values (°C, x±sd) ^a		Tm values (°C, x±sd)	n	%Double peaks				% Non-average major peaks ^b			
	<i>Prochlorococcus</i>	<i>Synechococcus</i>			All	3 m	DCM	DCM+40	All	3 m	DCM	DCM+40
<i>mpB</i> -HL	81.0 ± 0.3	81.3	80.9 ± 0.3	182	0.0	0.0	0.0	0.0	3.3	4.9	3.3	1.7
<i>rbcL</i> -HL	83.4 ± 2.4	84.4	83.8 ± 0.7	182	5.5	1.6	1.6	13.3	2.7	0.0	1.6	6.7
<i>psbA</i> -HL	81.7 ± 0.1	81.4	81.7 ± 0.2	182	0.0	0.0	0.0	0.0	3.8	0.0	8.2	3.3
<i>mpB</i> -LL	87.5 ± 0.7	88.1	86.7 ± 1.5	182	0.5	1.6	0.0	0.0	0.0	0.0	0.0	0.0
<i>rbcL</i> -LL	83.2 ± 2.5	89.7	84.4 ± 2.0	182	8.8	26.2	0.0	0.0	7.7	23.0	0.0	0.0
<i>psbA</i> -LL	85.6 ± 2.3	82.4	85.3 ± 1.5	182	1.1	0.0	0.0	3.3	1.1	3.3	0.0	0.0

^a) Separated values for each clade. Values from [S2 Table](#).

^b) Percentage of major peaks statistically different (2*sd) from the corresponding average Tm values.

doi:10.1371/journal.pone.0133207.t005

The cladograms in [Fig 4](#) compare the 126 sequenced amplicons with their closest sequences present in GenBank. For HL *rbcL* and *psbA* amplicons (left panels), field samples tend to form small clusters, loosely related to their geographical origin ([Fig 4b and 4c](#)). The relative position of these clusters and bona fide HL *Prochlorococcus* sequences suggests that the variability observed among field samples is similar to the one observed in natural HL *Prochlorococcus* isolates. For example, [Fig 4b and 4c](#) show most field sequences placed between MIT9301 and MED4. The same conclusion applies to HL *rnpB* sequences, despite the different topology of the corresponding cladogram ([Fig 4a](#)). Notice that sequences from LL *Prochlorococcus* strains (MIT9303, MIT9313, MIT9211, SS120, NATL1A and NATL2A) have been added to these HL cladograms as outgroups. They appear in separate branches in all three cladograms.

Cladograms for LL amplicons showed a clear distinction between *Synechococcus* and LL *Prochlorococcus* sequences for *rbcL* and *psbA* ([Fig 4e and 4f](#)). Consistently with the BLAST analyses, most field samples clustered with known *Prochlorococcus* sequences, with the exception of some samples from 3 m depth, which appear placed in *Synechococcus* branches. The topology of the LL *rnpB* cladogram differed from the other two, as it did not show a clear distinction between *Synechococcus* and *Prochlorococcus* sequences ([Fig 4d](#)). As a general conclusion, these cladograms indicate that all DCM and DCM+40 LL sequences corresponded to LL *Prochlorococcus* strains, whereas the results from 3 m samples showed a mixture of *Synechococcus* and LL *Prochlorococcus* sequences.

Effects of pollutant mixtures on *Prochlorococcus* gene expression

Exposure of experimental *Prochlorococcus* MED4 and MIT9313 cultures to either PAHs or OCIP mixtures did not affect significantly growth rates or chlorophyll contents ([S5 and S6 Tables](#)). However, treated cultures showed a decrease of the *rbcL* mRNA levels relative to control cultures, an effect not observed for *psbA* ([Table 6](#)). Combining data for all treatments, *rbcL* mRNA levels were reduced by 20% relative to the reference gene, a moderate, but significant decrease ([Table 6](#)). These data suggest that pollutants may alter the relative expression of photosynthetic genes, an effect likely to be added to the influence of other natural effectors we observed in the natural oceanic populations sampled at different depths ([Fig 2](#)). The analysis of the influence of different parameters on the effects of pollutants in *rbcL* expression (3-way ANOVA) indicated “time of exposure” as the only single factor with significant between-subject effects, as well as an interaction between “time of exposure” and “strain” ([Table 6](#)). These interactions can be visualized in [Fig 5](#), which shows a decrease of the *rbcL* expression and of

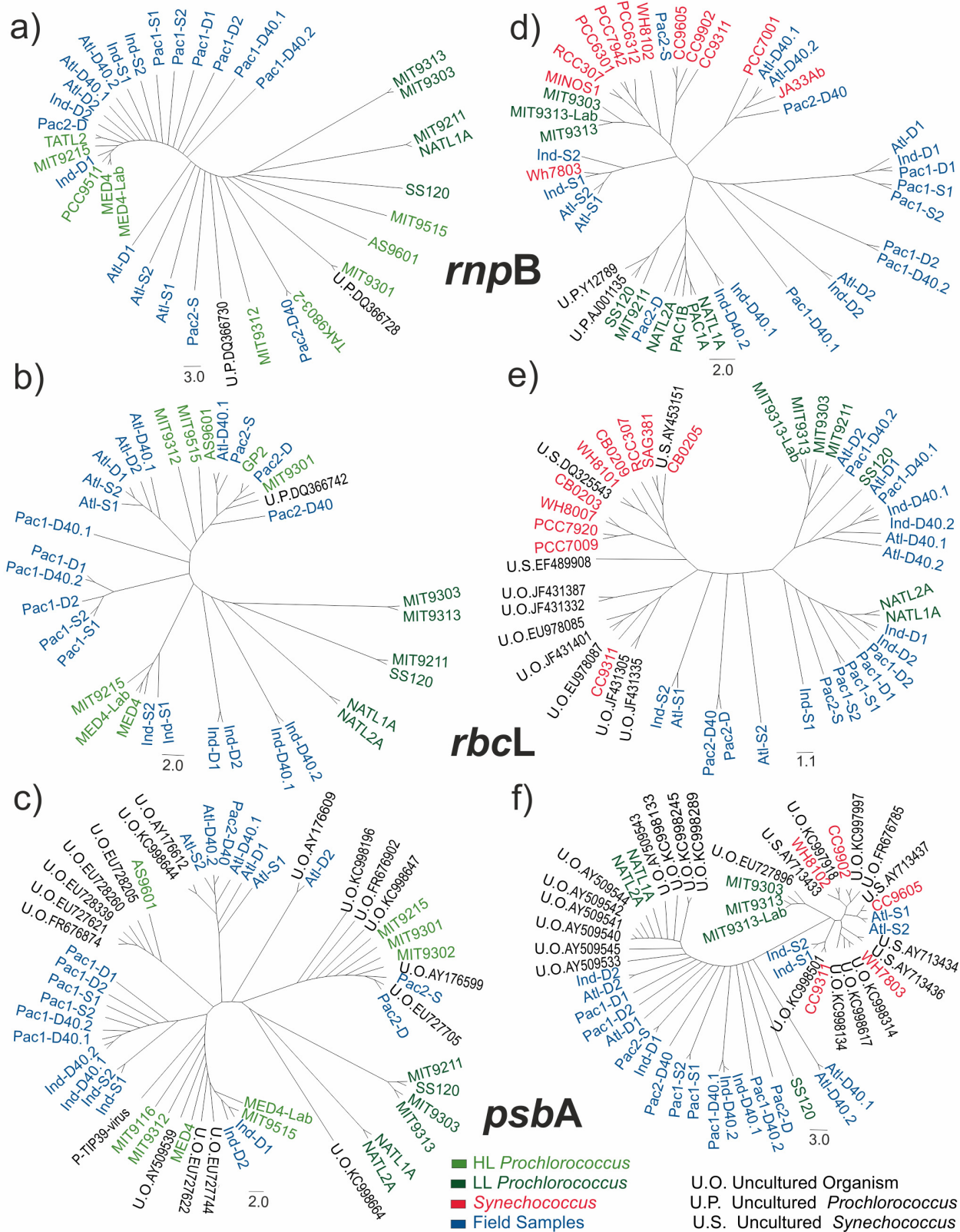


Fig 4. Cladograms of sequences from amplicons derived from field and laboratory samples compared to the closest sequences identified by BLAST. Two qRT-PCR products were sequenced for each sample from Atlantic, Indian and Pacific 1 stations, and only one for Pacific 2 samples (more

information in supplementary S4 Table). Amplicons from surface, DCM, and DCM+40 samples from the different stations are identified with "S", "D" and "D40" letters. Panels a-c correspond to HL amplicons, whereas panels d-f correspond to LL amplicons. For each ecotype, cladograms corresponding to *mmpB* (a,d), *rbcL* (b,e) and *psbA* (c,f) genes are presented. The trees also include sequences from the GenBank. LL *Prochlorococcus*, HL *Prochlorococcus*, and *Synechococcus* (*Syn.*) sequences are indicated by dark green, light green and red colors, respectively. Amplicons from MIT9313 and MED4 strains grown in the lab were also sequenced and included in the analysis as MIT9313-Lab and MED4-Lab, respectively. Sequences for uncultured microorganism isolates are marked as U.P. (identified as *Prochlorococcus*), U.Syn. (identified as *Synechococcus*) or U.O. (not identified).

doi:10.1371/journal.pone.0133207.g004

the *rbcL/psbA* ratio only after 24 hours of treatment for MIT9313, whereas for MED4 both values decreased during the first 0.5 h of exposure, recovering afterwards. While these results should be regarded only as indicative, they suggest that different *Prochlorococcus* strains may have different susceptibilities to organic pollutants, and that their temporal response to them may also differ.

None of these differences between treatments and controls were observed when DNA, rather than RNA, copies were evaluated (supplementary S7 Table, compare to Table 6), which confirms that the observed effects on mRNA levels were indeed reflecting changes in the expression of the target genes and not derived from gene alterations or cell abundances. We concluded that *Prochlorococcus* can be affected by relatively low concentrations of pollutants, and that *rbcL* expression and the *rbcL/psbA* ratio may be sensitive indicators of these effects even in conditions in which the overall growth and chlorophyll concentration remain unaffected.

Discussion

The ubiquitous presence of *Prochlorococcus* in tropical and subtropical oceans makes it extremely important for oceanic and global change studies. However, any global study on this genus needs to overcome the striking genetic variability of *Prochlorococcus* strains. This is especially important when using DNA or RNA-based markers, for they require exact base sequences for adequate amplifications. Our strategy in designing qRT-PCR primers was based in three considerations. First, the targeted genes had to be representative of the main

Table 6. Mean results from paired t-tests between treatment and their respective controls ($\Delta C_{p_{Treatment}}$ vs. $\Delta C_{p_{Control}}$) and from the General Linear Model (GLM) performed with the fold changes ($Copies_{Treatment}/Copies_{Control}$), setting as fixed factors of GLM "Treatment" (PAH, OCIP), "Strain" (MED4, MIT9313) and "Time" (0.5 h, 24 h) for each of the target genes and the ratio between them.

	Value/Factors	df	<i>rbcL</i>	<i>psbA</i> Mean \pm SD ^a	<i>rbcL/psbA</i>
Paired T-test	Ratio (T/C) ^b	21	0.82 \pm 0.41 **	1.09 \pm 0.45	0.78 \pm 0.29 ***
			F ^a		
GLM	Treatment	1	0.292	0.082	2.165
	Strain	1	0.083	0.410	0.338
	Time	1	0.727	0.137	4.762 *
	Treatment * Strain	1	0.130	0.110	0.006
	Treatment * Time	1	0.354	1.983	1.828
	Strain * Time	1	3.349	0.097	14.840 **
	Treatment * Strain * Time	1	0.163	0.039	1.153

^{a)} *, p<0.05;

** , p<0.01;

*** , p<0.001.

^{b)} Treatment versus control.

doi:10.1371/journal.pone.0133207.t006

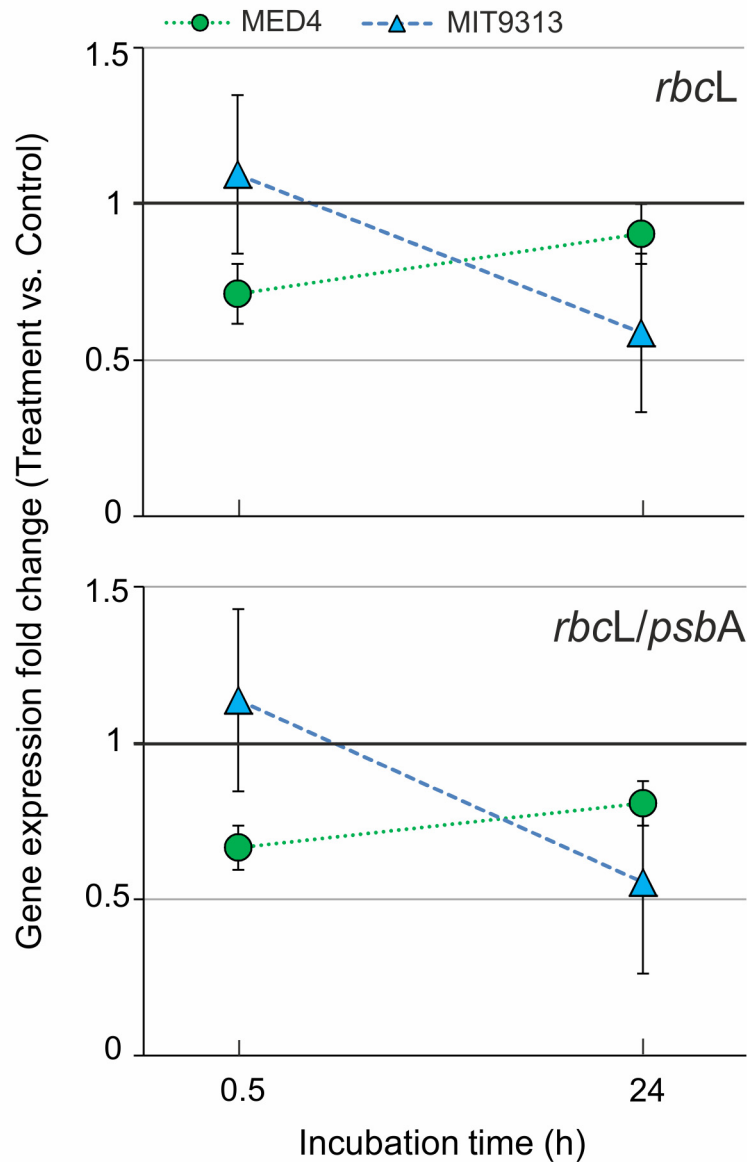


Fig 5. Graphic representation of the different temporal responses of MED4 (green circles) and MIT9313 (blue triangles) to pollutants. Results from PAH and OCIP treatments are pooled together. Significant differences in mRNA abundance between treated and untreated cultures are marked with asterisks (*, $p < 0.05$; **, $p < 0.01$). Whiskers indicate 95% confidence limits.

doi:10.1371/journal.pone.0133207.g005

physiological/ecological functions of *Prochlorococcus*. We selected photosystem functioning and carbon fixation, since these functions are intimately related to the major ecological function of *Prochlorococcus* as a main global primary producer, and because the photosynthetic machinery is at the core of the distinction between *Prochlorococcus* and other cyanobacteria. Secondly, we chose *psbA* and *rbcL* as representatives for PSII functionality and RuBisCo activity, respectively, as these two genes have been particularly well studied and sequenced for many cultured and non-cultured *Prochlorococcus* strains. Finally, we selected sequence regions that

were particularly well conserved among *Prochlorococcus* strains and different from homologous regions in other cyanobacteria. However, the complexity of *Prochlorococcus* populations made it necessary to design two complete sets of primers targeting HL and LL ecotypes. Thus, we designed two sets of primers and tested them in real oceanic samples to analyse relative abundances and sequence heterogeneity. We tested the sensitivity of the developed markers in detecting physiological alterations induced by environmental pollutants in controlled experiments. Our ultimate goal was to generate a molecular tool applicable to natural communities to study the effects of global change vectors on the oceanic carbon cycle.

Unlike DNA, which is relatively stable and that can be found in cellular debris and other dead materials, RNA is easily degraded and only found in living cells. In addition, RNA synthesis (transcription) is a highly controlled process and usually reflects the environmental inputs on the cell metabolism. Therefore, analysis of RNA levels gives information not only of the presence of the cells, but also of their physiological status and metabolic activity. On the other hand, the very rapid turnover of mRNA in prokaryotes makes it difficult to obtain mRNA samples truly representative of the original (i.e., undisturbed) physiological state. To circumvent this problem, we limited manipulation times to an operational minimum and maximized the integrity of the collected RNA by the use of PTFE filters, relatively small volumes of water (1 L), and preserving filtered samples in RNAlater. This strategy allowed to obtain RNA preparations concentrated enough for qRT-PCR analyses and with no signs of RNA degradation.

As for most photosynthesis genes in *Prochlorococcus*, abundances of mRNA for *psbA* and *rbcl* exhibit significant periodicity over the light:dark photoperiod [44, 49]. *rbcl* mRNA peaks at sunrise, while *psbA* peaks at noon. Given these strong and not identical diel periodicities, we were careful to collect field and culture samples at the same time of day (morning). Consequently, *psbA:rnpB*, *rbcl:rnpB*, and/or *psbA:rbcl* ratio data from samples collected at other times of the day may thus not be directly comparable to our data, and will need to be interpreted with caution.

A major issue when analyzing gene expression in *Prochlorococcus* in field communities is to differentiate their mRNA sequences from those from other cyanobacteria and, at the same time, to deal with the large genetic variability of the genus. The designed method allows quantification of mRNA levels of the selected genes belonging from a variety of *Prochlorococcus* strains, because the designed primers are general enough to amplify sequences of close related strains within the genus. As in previous works [28], *Synechococcus* cross-amplifications may occasionally occur, but it can be detected by using the differential T_m values of LL *rbcl* amplicons from both species. This improvement allows a fast and inexpensive method for the detection of cross-amplification that solves a major issue of the genetic similarity between these two genera, and simplifies the markers to just 6 pairs of primers instead of specific primers and/or probes for each strain [32–34, 37]. This simplification is of great importance when using large sets of samples, as in our survey of samples from the Malaspina expedition.

More than 90% of the 126 amplicons sequenced from oceanic samples corresponded to *Prochlorococcus* strains, as judged by sequence homology. The only ten cases in which the amplified fragment encompassed *Synechococcus* sequences corresponded to 3 m depth samples amplified with LL primers, a combination that in almost all cases result in very high Cp values—that is, low specific mRNA levels. The analysis of T_m profiles of the amplified fragments showed that most samples presented single peaks with T_m values close to the expected *Prochlorococcus* values. Again, most exceptions belonged to surface samples amplified with LL primers, in which the appearance of T_m values corresponding to *Synechococcus* amplicons was very usual. Combining both types of results, we concluded that the designed primers amplified the expected *Prochlorococcus* sequences, with very little cross amplification, except for *Synechococcus* strains, found only in surface samples and only with LL primers.

Our results indicate variations in the expression of photosynthetic genes in response to external effectors. In natural HL populations, both *rbcL* and *psbA* showed their maximal relative mRNA levels in 3 m samples, where the conditions of light and oligotrophy were optimal, and minimal levels when they are most unfavorable—at DCM+40. The effect is not so clear for the LL clade, although the *rbcL* gene mRNA was significantly underrepresented in 3 m samples, an unfavorable environment for these strains. It is also noticeable the relationship between poor environmental conditions and sequence heterogeneity. In both clades, the minimal variability (at least in terms of T_m values, arguably a not particularly sensitive parameter) corresponded to the optimal growth conditions (3 m for HL, DCM for LL strains). If we interpret the reduction in photosynthetic gene expression as acclimatization to suboptimal light and/or temperature conditions, strains variability can similarly be interpreted as an adaptation phenomenon, in which different ecotypes may occupy specific niches in the water column.

On top of the natural variation in the relative gene expression of photosynthetic genes, the results from experimental exposures of *Prochlorococcus* to pollutant mixtures suggest that *rbcL* mRNA levels tend to decrease upon the presence of pollutants. This effect was observed in both the LL and HL cultured strains, which showed some differences on the timing and amplitude of the response. Our data also indicates that the ratio between *rbcL* and *psbA* mRNA levels is particularly sensitive to the presence of pollutants. A recent paper demonstrates that the ratio between RuBisCo (RbcL) and PSII (PsbA) molar ratio is strongly correlated to the capacity for electron transport away from PSII, suggesting a limitation of the electron transport rate by the RuBisCo to PSII ratio [40]. This fact may determine the use of photosynthesis rate for carbon fixation or for ATP generation. While it is not possible to directly translate messenger levels to protein concentrations or activities, it is conceivable that the presence of pollutants may increase energy consumption for the cell (for example, for detoxification or extra metabolic activity), reducing its capacity to fix CO_2 and, therefore, the requirement for RuBisCo activity. Further experimental data is needed to test this hypothesis; however, we propose that both *rbcL* expression and the *rbcL/psbA* ratio may be useful indexes of the physiological status of the populations, likely responding to different environmental stressors.

We conclude that the proposed methodology allows not only quantifying the clade-specific photosynthetic potential of *Prochlorococcus* communities in the oceans, but also the assessment of changes in the gene expression due to environmental stressors, as light and/or nutrient conditions, and pollutants. The protocols for sample harvesting and preservation are adequate to ensure the integrity of RNA samples, and the qRT-PCR methodology allows processing of large number of samples at a reasonable cost. The developed methodology for quantifying the gene expression of the *rbcL* and *psbA* genes of *Prochlorococcus*, therefore, represents a first step for a future field and laboratory assessment of the different drivers and stressors affecting the photosynthetic function in *Prochlorococcus*, an essential issue for our understanding of the marine carbon cycle and its modulation in the current scenario of global change.

Supporting Information

S1 Fig. Permeability tests. Three filters of 0.2 μm pore-size and 47 mm of diameter made of different materials were used. Aliquots of 1 L of a seawater sample collected on 08/11/10 in Mediterranean Sea (41 39.7 N 02 54.6) were filtered onto each filter. (TIF)

S2 Fig. DNA recovery. Two water samples of Mediterranean Sea (41 39.7 N 02 54.6) were collected on 11/11/2010 (Nov) and 16/12/2010 (Dec), respectively, using three filters of 0.2 μm pore size and 47 mm of diameter made of different materials. 1 L of water was filtered through

each filter.
(TIF)

S3 Fig. RNA recovery. RNA isolated from both the filter and the pellet resulting from centrifugation of RNAlater where the filter was immersed. Three replicates of 1 L of Mediterranean Sea (41 39.7 N 02 54.6) water filtered onto Ominipore (PTFE) filters are shown. RNA concentration obtained from Ominipore (PTFE) filter is between two and three fold that obtained from the pellet.
(TIF)

S4 Fig. Genetic marker for *Synechococcus*. The graphs shows melting temperature curves of amplicons using *rbcl* LL primers. Melting peaks for *Prochlorococcus* str. MIT9313 and *Synechococcus* str. WH7803 in the same sample are shown
(TIF)

S1 Table. Model strain list. *Prochlorococcus* and *Synechococcus* strains used to design specific primers for *Prochlorococcus* of the selected genes by sequences alignments using Software *Genious*.
(DOCX)

S2 Table. Coordinates and collection data of field samples collected during Malaspina cruise.
(DOCX)

S3 Table. Primer specificity. Specificity of the designed primers checked by qRT-PCR using pure cultures of some *Prochlorococcus* and *Synechococcus* strains at equal cell concentrations and verification of the method validity in natural communities*.
(DOCX)

S4 Table. Amplicon sequences. Sequencing of single amplicons from samples collected at three different depths and three sampled station during Malaspina circumnavigation.
(DOCX)

S5 Table. Cell concentration variability. Paired-sample Wilcoxon signed rank tests achieved with control- treatment pairs of cell concentration measured.
(DOCX)

S6 Table. Changes in chlorophyll a concentrations. Results from comparing means test achieved with control- treatment pairs of chlorophyll a fluorescence.
(DOCX)

S7 Table. Quantitative Genomic changes. Relative changes in DNA abundance for different genes in MIT9313 and MED4 *Prochlorococcus* cultures when challenged with organic pollutant treatments (PAHs or OCIP), evaluated at two different incubation times (0.5 and 24 hours). Results from four different paired T-tests are shown: one of the whole data set (“All samples”) and one of each data subset (“Strain”, “Treatment” and “Incubation time”).
(DOCX)

Acknowledgments

We thank the kind collaboration of J. W. Chandler, L. Ma and M.J. Szul for help with *Prochlorococcus* cultures, and the R/V Hesperides crew for help during the cruise. We also thank J. Gasol for his support on the first steps of the development of the DNA/RNA isolation protocol, and María Vila-Costa for critical reading of the manuscript. This work was funded by the

Ministry of Economy and Competiveness through grants CSD2008-00077 (Malaspina 2010 research project) and CTM2014-51985-R (EMRISK). Additional funding was received from the National Science Foundation (grants OCE0526072 and OCE1030518). M.C. Fernández-Pinos acknowledges a predoctoral fellowship from the Spanish National Research Council (CSIC).

Author Contributions

Conceived and designed the experiments: JD BP ERZ. Performed the experiments: MCFP MC GC. Analyzed the data: MCFP JD BP. Contributed reagents/materials/analysis tools: MCFP MC JP ERZ BP. Wrote the paper: MCFP ERZ JD BP.

References

1. Chavez FP, Messié M, Pennington JT. Marine Primary Production in Relation to Climate Variability and Change. *Annual Review of Marine Science*. 2011; 3(1):227–60. doi: [10.1146/annurev.marine.010908.163917](https://doi.org/10.1146/annurev.marine.010908.163917)
2. Li WK. Composition of ultraphytoplankton in the central north Atlantic. *Marine Ecology Progress Series*. 1995; 122(1–3):1–8.
3. Veldhuis MJW, Kraay GW, Van Bleijswijk JDL, Baars MA. Seasonal and spatial variability in phytoplankton biomass, productivity and growth in the northwestern Indian ocean: The southwest and northeast monsoon, 1992–1993. *Deep-Sea Research Part I: Oceanographic Research Papers*. 1997; 44(3):425–49.
4. Goericke R, Welschmeyer NA. The marine prochlorophyte *Prochlorococcus* contributes significantly to phytoplankton biomass and primary production in the Sargasso Sea. *Deep-Sea Research Part I: Oceanographic Research Papers*. 1993; 40(11/12):2283–94.
5. Liu H, Nolla HA, Campbell L. *Prochlorococcus* growth rate and contribution to primary production in the equatorial and subtropical North Pacific Ocean. *Aquatic Microbial Ecology*. 1997; 12(1):39–47.
6. Partensky F, Hess WR, Vault D. *Prochlorococcus*, a marine photosynthetic prokaryote of global significance. *Microbiology and Molecular Biology Reviews*. 1999; 63(1):106–27. PMID: [10066832](https://pubmed.ncbi.nlm.nih.gov/10066832/)
7. Bouman HA, Ulloa O, Scanlan DJ, Zwirgmaier K, Li WK, Platt T, et al. Oceanographic basis of the global surface distribution of *Prochlorococcus* ecotypes. *Science*. 2006; 312(5775):918–21. doi: [10.1126/science.1122692](https://doi.org/10.1126/science.1122692) PMID: [16690867](https://pubmed.ncbi.nlm.nih.gov/16690867/).
8. Campbell L, Liu H, Nolla HA, Vault D. Annual variability of phytoplankton and bacteria in the subtropical North Pacific Ocean at Station ALOHA during the 1991–1994 ENSO event. *Deep-Sea Research Part I: Oceanographic Research Papers*. 1997; 44(2):167–92.
9. Crosbie ND, Furnas MJ. Abundance, distribution and flow-cytometric characterization of picophytoplankton populations in central (17°S) and southern (20°S) shelf waters of the Great Barrier Reef. *Journal of Plankton Research*. 2001; 23(8):809–28.
10. Garczarek L, Dufresne A, Rousvoal S, West NJ, Mazard S, Marie D, et al. High vertical and low horizontal diversity of *Prochlorococcus* ecotypes in the Mediterranean Sea in summer. *FEMS microbiology ecology*. 2007; 60(2):189–206. doi: [10.1111/j.1574-6941.2007.00297.x](https://doi.org/10.1111/j.1574-6941.2007.00297.x) PMID: [17391326](https://pubmed.ncbi.nlm.nih.gov/17391326/).
11. Johnson ZI, Zinser ER, Coe A, McNulty NP, Woodward EMS, Chisholm SW. Niche partitioning among *Prochlorococcus* ecotypes along ocean-scale environmental gradients. *Science*. 2006; 311(5768):1737–40. PMID: [16556835](https://pubmed.ncbi.nlm.nih.gov/16556835/)
12. Malmstrom RR, Coe A, Kettler GC, Martiny AC, Frias-Lopez J, Zinser ER, et al. Temporal dynamics of *Prochlorococcus* ecotypes in the Atlantic and Pacific oceans. *The ISME journal*. 2010; 4(10):1252–64. doi: [10.1038/ismej.2010.60](https://doi.org/10.1038/ismej.2010.60) PMID: [20463762](https://pubmed.ncbi.nlm.nih.gov/20463762/).
13. Martiny AC, Kathuria S, Berube PM. Widespread metabolic potential for nitrite and nitrate assimilation among *Prochlorococcus* ecotypes. *Proceedings of the National Academy of Sciences of the United States of America*. 2009; 106(26):10787–92. doi: [10.1073/pnas.0902532106](https://doi.org/10.1073/pnas.0902532106) PMID: [19549842](https://pubmed.ncbi.nlm.nih.gov/19549842/); PubMed Central PMCID: PMC2705535.
14. Vault D, Marie D, Olson RJ, Chisholm SW. Growth of *prochlorococcus*, a photosynthetic prokaryote, in the equatorial Pacific Ocean. *Science*. 1995; 268(5216):1480–2. PMID: [17843668](https://pubmed.ncbi.nlm.nih.gov/17843668/)
15. Durand MD, Olson RJ, Chisholm SW. Phytoplankton population dynamics at the Bermuda Atlantic Time-series station in the Sargasso Sea. *Deep-Sea Research Part II: Topical Studies in Oceanography*. 2001; 48(8–9):1983–2003.

16. Moore LR, Rocap G, Chisholm SW. Physiology and molecular phylogeny of coexisting *Prochlorococcus* ecotypes. *Nature*. 1998; 393(6684):464–7. PMID: [9624000](#)
17. Moore LR, Chisholm SW. Photophysiology of the marine cyanobacterium *Prochlorococcus*: Ecotypic differences among cultured isolates. *Limnology and Oceanography*. 1999; 44(3 1):628–38.
18. West NJ, Scanlan DJ. Niche-partitioning of *Prochlorococcus* populations in a stratified water column in the eastern North Atlantic Ocean? *Applied and Environmental Microbiology*. 1999; 65(6):2585–91. PMID: [10347047](#)
19. Moore LR, Post AF, Rocap G, Chisholm SW. Utilization of different nitrogen sources by the marine cyanobacteria *Prochlorococcus* and *Synechococcus*. *Limnology and Oceanography*. 2002; 47(4):989–96.
20. Rocap G, Distel DL, Waterbury JB, Chisholm SW. Resolution of *Prochlorococcus* and *Synechococcus* Ecotypes by Using 16S-23S Ribosomal DNA Internal Transcribed Spacer Sequences. *Applied and Environmental Microbiology*. 2002; 68(3):1180–91. doi: [10.1128/aem.68.3.1180-1191.2002](#) PMID: [11872466](#)
21. Partensky F, La Roche J, Wyman K, Falkowski PG. The divinyl-chlorophyll *a/b*-protein complexes of two strains of the oxyphototrophic marine prokaryote *Prochlorococcus*—Characterization and response to changes in growth irradiance. *Photosynthesis research*. 1997; 51(3):209–22.
22. Zinser ER, Coe A, Johnson ZI, Martiny AC, Fuller NJ, Scanlan DJ, et al. *Prochlorococcus* ecotype abundances in the North Atlantic Ocean as revealed by an improved quantitative PCR method. *Appl Environ Microbiol*. 2006; 72(1):723–32. doi: [10.1128/AEM.72.1.723-732.2006](#) PMID: [16391112](#); PubMed Central PMCID: PMC1352191.
23. Zeidner G, Preston CM, Delong EF, Massana R, Post AF, Scanlan DJ, et al. Molecular diversity among marine picophytoplankton as revealed by *psbA* analyses. *Environmental microbiology*. 2003; 5(3):212–6. PMID: [12588300](#)
24. West NJ, Lebaron P, Strutton PG, Suzuki MT. A novel clade of *Prochlorococcus* found in high nutrient low chlorophyll waters in the South and Equatorial Pacific Ocean. *The ISME journal*. 2011; 5(6):933–44. doi: [10.1038/ismej.2010.186](#) PMID: [21124492](#); PubMed Central PMCID: PMC3131852.
25. Malmstrom RR, Rodrigue S, Huang KH, Kelly L, Kern SE, Thompson A, et al. Ecology of uncultured *Prochlorococcus* clades revealed through single-cell genomics and biogeographic analysis. *ISME Journal*. 2013; 7(1):184–98. doi: [10.1038/ismej.2012.89](#) PMID: [22895163](#)
26. Huang S, Wilhelm SW, Harvey HR, Taylor K, Jiao N, Chen F. Novel lineages of *Prochlorococcus* and *Synechococcus* in the global oceans. *The ISME journal*. 2012; 6(2):285–97. doi: [10.1038/ismej.2011.106](#) PMID: [21955990](#); PubMed Central PMCID: PMC3260499.
27. Rusch DB, Martiny AC, Dupont CL, Halpern AL, Venter JC. Characterization of *Prochlorococcus* clades from iron-depleted oceanic regions. *Proceedings of the National Academy of Sciences of the United States of America*. 2010; 107(37):16184–9. doi: [10.1073/pnas.1009513107](#) PMID: [20733077](#)
28. Martiny AC, Tai AP, Veneziano D, Primeau F, Chisholm SW. Taxonomic resolution, ecotypes and the biogeography of *Prochlorococcus*. *Environmental microbiology*. 2009; 11(4):823–32. doi: [10.1111/j.1462-2920.2008.01803.x](#) PMID: [19021692](#).
29. Steglich C, Mullineaux CW, Teuchner K, Hess WR, Lokstein H. Photophysical properties of *Prochlorococcus marinus* SS120 divinyl chlorophylls and phycoerythrin in vitro and in vivo. *FEBS Letters*. 2003; 553(1–2):79–84. doi: [10.1016/s0014-5793\(03\)00971-2](#) PMID: [14550550](#)
30. Ahlgren NA, Rocap G, Chisholm SW. Measurement of *Prochlorococcus* ecotypes using real-time polymerase chain reaction reveals different abundances of genotypes with similar light physiologies. *Environmental microbiology*. 2006; 8(3):441–54. doi: [10.1111/j.1462-2920.2005.00910.x](#) PMID: [16478451](#).
31. West NJ, Schönhuber WA, Fuller NJ, Amann RI, Rippka R, Post AF, et al. Closely related *Prochlorococcus* genotypes show remarkably different depth distributions in two oceanic regions as revealed by in situ hybridization using 16S rRNA-targeted oligonucleotides. *Microbiology*. 2001; 147(7):1731–44.
32. Berg GM, Shrager J, van Dijken G, Mills MM, Arrigo KR, Grossman AR. Responses of *psbA*, *hli* and *ptox* genes to changes in irradiance in marine *Synechococcus* and *Prochlorococcus*. *Aquatic Microbial Ecology*. 2011; 65(1):1–14. doi: [10.3354/ame01528](#)
33. Bibby TS, Mary I, Nield J, Partensky F, Barber J. Low-light-adapted *Prochlorococcus* species possess specific antennae for each photosystem. *Nature*. 2003; 424(6952):1051–4. PMID: [12944966](#)
34. Coleman ML, Chisholm SW. Ecosystem-specific selection pressures revealed through comparative population genomics. *Proceedings of the National Academy of Sciences of the United States of America*. 2010; 107(43):18634–9. doi: [10.1073/pnas.1009480107](#) PMID: [20937887](#)
35. Gómez-Baena G, López-Lozano A, Gil-Martínez J, Lucena JM, Díez J, Candau P, et al. Glucose uptake and its effect on gene expression in *Prochlorococcus*. *PLOS ONE*. 2008; 3(10).
36. Kolowrat C, Partensky F, Mella-Flores D, Le Corquille G, Boutte C, Blot N, et al. Ultraviolet stress delays chromosome replication in light/dark synchronized cells of the marine cyanobacterium

- Prochlorococcus marinus* PCC9511. BMC microbiology. 2010; 10:204. doi: [10.1186/1471-2180-10-204](https://doi.org/10.1186/1471-2180-10-204) PMID: [20670397](https://pubmed.ncbi.nlm.nih.gov/20670397/); PubMed Central PMCID: PMC2921402.
37. Rangel OA, Gomez-Baena G, Lopez-Lozano A, Diez J, Garcia-Fernandez JM. Physiological role and regulation of glutamate dehydrogenase in *Prochlorococcus* sp. strain MIT9313. Environmental microbiology reports. 2009; 1(1):56–64. doi: [10.1111/j.1758-2229.2008.00005.x](https://doi.org/10.1111/j.1758-2229.2008.00005.x) PMID: [23765721](https://pubmed.ncbi.nlm.nih.gov/23765721/).
 38. John DE, Patterson SS, Paul JH. Phytoplankton-group specific quantitative polymerase chain reaction assays for RuBisCO mRNA transcripts in seawater. Marine biotechnology. 2007; 9(6):747–59. doi: [10.1007/s10126-007-9027-z](https://doi.org/10.1007/s10126-007-9027-z) PMID: [17694413](https://pubmed.ncbi.nlm.nih.gov/17694413/).
 39. John DE, Wang ZA, Liu X, Byrne RH, Corredor JE, Lopez JM, et al. Phytoplankton carbon fixation gene (RuBisCO) transcripts and air-sea CO₂ flux in the Mississippi River plume. The ISME journal. 2007; 1(6):517–31. doi: [10.1038/ismej.2007.70](https://doi.org/10.1038/ismej.2007.70) PMID: [18043653](https://pubmed.ncbi.nlm.nih.gov/18043653/).
 40. Zorz JK, Allanach JR, Murphy CD, Roodvoets MS, Campbell DA, Cockshutt AM. The RUBISCO to photosystem ii ratio limits the maximum photosynthetic rate in picocyanobacteria. Life. 2015; 5(1):403–17. doi: [10.3390/life5010403](https://doi.org/10.3390/life5010403) PMID: [25658887](https://pubmed.ncbi.nlm.nih.gov/25658887/)
 41. Pichard SL, Campbell L, Kang JB, Tabita FR, Paul JH. Regulation of ribulose biphosphate carboxylase gene expression in natural phytoplankton communities. I. Diel rhythms. Marine Ecology Progress Series. 1996; 139(1–3):257–65.
 42. Pichard SL, Campbell L, Carder K, Kang JB, Patch J, Tabita FR, et al. Analysis of ribulose biphosphate carboxylase gene expression in natural phytoplankton communities by group-specific gene probing. Marine Ecology Progress Series. 1997; 149(1–3):239–53.
 43. Bruyant F, Babin M, Genty B, Prasil O, Behrenfeld MJ, Claustre H, et al. Diel variations in the photosynthetic parameters of *Prochlorococcus* strain PCC 9511: Combined effects of light and cell cycle. Limnology and Oceanography. 2005; 50(3):850–63.
 44. Zinser ER, Lindell D, Johnson ZI, Futschik ME, Steglich C, Coleman ML, et al. Choreography of the transcriptome, photophysiology, and cell cycle of a minimal photoautotroph, *Prochlorococcus*. PLOS ONE. 2009; 4(4).
 45. Tolonen AC, Aach J, Lindell D, Johnson ZI, Rector T, Steen R, et al. Global gene expression of *Prochlorococcus* ecotypes in response to changes in nitrogen availability. Molecular systems biology. 2006; 2:53. doi: [10.1038/msb4100087](https://doi.org/10.1038/msb4100087) PMID: [17016519](https://pubmed.ncbi.nlm.nih.gov/17016519/); PubMed Central PMCID: PMC1682016.
 46. Mulo P, Sirpiö S, Suorsa M, Aro E-M. Auxiliary proteins involved in the assembly and sustenance of photosystem II. Photosynthesis research. 2008; 98(1–3):489–501. doi: [10.1007/s11120-008-9320-3](https://doi.org/10.1007/s11120-008-9320-3) PMID: [18618287](https://pubmed.ncbi.nlm.nih.gov/18618287/)
 47. Clarke AK, Campbell D, Gustafsson P, Oquist G. Dynamic responses of Photosystem II and phycobilisomes to changing light in the cyanobacterium *Synechococcus* sp. PCC 7942. Planta. 1995; 197(3):553–62.
 48. Garczarek L, Dufresne A, Blot N, Cockshutt AM, Peyrat A, Campbell DA, et al. Function and evolution of the psbA gene family in marine *Synechococcus*: *Synechococcus* sp. WH7803 as a case study. ISME Journal. 2008; 2(9):937–53. doi: [10.1038/ismej.2008.46](https://doi.org/10.1038/ismej.2008.46) PMID: [18509382](https://pubmed.ncbi.nlm.nih.gov/18509382/)
 49. Garczarek L, Partensky F, Irlbacher H, Holtzendorff J, Babin M, Mary I, et al. Differential expression of antenna and core genes in *Prochlorococcus* PCC 9511 (Oxyphotobacteria) grown under a modulated light-dark cycle. Environmental microbiology. 2001; 3(3):168–75. PMID: [11321533](https://pubmed.ncbi.nlm.nih.gov/11321533/)
 50. García-Fernández JM, Hess WR, Houmard J, Partensky F. Expression of the psbA gene in the marine oxyphotobacteria *Prochlorococcus* spp. Archives of Biochemistry and Biophysics. 1998; 359(1):17–23. PMID: [9799555](https://pubmed.ncbi.nlm.nih.gov/9799555/)
 51. Schon A. Conserved and variable domains within divergent RNase P RNA gene sequences of *Prochlorococcus* strains. International Journal of Systematic and Evolutionary Microbiology. 2002; 52(4):1383–9. doi: [10.1099/ijs.0.01983-0](https://doi.org/10.1099/ijs.0.01983-0)
 52. Martiny AC, Coleman ML, Chisholm SW. Phosphate acquisition genes in *Prochlorococcus* ecotypes: evidence for genome-wide adaptation. Proceedings of the National Academy of Sciences of the United States of America. 2006; 103(33):12552–7. doi: [10.1073/pnas.0601301103](https://doi.org/10.1073/pnas.0601301103) PMID: [16895994](https://pubmed.ncbi.nlm.nih.gov/16895994/); PubMed Central PMCID: PMC1567916.
 53. Gomez-Baena G, Rangel OA, Lopez-Lozano A, Garcia-Fernandez JM, Diez J. Stress responses in *Prochlorococcus* MIT9313 vs. SS120 involve differential expression of genes encoding proteases ClpP, FtsH and Lon. Research in microbiology. 2009; 160(8):567–75. doi: [10.1016/j.resmic.2009.08.009](https://doi.org/10.1016/j.resmic.2009.08.009) PMID: [19732824](https://pubmed.ncbi.nlm.nih.gov/19732824/).
 54. Osburne MS, Holmbeck BM, Frias-Lopez J, Steen R, Huang K, Kelly L, et al. UV hyper-resistance in *Prochlorococcus* MED4 results from a single base pair deletion just upstream of an operon encoding nudix hydrolase and photolyase. Environmental microbiology. 2010; 12(7):1978–88. doi: [10.1111/j.1462-2920.2010.02203.x](https://doi.org/10.1111/j.1462-2920.2010.02203.x) PMID: [20345942](https://pubmed.ncbi.nlm.nih.gov/20345942/); PubMed Central PMCID: PMC2955971.

55. Steglich C, Futschik M, Rector T, Steen R, Chisholm SW. Genome-wide analysis of light sensing in *Prochlorococcus*. *Journal of bacteriology*. 2006; 188(22):7796–806. doi: [10.1128/JB.01097-06](https://doi.org/10.1128/JB.01097-06) PMID: [16980454](https://pubmed.ncbi.nlm.nih.gov/16980454/); PubMed Central PMCID: PMC1636322.
56. Jurado E, Dachs J. Seasonality in the "grasshopping" and atmospheric residence times of persistent organic pollutants over the oceans. *Geophysical Research Letters*. 2008; 35(17).
57. Jurado E, Jaward F, Lohmann R, Jones KC, Simó R, Dachs J. Wet deposition of persistent organic pollutants to the global oceans. *Environmental Science and Technology*. 2005; 39(8):2426–35. PMID: [15884331](https://pubmed.ncbi.nlm.nih.gov/15884331/)
58. Lohmann R, Breivik K, Dachs J, Muir D. Global fate of POPs: current and future research directions. *Environmental pollution*. 2007; 150(1):150–65. doi: [10.1016/j.envpol.2007.06.051](https://doi.org/10.1016/j.envpol.2007.06.051) PMID: [17698265](https://pubmed.ncbi.nlm.nih.gov/17698265/).
59. Berrojalbiz N, Dachs J, Ojeda MJ, Valle MC, Castro-Jiménez J, Wollgast J, et al. Biogeochemical and physical controls on concentrations of polycyclic aromatic hydrocarbons in water and plankton of the Mediterranean and Black Seas. *Global Biogeochemical Cycles*. 2011; 25(4):n/a-n/a. doi: [10.1029/2010gb003775](https://doi.org/10.1029/2010gb003775)
60. Galbán-Malagón C, Berrojalbiz N, Ojeda MJ, Dachs J. The oceanic biological pump modulates the atmospheric transport of persistent organic pollutants to the Arctic. *Nature communications*. 2012; 3:862. doi: [10.1038/ncomms1858](https://doi.org/10.1038/ncomms1858) PMID: [22643889](https://pubmed.ncbi.nlm.nih.gov/22643889/)
61. Galbán-Malagón CJ, Del Vento S, Berrojalbiz N, Ojeda MJ, Dachs J. Polychlorinated biphenyls, hexachlorocyclohexanes and hexachlorobenzene in seawater and phytoplankton from the Southern Ocean (Weddell, South Scotia, and Bellingshausen Seas). *Environmental Science and Technology*. 2013; 47(11):5578–87. doi: [10.1021/es400030q](https://doi.org/10.1021/es400030q) PMID: [23627767](https://pubmed.ncbi.nlm.nih.gov/23627767/)
62. Frouin H, Dangerfield N, Macdonald RW, Galbraith M, Crewe N, Shaw P, et al. Partitioning and bioaccumulation of pcbs and pbdes in marine plankton from the strait of Georgia, British Columbia, Canada. *Progress in Oceanography*. 2013; 115:65–75.
63. Wallberg P, Andersson A. Determination of adsorbed and absorbed polychlorinated biphenyls (PCBs) in seawater microorganisms. *Marine Chemistry*. 1999; 64(4):287–99.
64. Del Vento S, Dachs J. Prediction of uptake dynamics of persistent organic pollutants by bacteria and phytoplankton. *Environmental Toxicology and Chemistry*. 2002; 21(10):2099–107. PMID: [12371486](https://pubmed.ncbi.nlm.nih.gov/12371486/)
65. Echeveste P, Agusti S, Dachs J. Cell size dependent toxicity thresholds of polycyclic aromatic hydrocarbons to natural and cultured phytoplankton populations. *Environmental pollution*. 2010; 158(1):299–307. doi: [10.1016/j.envpol.2009.07.006](https://doi.org/10.1016/j.envpol.2009.07.006) PMID: [19631432](https://pubmed.ncbi.nlm.nih.gov/19631432/).
66. Echeveste P, Dachs J, Berrojalbiz N, Agusti S. Decrease in the abundance and viability of oceanic phytoplankton due to trace levels of complex mixtures of organic pollutants. *Chemosphere*. 2010; 81(2):161–8. doi: [10.1016/j.chemosphere.2010.06.072](https://doi.org/10.1016/j.chemosphere.2010.06.072) PMID: [20673958](https://pubmed.ncbi.nlm.nih.gov/20673958/).
67. Steglich C, Lindell D, Futschik M, Rector T, Steen R, Chisholm SW. Short RNA half-lives in the slow-growing marine cyanobacterium *Prochlorococcus*. *Genome biology*. 2010; 11(5):R54. doi: [10.1186/gb-2010-11-5-r54](https://doi.org/10.1186/gb-2010-11-5-r54) PMID: [20482874](https://pubmed.ncbi.nlm.nih.gov/20482874/); PubMed Central PMCID: PMC2897979.
68. Kelly L, Huang KH, Ding H, Chisholm SW. ProPortal: a resource for integrated systems biology of *Prochlorococcus* and its phage. *Nucleic Acids Res*. 2012; 40(Database issue):D632–40. doi: [10.1093/nar/gkr1022](https://doi.org/10.1093/nar/gkr1022) PMID: [22102570](https://pubmed.ncbi.nlm.nih.gov/22102570/); PubMed Central PMCID: PMC3245167.
69. Morris JJ, Johnson ZI, Szul MJ, Keller M, Zinser ER. Dependence of the cyanobacterium *Prochlorococcus* on hydrogen peroxide scavenging microbes for growth at the ocean's surface. *PLOS ONE*. 2011; 6(2).
70. Mella-Flores D, Mazard S, Humily F, Partensky F, Mahé F, Bariat L, et al. Is the distribution of *Prochlorococcus* and *Synechococcus* ecotypes in the Mediterranean Sea affected by global warming? *Biogeosciences*. 2011; 8(9):2785–804. doi: [10.5194/bg-8-2785-2011](https://doi.org/10.5194/bg-8-2785-2011)
71. Zwirgmaier K, Jardillier L, Ostrowski M, Mazard S, Garczarek L, Vaulot D, et al. Global phylogeography of marine *Synechococcus* and *Prochlorococcus* reveals a distinct partitioning of lineages among oceanic biomes. *Environmental microbiology*. 2008; 10(1):147–61. doi: [10.1111/j.1462-2920.2007.01440.x](https://doi.org/10.1111/j.1462-2920.2007.01440.x) PMID: [17900271](https://pubmed.ncbi.nlm.nih.gov/17900271/).
72. Claustre H, Bricaud A, Babin M, Bruyant F, Guillou L, Le Gall F, et al. Diel variations in *Prochlorococcus* optical properties. *Limnology and Oceanography*. 2002; 47(6):1637–47.
73. Rocap G, Larimer FW, Lamerdin J, Malfatti S, Chain P, Ahlgren NA, et al. Genome divergence in two *Prochlorococcus* ecotypes reflects oceanic niche differentiation. *Nature*. 2003; 424(6952):1042–7. PMID: [12917642](https://pubmed.ncbi.nlm.nih.gov/12917642/)
74. UNECE. Protocol to the 1979 Convention on Long-Range Transboundary Air Pollution on Persistent Organic Pollutants. In: Europe UNECF, editor. Geneva1998.

75. Berrojalbiz N, Dachs J, Del Vento S, Ojeda MJ, Valle MC, Castro-Jimenez J, et al. Persistent organic pollutants in Mediterranean seawater and processes affecting their accumulation in plankton. *Environ Sci Technol*. 2011; 45(10):4315–22. doi: [10.1021/es103742w](https://doi.org/10.1021/es103742w) PMID: [21526777](https://pubmed.ncbi.nlm.nih.gov/21526777/).
76. (EPA). USEPA. <http://www.epa.gov>.
77. Chisholm SW, Olson RJ, Zettler ER, Goericke R, Waterbury JB, Welschmeyer NA. A novel free-living prochlorophyte abundant in the oceanic euphotic zone. *Nature*. 1988; 334(6180):340–3.
78. Lindell D, Jaffe JD, Johnson ZI, Church GM, Chisholm SW. Photosynthesis genes in marine viruses yield proteins during host infection. *Nature*. 2005; 438(7064):86–9. doi: [10.1038/nature04111](https://doi.org/10.1038/nature04111) PMID: [16222247](https://pubmed.ncbi.nlm.nih.gov/16222247/).
79. Pfaffl MW. A new mathematical model for relative quantification in real-time RT-PCR. *Nucleic Acids Research*. 2001; 29(9).
80. Real-time PCR. Dorak MT, editor: Taylor & Francis; 2007.
81. Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. Basic local alignment search tool. *Journal of Molecular Biology*. 1990; 215(3):403–10. PMID: [2231712](https://pubmed.ncbi.nlm.nih.gov/2231712/)
82. Jeffrey SW, Humphrey GF. New Spectrophotometric Equations For Determining Chlorophylls A, B, C1 And C2 In Higher-Plants, Algae And Natural Phytoplankton. *Biochemie Und Physiologie Der Pflanzen*. 1975; 167(2):191–4. WOS:A1975AF61700012.