

RESEARCH ARTICLE

A Network Biology Approach to Denitrification in *Pseudomonas aeruginosa*

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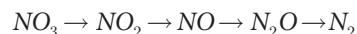
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Abstract

Pseudomonas aeruginosa is a metabolically flexible member of the Gammaproteobacteria. Under anaerobic conditions and the presence of nitrate, *P. aeruginosa* can perform (complete) denitrification, a respiratory process of dissimilatory nitrate reduction to nitrogen gas via nitrite (NO_2), nitric oxide (NO) and nitrous oxide (N_2O). This study focuses on understanding the influence of environmental conditions on bacterial denitrification performance, using a mathematical model of a metabolic network in *P. aeruginosa*. To our knowledge, this is the *first* mathematical model of denitrification for this bacterium. Analysis of the long-term behavior of the network under changing concentration levels of oxygen (O_2), nitrate (NO_3), and phosphate (PO_4) suggests that PO_4 concentration strongly affects denitrification performance. The model provides three predictions on denitrification activity of *P. aeruginosa* under various environmental conditions, and these predictions are either experimentally validated or supported by pertinent biological literature. One motivation for this study is to capture the effect of PO_4 on a denitrification metabolic network of *P. aeruginosa* in order to shed light on mechanisms for greenhouse gas N_2O accumulation during seasonal oxygen depletion in aquatic environments such as Lake Erie (Laurentian Great Lakes, USA). Simulating the microbial production of greenhouse gases in anaerobic aquatic systems such as Lake Erie allows a deeper understanding of the contributing environmental effects that will inform studies on, and remediation strategies for, other hypoxic sites worldwide.

Introduction

Denitrification is a facultative anaerobic process in which nitrate is utilized as an alternative terminal electron receptor and dissimilatory nitrate is reduced to nitrogen gas via nitrogen oxides [1–3].



Since denitrification is one of the few pathways for producing atmospheric N_2 , it is a major component of the nitrogen cycle [4]. Denitrification occurs in several habitats such as soils, lakes, rivers and oceans [5]. Nitrogen fluxes from marine systems to the atmosphere are between 25×10^9 and 179×10^9 kilograms per year via microbial denitrification [6]. *Pseudomonas aeruginosa*, a facultative ubiquitous, and metabolically flexible member of the Gammaproteobacteria, can perform (complete) denitrification under anaerobic conditions and the presence of nitrate. Complete denitrification consists of four sequential steps to reduce nitrate (NO_3^-) to dinitrogen (N_2) via nitrite (NO_2^-), nitric oxide (NO), and nitrous oxide (N_2O), and each step of the pathway is catalyzed by (denitrification) enzymes such as nitrate reductase (*nar*), nitrite reductase (*nir*), nitric oxide reductase (*nor*), and nitrous oxide reductase (*nos*). The identification and transcriptional control of denitrification genes encoding *nar*, *nir*, *nor* and *nos* has been largely established. Transcription is dependent on a hierarchy of the FNR-like Crp family transcription factors Anr and Dnr, the two-component system NarXL, and the CbbQ family protein NirQ [7,8], summarized in [4], allowing for experimental validation of N_2O yield as environmental parameters change.

We have built a combined gene regulatory and metabolic network for the denitrification pathway in *Pseudomonas aeruginosa* PAO1, a well-studied denitrifier strain (Fig. 1). With this study, we hope to shed light on the environmental factors contributing to greenhouse gas N_2O accumulation, of particular interest in Lake Erie (Laurentian Great Lakes, USA). Environments such as Lake Erie experience seasonal periods of hypoxic conditions favorable for

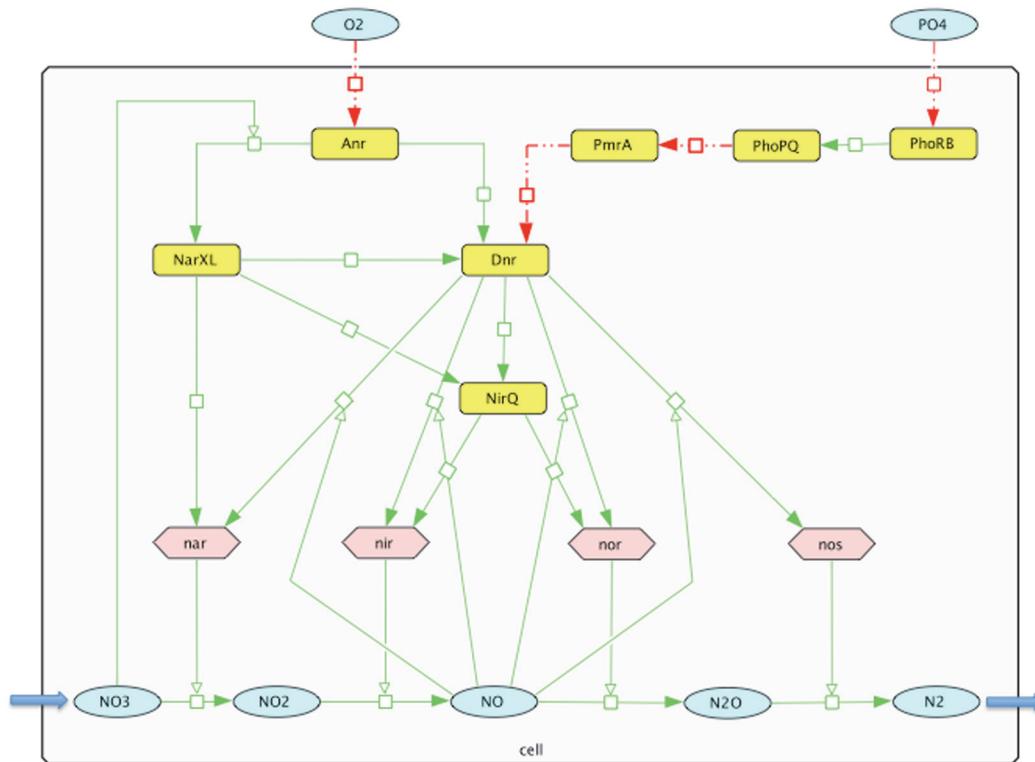


Fig 1. Denitrification regulatory network of *P. aeruginosa*. Green solid arrows indicate upregulation and red dashed arrows indicate downregulation. Model components are PhoPQ, PmrA, Anr, NarXL, Dnr, NirQ, *nar*, *nir*, *nor*, *nos*, NO_2 , NO , N_2O , and N_2 . Our interest lies in perturbation of the external parameters (O_2 , PO_4 , NO_3) and their effect on the long-term behavior of the network.

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denitrification, and the endemic microbial community regulates expression of alternative respiratory pathways to adapt to low oxygen (O_2) tension.

We are interested in using the model to investigate the effect of PO_4 on the denitrification performance of *P. aeruginosa* under anaerobic conditions with high NO_3^- . Although there are several studies on regulation of denitrification by kinetic mathematical modeling approaches (e.g. [9–12]), these attempts are not enough to cover the phenomenon at different levels [2]. One of the challenges in building kinetic mathematical models of networks, such as systems of differential equations, is that many of the needed parameters are either not known or unmeasurable. Furthermore, for large networks, kinetic models are difficult to analyze mathematically. Therefore, we take a qualitative approach to model denitrification distinct from the quantitative denitrification models attempted previously. We use a discrete model framework that provides coarse-grained information about the temporal biochemical output of the network in response to environmental conditions. This framework captures attractors (and their biological correspondence, phenotypes) yet it does not render any measurements of time or concentration. In particular, we prefer a time-discrete and multi-state deterministic framework, Polynomial Dynamical System (PDS) [13], to model our denitrification network in *Pseudomonas aeruginosa*.

Results

The denitrification network consists of molecules, proteins and genes all of which can play an important role in the denitrification process in *Pseudomonas aeruginosa*. Fig. 1 illustrates a static representation of the variables and their regulations. The blue circular nodes are molecules (O_2 , PO_4 , NO_3^- , NO_2 , NO , N_2O , N_2), the yellow rectangular nodes are proteins (PhoRB, PhoPQ, PmrA, Anr, Dnr, NarXL, NirQ) and the pink hexagonal nodes are genes (*nar*, *nir*, *nor*, *nos*) in the network. The large gray rectangle represents the bacterial cell. The regulatory edges between the nodes are either upregulation/activation (green solid arrows) or downregulation/inhibition (red dashed arrows). The pathway begins with the phosphate-sensing two component regulatory system PhoRB [14]. PhoRB, the main PO_4 sensor activating the pho regulon, has been recently shown to be a regulator of PhoPQ transcription in the gammaproteobacterium *Escherichia coli* [15]. In light of the fact that *Pseudomonas aeruginosa* possesses a similar regulatory system to PhoRB in *E. coli* [16], it is appropriate to label the PO_4 -sensing regulatory protein as PhoRB in the denitrification network. In this case, the red dashed arrow from PO_4 to PhoRB means that the availability of phosphate, PO_4 , reduces PhoRB function, and the green arrow from PhoRB to PhoPQ means that PhoPQ is activated by PhoRB. Thus, the availability of PO_4 downregulates PhoPQ via PhoRB. The green solid arrow from Anr (anaerobic regulation of arginine deiminase and nitrate reduction) to NarXL and the green solid arrow from NO_3^- to the arrow between Anr and NarXL indicate that Anr activates NarXL in the presence of NO_3^- . In the same setting, PhoPQ inhibits the expression of PmrA [17]. Low oxygen (O_2) tension, which is the major initial signal to turn on the denitrification pathway, can be sensed by Anr [1]. Under anaerobic conditions, Anr primarily promotes Dnr (dissimilatory nitrate respiration regulator) transcription [4]. The effect of Anr on Dnr can be amplified by NarXL [8]. The mechanism of inhibitory effect of PmrA on Dnr [17] is not known, so we assumed that the effect of Anr on Dnr can be reduced by PmrA. The regulatory protein NirQ, which can be activated by NarXL or Dnr, regulates *nir* and *nor* coordinately to keep the level of NO low because of toxicity of NO [4]. A NO_3^- -responding regulatory protein, NarXL, directly activates *nar*, and indirectly activates *nir* and *nor* via NirQ [4]. The main regulator of the system, Dnr, controls the expression of all denitrification genes (*nar*, *nir*, *nor*, *nos*) in the presence of NO [18]. Of particular note is the influence of the two-component system PhoPQ on PmrA expression and,

subsequently, Dnr expression [17], suggesting that phosphorus (P) availability influences denitrification gene expression (see Fig. 1). This is particularly relevant, since linkages between anaerobic Fe(III) reduction and P release adsorbed to $FeOOH$ in sediments have been recognized for many years [19, 20], and recently documented in Lake Erie by stable isotope methods [21].

The actual mechanisms of the relationships in the denitrification network (Fig. 1) may be quite complex and involve several intermediates. Thus, the network does not represent a biochemical reaction network, for instance, but rather captures the regulatory logic driving the network in a similar way that a circuit diagram explains the function of a circuit board. In the network (Fig. 1), O_2 , PO_4 and NO_3 are *external parameters* and the remaining nodes are *variables*. In the discrete setting that is used to model the denitrification network, each node (e.g. an external parameter O_2 or a variable *nos*) can take up to three states (low, medium, high), and time is implicit and progresses in discrete steps. Our interest lies in perturbation of the external parameters and their effect on the long-term behavior of the variables in the system. S1 Table indicates the discretization values (low/medium/high) for external parameters and nitrogen oxides. Such values incorporate appropriate ranges of long-term nutrient and seasonal oxygen concentrations for Lake Erie [22, 23].

The denitrification network is an open system; it exchanges molecules with the outside environment and responds to external stimuli [24]. The molecule NO_3 enters the bacterium and N_2 exits the system once the system is triggered by low O_2 . The model predicts the long-term behavior of the denitrification pathway under various environmental conditions and these predictions are either supported by the literature or validated by experimental results. Fig. 2 indicates the (predicted) attractors of the system under some possible configuration of the external parameters. There are two conditions that we did not focus on. The low NO_3 and low PO_4 condition and the low NO_3 and high PO_4 condition, while possible, are less likely in freshwaters based on a worldwide survey of lakes revealing that N:P stoichiometric ratios average above the ideal Redfield ratio of 16 [25]. Besides, these conditions would be less relevant to current conditions in Lake Erie, for example, as current measurements of nitrate concentrations (averaging $14\mu M$) typically exceed the K_m (Michaelis-Menten constant) for nitrate-dependent denitrification in *Pseudomonas* spp. (for more information, see [26, 27]). However, a high P, high NO_3 condition can arise in lakes affected by agricultural nutrient inputs and deposition of P in sediments.

- **Prediction 1:** If the concentration levels of O_2 and PO_4 are low, and NO_3 is high, then it is a perfect condition for complete denitrification to N_2 . The model suggests that all variables in the network except PmrA are expected to be high and the bacterium reduces NO_3 to N_2 via nitrogen oxides. This prediction is *supported* by the following studies [1, 4, 8]. In this condition, Anr senses low O_2 and activates NarXL in the presence of NO_3 [4]. Since the effect of Anr on Dnr is amplified by NarXL but is not reduced by PmrA under low PO_4 conditions, Dnr is highly expressed [8]. The main regulator of the system, Dnr, promotes activation of all denitrification genes (*nar*, *nir*, *nor*, *nos*), so NO_3 is reduced to N_2 via NO_2 , NO and N_2O [1].
- **Prediction 2:** If the concentration level of O_2 is low, and PO_4 and NO_3 are high, then the model suggests that all variables except PhoRB-PhoPQ are medium or high. Thus, lower complete denitrification activity to N_2 is expected because the *nar*, *nir* and *nor* levels are high whereas the *nos* level is intermediate. This can cause lower rates of reduction of N_2O to N_2 i.e. higher rates of accumulation of N_2O . These predictions *coincide* with the following studies [8, 17] and experimentation. In this condition, Dnr level is intermediate and induces the expression of denitrification genes (*nar*, *nir*, *nor*, *nos*) due to the fact that the effect of Anr on Dnr is amplified by NarXL and is reduced by PmrA [8, 17]. Moreover, our experimental results in Table 1 show a modest increase in N_2O production with a high PO_4 level. There is

EXTERNAL PARAMETERS			VARIABLES														
O ₂	P _{O4}	N _{O3}	PhoRB	PhoPQ	PmrA	Anr	NarXL	Dnr	NirQ	nar	nir	nor	nos	NO ₂	NO	N ₂ O	N ₂
low	low	high	high	high	low	high	high	high	high	high	high	high	high	high	high	high	high
low	high	high	low	low	high	high	high	medium	high	high	high	high	medium	high	high	high	medium
high	low	low	high	high	low	low	low	low	low	low	low	low	low	low	low	low	low
high	low	high	high	high	low	low	low	low	low	low	low	low	low	low	low	low	low
high	high	low	low	low	high	low	low	low	low	low	low	low	low	low	low	low	low
high	high	high	low	low	high	low	low	low	low	low	low	low	low	low	low	low	low

Fig 2. Steady states of the denitrification network under different environmental conditions. The first condition (low O₂, low P_{O4} and high N_{O3}) corresponds to the perfect condition for denitrification and the second condition (low O₂, high P_{O4} and high N_{O3}) corresponds to the denitrification condition disrupted by P_{O4} availability. The remaining conditions can be labeled as aerobic conditions.

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about a 2-fold increase in N₂O concentration in comparison of the anaerobic *P. aeruginosa* culture with 1.0mM P_{O4} to the anaerobic *P. aeruginosa* culture with 7.5mM P_{O4}. Under these conditions, the culture at 1.0mM P_{O4} is grown under the ideal total N:P ratio of 16 reflecting the 16:1 N:P elemental stoichiometry of aquatic plankton [28]. The cultures grown at elevated P_{O4} (3.0mM and 7.5mM) thus reflect a condition in which P_{O4} is available at surplus levels that repress the PhoRB-dependent gene activation. This is an example of how P_{O4} can influence the expression of denitrification gene, nos, distant from P_{O4} acquisition and subsequently greenhouse gas N₂O accumulation.

- Prediction 3:** If the concentration level of O₂ is high, then, the model suggests that there is no denitrification activity regardless of the values of the other external parameters (P_{O4} or N_{O3}). This prediction is supported by Zumft's extensive review paper, which states that under aerobic conditions, *Pseudomonas aeruginosa* cannot perform denitrification because Anr cannot activate the main regulator of the system, Dnr, in the presence of oxygen [1].

[Fig 2](#) indicates the attractors of the system under different environmental condition. These attractors indeed are steady states, each of which corresponds to one environmental condition. This agrees with biology; Palsson highlighted that open systems eventually reach a (homeostatic) steady state and are in balance with their environment until the environmental conditions are perturbed [24]. Phenotypes, biological interpretations of the long-term behavior (steady states), of the system under various environmental conditions can be found in [Table 2](#). Based

Table 1. Nitrous oxide concentration in *P. aeruginosa* cultures grown in glucose minimal medium at varying phosphate concentrations, normalized to 10⁸ cells.

Culture (mM P _{O4})	[N ₂ O] ppm, 24 h	[N ₂ O] ppm, 72 h
1.0 mM	760.3 +/- 109.3	813.8 +/- 52.1
3.0 mM	856.0 +/- 121.5	872.3 +/- 63.3
7.5 mM	1484.0 +/- 146.2	1786 +/- 98.0

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Table 2. Biological interpretation of the steady states of the system under different environmental conditions.

O ₂	P _{O4}	N _{O3}	BIOLOGICAL INTERPRETATION
low	low	high	high denitrification performance
low	high	high	low denitrification performance
high	low	low	no denitrification
high	low	high	no denitrification
high	high	low	no denitrification
high	high	high	no denitrification

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on the steady state analysis above, the *Pseudomonas* network model predicts that elevated PO_4 , hypothesized to increase under hypoxia, acts to modulate the transcriptional network to limit *nos* gene expression. Thus, the physiological output under this condition will be an increased yield of N_2O relative to N_2 . Given the prediction that increased PO_4 will influence the N_2O yield, our experimental results thus far indicate that PO_4 availability modestly, but significantly increases N_2O yield in this model species (ANOVA $p = 0.012$; [Table 1](#)). While other studies have suggested linkages between N_2O accumulation and factors such as *nosZ* vs. *nirS/K* abundance [29, 30], *nirS* (heme dependent nitrite reductase) genetic diversity [31], or soil pH [32], the data presented here are the first to suggest a role for PO_4 in regulating the denitrification pathway. Given the elevated PO_4 release from $FeOOH$ complexes following sedimentary anaerobic Fe(III) reduction [19, 20], hypoxia may yield a high P, high NO_3^- condition that enhances N_2O production.

Discussion

In an aquatic system, oxygen dissolves in water to be available to living aerobic organisms. Hypoxia is the phenomenon of dissolved oxygen below 4 mg O₂ per liter. Common reasons for hypoxia include aerobic respiration of decaying algal biomass from bloom events. Such blooms are fueled by increased availability of N and P due to anthropogenic inputs such as agricultural runoff and industrial pollutants [33]. The linkage between high nutrient (N, P) loads and N losses (N_2 and N_2O) through dissimilatory anaerobic processes was described recently [34]. Hypoxic (low-oxygen) areas, so-called dead zones, often occur in several large bodies of water affected by human activity, including Lake Erie, which is of particular interest. Establishing a better understanding of the nutrient cycling of Lake Erie has quite wide ranging socioeconomic impacts on its recreational area and economy, primarily fisheries. Through denitrification, dead zones lead to microbial production of the greenhouse gas nitrous oxide (N_2O), which plays a crucial role in ozone layer depletion and climate change. Simulating the microbial production of greenhouse gases in anaerobic aquatic systems such as Lake Erie allows a deeper understanding of the contributing environmental effects that will inform studies on, and remediation strategies for, other hypoxic sites worldwide. During hypoxia, the denitrification rate in Lake Erie is about $150 \mu\text{mol} N_2 m^{-2} h^{-1}$ [35]. In addition to oxygen, the intersections of the nitrogen cycle with other geochemical cycles may be important factors influencing denitrification and nitrogen (N) sinks in aquatic systems. In particular, the increased availability of phosphorus (P) has been shown to dictate the rate of nitrogen removal in aquatic systems [34]. Indeed, the transcriptional regulatory network developed for *P. aeruginosa* indicates that bioavailable phosphate (PO_4) is an environmental factor that should be considered.

The bacterium *Pseudomonas aeruginosa* is an example of an abundant microbe in aquatic systems [36], and analysis of Lake Erie metagenomic data sets reveals abundant pseudomonads capable of denitrification (Unpublished data, DOE-JGI). This study describes a computational model of a denitrification network of this bacterium to capture the effect of PO_4 on its denitrification performance in order to shed light on greenhouse gas N_2O accumulation during oxygen depletion. To our knowledge, this is the *first* mathematical model of denitrification for this bacterium. Transcription is dependent on a hierarchy of the FNR-like Crp family transcription factors Anr and Dnr, the two-component system NarXL, and the CbbQ family protein NirQ [7, 8, 37], allowing for experimental measurement of N_2O as external (environmental) parameters change. The model was constructed based on the pertinent biological literature. Model predictions either agree with current published results or are validated by experimentation. The new biology that our model discovers is that PO_4 availability strongly affects the denitrification activity of *P. aeruginosa* under anaerobic conditions and the presence of nitrate; high PO_4 can

cause less N_2O reduction to N_2 during denitrification. The data presented here are the first to suggest a role for PO_4 in regulating the denitrification pathway in *Pseudomonas aeruginosa*.

Current efforts will be expanded to determine how PO_4 affects greenhouse gas N_2O accumulation during denitrification in *P. aeruginosa*. According to the model, the activation of Dnr by Anr or the activation of nos in the presence of NO by Dnr can be prevented by high PO_4 . These hypotheses will be tested utilizing quantitative reverse transcriptase PCR (qRT-PCR) to determine Dnr, *norB* (nitric oxide reductase large subunit gene) and *nosZ* (encoding nitrous oxide reductase) transcript levels in denitrifying cultures grown in increasing *P.* Synergistic interactions between individual members of population of *Pseudomonas aeruginosa* may need to be taken into account and incorporated to the model. For instance, Toyofuku and his colleagues stated that denitrification performance of *P. aeruginosa* does not only depend upon activation of denitrification genes (*nar*, *nir*, *nor*, *nos*) but also cell-cell communications under denitrifying conditions [38].

The model described here works well for cultured *Pseudomonas*, and the next step is to test natural complex microbial communities from different denitrification sites. The effects of PO_4 on N_2O production will be tested in mesocosms of hypoxic Lake Erie water samples to see if the model described here predicts the community as a whole. By testing the model on environmental samples in mesocosms from Lake Erie and elsewhere, the study can likely be applied broadly to other marine dead zones such as those that routinely occur in the Gulf of Mexico.

Materials and Methods

Computational Methods

Our network consists of two different sub-networks (metabolic and gene regulatory) and consequently different time scales. From a discrete modeling perspective, this issue can be tackled or ignored only if the long-term behavior of the system is of interest. One could address this issue either (1) using a stochastic framework such as Stochastic Discrete Dynamical System (SDDS) [39] if how fast/slow the reactions are in the network are known/inferred out of a time-course experimental data or (2) introducing time delays by an asynchronous update schedule. Due to inadequate information on the reaction rates, we do not focus on a stochastic framework. Even with a fully asynchronous update schedule, the attractors are preserved for each configuration of external parameters; however, this asynchronous update schedule requires more time steps to reach a steady state than a synchronous update schedule does. Since an asynchronous update schedule provides us more on transient behavior of the system and we are interested in long-term behavior of the system, we prefer to use a deterministic framework with a synchronous update schedule, Polynomial Dynamical System (PDS), which allows us to model regulatory networks over a finite field [13].

Definition 1 Let x_1, x_2, \dots, x_n be variables which can take values in finite fields X_1, X_2, \dots, X_n respectively. Let $\mathbf{X} = X_1 \times \dots \times X_n$ be the Cartesian product. For each $i = 1, 2, \dots, n$, we define $f_i: \mathbf{X} \rightarrow X_i$ which is an update function that describes the regulation of x_i through interaction with other variables in the system. A Polynomial Dynamical System is a collection of n update functions

$$f = (f_1, f_2, \dots, f_n) : \mathbf{X} \rightarrow \mathbf{X}$$

In the model, all external parameters (O_2, PO_4, NO_3) and some variables (PhoPQ, PmrA, Anr, NarXL) are Boolean (low or high), and other variables are ternary (low, medium or high). There are 14 variables, each of which is labeled for the mathematical formulation. [Table 3](#) indicates the variables, their discretization, update rules and the literature evidence that support these update rules. Inflow substances (i.e. external parameters: O_2, PO_4, NO_3) in this model

Table 3. Summary of the model variables, their discretization, update rules and supportive argument. The update rules with an asterix (*) means this update rule is very close to the biological correspondence but not quite. The transition tables of the variables having update rules with an asterix (*) can be found in the Supplementary material.

Index	Variables	Discretization	Update Rules	Literature Evidence
1	PhoRB	Boolean	$\text{NOT}(PO_4)$	PhoRB is a phosphate-sensing two component regulatory system [14]
2	PhoPQ	Boolean	PhoRB	PhoPQ regulates acid phosphatase under P starvation [48, 49], and PhoB regulates PhoQ transcription [15]
3	PmrA	Boolean	$\text{NOT}(\text{PhoPQ})$	PhoPQ inhibits the expression of PmrA [17]
4	Anr	Boolean	$\text{NOT}(O_2)$	Low oxygen (O_2) tension can be sensed by Anr [1]
5	NarXL	Boolean	$\text{MIN}(\text{Anr}, NO_3)$	Anr activates NarXL in the presence of NO_3 [4]
6	Dnr	Ternary	$*\text{MIN}(\text{Anr}, \text{MAX}(\text{NarXL}, \text{NOT}(\text{PmrA})))$	The effect of Anr on Dnr can be reduced by PmrA (model assumption) and Anr and NarXL cooperatively activate <i>dnr</i> [8, 17]
7	NirQ	Ternary	$*\text{MAX}(\text{NarXL}, \text{Dnr})$	NirQ can be activated by NarXL or Dnr [4]
8	<i>nar</i>	Ternary	$*\text{MIN}(\text{NarXL}, \text{MIN}(\text{Dnr}, \text{NO}))$	<i>nar</i> is directly activated by NarXL and Dnr in the presence of NO [4, 18]
9	<i>nir</i>	Ternary	$\text{MAX}(\text{NirQ}, \text{MIN}(\text{Dnr}, \text{NO}))$	<i>nir</i> is activated by NirQ or Dnr in the presence of NO [4, 18]
10	<i>nor</i>	Ternary	$\text{MAX}(\text{NirQ}, \text{MIN}(\text{Dnr}, \text{NO}))$	<i>nor</i> is activated by NirQ or Dnr in the presence of NO [4, 19]
11	<i>nos</i>	Ternary	$\text{MIN}(\text{Dnr}, \text{NO})$	<i>nos</i> is activated by Dnr in the presence of NO [18]
12	NO_2	Ternary	$*\text{MIN}(\text{NO}_3, \text{nar})$	NO_3 is reduced to NO_2 by <i>nar</i> [1]
13	NO	Ternary	$\text{MIN}(\text{NO}_2, \text{nir})$	NO_2 is reduced to NO by <i>nir</i> [1]
14	N_2O	Ternary	$\text{MIN}(\text{NO}, \text{nor})$	NO is reduced to N_2O by <i>nor</i> [1]
15	N_2	Ternary	$\text{MIN}(\text{N}_2\text{O}, \text{nos})$	N_2O is reduced to N_2 by <i>nos</i> [1]

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give inputs to variables and are involved in the update rules. They do not have update rules because not only they do not have regulators but also we are interested in analyzing the long-term behavior of the model under different configurations of them. The model has only one outflow substance, N_2 , whose regulation depends upon the greenhouse gas N_2O and its reductase, *nos*.

Based on the literature, we formulate the regulation of the variables with MIN, MAX and NOT, which correspond to AND, OR and NOT in a Boolean setting. The following are examples for how the update rules are decided:

- An update rule of NarXL can be defined as “ $\text{MIN}(\text{Anr}, NO_3)$ ” because NarXL is activated by Anr only in the presence of NO_3 , i.e. both Anr and NO_3 need to be high for NarXL regulation.
- An update rule of *nir* can be labeled as “ $\text{MAX}(\text{NirQ}, \text{MIN}(\text{Dnr}, \text{NO}))$ ” due to the fact that *nir* is activated by NirQ or Dnr in the presence of NO.
- An update rule of PhoRB can be “ $\text{NOT}(PO_4)$ ” since PO_4 downregulates PhoRB, i.e. one is low when another is high.

From the update rules in Table 3, for each network variable, we constructed a corresponding transition table, which describes how a specific variable responds to different configurations of their regulators. Although the regulations for most variables can be formulated by MIN, MAX and/or NOT, the regulations of a few variables are very close to some formulation but not quite. For the sake of consistency with biology, we decided to slightly modify the transition table of Dnr, NirQ, *nar* and NO_2 , whose update rules are marked with an asterix (*) in Table 3. The transition tables of these variables and more explanation on why the changes were necessary can be found in S2 Table, S3 Table, S4 Table and S5 Table respectively.

Besides, if the variable takes three states (low, medium, high), the current state of the variable is included its own transition table. This does not mean autoregulation/self-regulation; but it is to prevent the variable from jumps between the low (0) state and the high (2) state at the next time step. In other words, including the current state of a ternary variable in its transition table provides a smooth transition among its own states. On the other hand, such jumps cannot occur in a Boolean variable.

After constructing a transition table for each variable x_i , an update function can be obtained by interpolating its transition table using the polynomial form:

$$f_i(x) = \sum_{(c_{i_1}, \dots, c_{i_r}) \in F_p^r} f_i(c_{i_1}, \dots, c_{i_r}) \prod_{j \in \{i_1, \dots, i_r\}} (1 - (x_j - c_j)^{p-1}) \pmod{p} \quad (1)$$

where $\mathbf{x} = (x_{i_1}, \dots, x_{i_r})$ is a vector; c_{i_1}, \dots, c_{i_r} are the values of the variables x_{i_1}, \dots, x_{i_r} , which affect the update of x_i in the transition table of x_i ; $f_i(c_{i_1}, \dots, c_{i_r})$ is the value in the last column of the transition table of x_i ; p is the maximum (prime) number of the different discrete values that all variables can take on [40]. In our model, all computations were done in modulo 3.

After having all update functions (see S1 Text), we computed the basin of attraction of the whole system under the environmental conditions of interest (see Fig. 2). For model construction and steady state analysis, we used customized Ruby and Perl scripts, which are a part of the source code of Analysis of Dynamic Algebraic Models (ADAM, available at <http://adam.plantsimlab.org/>), a free of charge web-tool to analyze the dynamics of discrete biological systems [41].

Experimental Methods

Pseudomonas aeruginosa PAO1 cultures were grown in stoppered 20mL serum vials containing glucose minimal medium [42] supplemented with 110mM glucose and 16mM nitrate (NO_3^-). Phosphate (PO_4^{3-}) concentration varied from 1.0mM to 7.5mM, and triplicate culture vials were sampled for headspace gases at 24h and 72h post-inoculation. Gases were dispensed into evacuated exetainers and assayed for nitrous oxide by gas chromatography. Gas production was normalized to cell counts obtained by flow cytometry of culture fluids.

Supporting Information

S1 Table. Discretization of external parameters and nitrogen oxides. Information in the table was obtained from [43–47]
(XLS)

S2 Table. Transition table of Dnr.
(XLS)

S3 Table. Transition table of NirQ.
(XLS)

S4 Table. Transition table of nar.
(XLS)

S5 Table. Transition table of NO₂.
(XLS)

S1 Text. Update functions of all variables in the denitrification network.
(TXT)

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Author Contributions

Conceived and designed the experiments: GSB. Performed the experiments: GSB. Analyzed the data: SA GSB. Contributed reagents/materials/analysis tools: SA RL. Wrote the paper: SA GSB RL.

References

1. Zumft WG (1997) Cell biology and molecular basis of denitrification. *Microbiol Mol Biol Rev* 61: 533–616.
2. Groffman P, Davidson E, S S (2009) New approaches to modeling denitrification. *Biogeochemistry* 93: 1–5. doi: [10.1007/s10533-009-9285-0](https://doi.org/10.1007/s10533-009-9285-0)
3. Fennel K, Brady D, DiToro D, Fulweiler R, Gardner W, et al. (2009) Modeling denitrification in aquatic sediments. *Biogeochemistry* 93: 159–178. doi: [10.1007/s10533-008-9270-z](https://doi.org/10.1007/s10533-008-9270-z)
4. Arai H (2011) Regulation and function of versatile aerobic and anaerobic respiratory metabolism in *pseudomonas aeruginosa*. *Front Microbiol* 2: 103. doi: [10.3389/fmicb.2011.00103](https://doi.org/10.3389/fmicb.2011.00103) PMID: [21833336](https://pubmed.ncbi.nlm.nih.gov/21833336/)
5. Seitzinger S, Harrison JA, Bohlke JK, Bouwman AF, Lowrance R, et al. (2006) Denitrification across landscapes and waterscapes: a synthesis. *Ecological Applications* 16: 2064–2090. doi: [10.1890/1051-0761\(2006\)016%5B2064:DALAWA%5D2.0.CO;2](https://doi.org/10.1890/1051-0761(2006)016%5B2064:DALAWA%5D2.0.CO;2) PMID: [17205890](https://pubmed.ncbi.nlm.nih.gov/17205890/)
6. Kupchella CE, Hyland MC (1989) Environmental science: Living within the system of nature. *Ecology*.
7. Arai H, Mizutani M, Igarashi Y (2003) Transcriptional regulation of the nos genes for nitrous oxide reductase in *pseudomonas aeruginosa*. *Microbiology* 149: 29–36. doi: [10.1099/mic.0.25936-0](https://doi.org/10.1099/mic.0.25936-0) PMID: [12576577](https://pubmed.ncbi.nlm.nih.gov/12576577/)
8. Schreiber K, Krieger R, Benkert B, Eschbach M, Arai H, et al. (2007) The anaerobic regulatory network required for *pseudomonas aeruginosa* nitrate respiration. *J Bacteriol* 189: 4310–4. doi: [10.1128/JB.00240-07](https://doi.org/10.1128/JB.00240-07) PMID: [17400734](https://pubmed.ncbi.nlm.nih.gov/17400734/)
9. Kampschreur MJ, Kleerebezem R, Picoreanu C, Bakken L, Bergaust L, et al. (2012) Metabolic modeling of denitrification in *agrobacterium tumefaciens*: a tool to study inhibiting and activating compounds for the denitrification pathway. *Front Microbiol* 3: 370. doi: [10.3389/fmicb.2012.00370](https://doi.org/10.3389/fmicb.2012.00370) PMID: [23087683](https://pubmed.ncbi.nlm.nih.gov/23087683/)
10. Jayakumar A, O'Mullan GD, Naqvi SW, Ward BB (2009) Denitrifying bacterial community composition changes associated with stages of denitrification in oxygen minimum zones. *Microb Ecol* 58: 350–62. doi: [10.1007/s00248-009-9487-y](https://doi.org/10.1007/s00248-009-9487-y) PMID: [19238477](https://pubmed.ncbi.nlm.nih.gov/19238477/)
11. Mampaey KE, Beuckels B, Kampschreur MJ, Kleerebezem R, van Loosdrecht MC, et al. (2013) Modelling nitrous and nitric oxide emissions by autotrophic ammonia-oxidizing bacteria. *Environ Technol* 34: 1555–66. doi: [10.1080/09593330.2012.758666](https://doi.org/10.1080/09593330.2012.758666) PMID: [24191490](https://pubmed.ncbi.nlm.nih.gov/24191490/)
12. Pan Y, Ni BJ, Yuan Z (2013) Modeling electron competition among nitrogen oxides reduction and n₂O accumulation in denitrification. *Environ Sci Technol* 47: 11083–11091. doi: [10.1021/es402348n](https://doi.org/10.1021/es402348n) PMID: [24001217](https://pubmed.ncbi.nlm.nih.gov/24001217/)
13. Laubenbacher R, Stigler B (2004) A computational algebra approach to the reverse engineering of gene regulatory networks. *J Theor Biol* 229: 523–37. doi: [10.1016/j.jtbi.2004.04.037](https://doi.org/10.1016/j.jtbi.2004.04.037) PMID: [15246788](https://pubmed.ncbi.nlm.nih.gov/15246788/)
14. Wanner BL (1993) Gene regulation by phosphate in enteric bacteria. *J Cell Biochem* 51: 47–54. doi: [10.1002/jcb.240510110](https://doi.org/10.1002/jcb.240510110) PMID: [8432742](https://pubmed.ncbi.nlm.nih.gov/8432742/)
15. Marzan LW, Hasan CM, Shimizu K (2013) Effect of acidic condition on the metabolic regulation of *escherichia coli* and its phob mutant. *Arch Microbiol* 195: 161–71. doi: [10.1007/s00203-012-0861-7](https://doi.org/10.1007/s00203-012-0861-7) PMID: [23274360](https://pubmed.ncbi.nlm.nih.gov/23274360/)

16. Filloux A, Bally M, Soscia C, Murgier M, Lazdunski A (1988) Phosphate regulation in *pseudomonas aeruginosa*: cloning of the alkaline phosphatase gene and identification of phob- and phor-like genes. Mol Genet 212: 510–3. doi: [10.1007/BF00330857](https://doi.org/10.1007/BF00330857) PMID: [3138529](#)
17. Galan-Vasquez E, Luna B, Martinez-Antonio A (2011) The regulatory network of *pseudomonas aeruginosa*. Microb Inform Exp 1: 3. doi: [10.1186/2042-5783-1-3](https://doi.org/10.1186/2042-5783-1-3) PMID: [22587778](#)
18. Giardina G, Rinaldo S, Johnson KA, Di Matteo A, Brunori M, et al. (2008) No sensing in *pseudomonas aeruginosa*: structure of the transcriptional regulator dnr. J Mol Biol 378: 1002–15. doi: [10.1016/j.jmb.2008.03.013](https://doi.org/10.1016/j.jmb.2008.03.013) PMID: [18420222](#)
19. Mortimer CH (1941) The exchange of dissolved substances between mud and water in lakes. J Ecol: 280–329. doi: [10.2307/2256395](https://doi.org/10.2307/2256395)
20. Mortimer CH (1942) The exchange of dissolved substances between mud and water in lakes. J Ecol: 147–201. doi: [10.2307/2256691](https://doi.org/10.2307/2256691)
21. Elsbury KE, Paytan A, Ostrom N, Kendall C, Young MB, et al. (2009) Using oxygen isotopes of phosphate to trace phosphorus sources and cycling in lake erie. Environ Sci Technol: 3108–3114. doi: [10.1021/es8034126](https://doi.org/10.1021/es8034126) PMID: [19534121](#)
22. Makarewicz J, Bertram P, TW L (2000) Chemistry of the offshore surface waters of lake erie: Pre- and post-dreissena introduction (1983–1993). J Great Lakes Res 26: 82–93. doi: [10.1016/S0380-1330\(00\)70675-7](https://doi.org/10.1016/S0380-1330(00)70675-7)
23. Wilhelm S, Bullerjahn G, Eldridge M, Rinta-Kanto J, Poorvin L, et al. (2006) Seasonal hypoxia and the genetic diversity of prokaryote populations in the central basin hypolimnion of lake erie: evidence for abundant cyanobacteria and photosynthesis. J Great Lakes Res 32: 657–671. doi: [10.3394/0380-1330\(2006\)32%5B657:SHATGD%5D2.0.CO;2](https://doi.org/10.3394/0380-1330(2006)32%5B657:SHATGD%5D2.0.CO;2)
24. Palsson B (2011) Systems Biology: Simulation of Dynamic Network States. Cambridge University Press.
25. Hecky RE, Campbell P, Hendzel LL (1993) The stoichiometry of carbon, nitrogen and phosphorus in particulate matter of lakes and oceans. Limnol Oceanogr 38: 709–724. doi: [10.4319/lo.1993.38.4.0709](https://doi.org/10.4319/lo.1993.38.4.0709)
26. Almeida JS, Reis MAM, Carrondo MJT (1995) Competition between nitrate and nitrite reduction in denitrification by *pseudomonas fluorescens*. Biotechnol Bioeng 46: 476484.
27. Bootsma H (2006) Naturally occurring chemicals: Nutrients. State of the Lakes Ecosystem Conference, Milwaukee, WI (1–3 November, 2006).
28. Loladze I, Elser JJ (2011) The origins of the redfield nitrogen-to-phosphorus ratio are in a homoeo-static protein-to-rRNA ratio. Ecol Lett 14: 244–50. doi: [10.1111/j.1461-0248.2010.01577.x](https://doi.org/10.1111/j.1461-0248.2010.01577.x) PMID: [21244593](#)
29. Philippot L, uhel J, Saby NPA, Chneby D, Chrokov A, et al. (2009) Mapping field-scale spatial patterns of size and activity of the denitrifier community. Environ Microbiol 11: 1518–1526. doi: [10.1111/j.1462-2920.2009.01879.x](https://doi.org/10.1111/j.1462-2920.2009.01879.x) PMID: [19260937](#)
30. Philippot L, Andert J, Jones C, D B, S H (2011) Importance of denitrifiers lacking the genes for nitrous oxide reductase for n₂O emissions from soil. Glob Change Biol 17: 1497–1504. doi: [10.1111/j.1365-2486.2010.02334.x](https://doi.org/10.1111/j.1365-2486.2010.02334.x)
31. Salles JF, Le Roux X, Poly F (2012) Relating phylogenetic and functional diversity among denitrifiers and quantifying their capacity to predict community functioning. Frontiers in Microbiology 3. doi: [10.3389/fmicb.2012.00209](https://doi.org/10.3389/fmicb.2012.00209)
32. Bakken LR, Bergaust L, Liu B, Frostegrd A (2011) Regulation of denitrification at the cellular level: a clue to the understanding of n₂O emissions from soils. Phil Trans R Soc B 367: 1226–1234. doi: [10.1098/rstb.2011.0321](https://doi.org/10.1098/rstb.2011.0321)
33. Dybas CL (2005) Dead zones spreading in world oceans. Bioscience 55: 552–557. doi: [10.1641/0006-3568\(2005\)055%5B0552:DZSIWO%5D2.0.CO;2](https://doi.org/10.1641/0006-3568(2005)055%5B0552:DZSIWO%5D2.0.CO;2)
34. Finlay J, Small G, Sterner R (2013) Human influences on nitrogen removal in lakes. Science 342: 247–250. doi: [10.1126/science.1242575](https://doi.org/10.1126/science.1242575) PMID: [24115440](#)
35. McCarthy M, Gardner W, Lavrentyev P, Moats K, Jochem F, et al. (2007) Effects of hydrological flow regime on sediment-water interface and water column nitrogen dynamics in a great lakes coastal wetland (old woman creek, lake erie). Journal of Great Lakes Research 33: 219–231. doi: [10.3394/0380-1330\(2007\)33%5B219:EOHFRO%5D2.0.CO;2](https://doi.org/10.3394/0380-1330(2007)33%5B219:EOHFRO%5D2.0.CO;2)
36. Romling U, Wingender Y, Muller H, Tummler B (1994) A major *pseudomonas aeruginosa* clone common to patients and aquatic habitats. Applied and Environmental Microbiology 60: 1734–1738. PMID: [8031075](#)
37. Rodionov DA, Dubchak IL, Arkin AP, Alm EJ, Gelfand MS (2005) Dissimilatory metabolism of nitrogen oxides in bacteria: comparative reconstruction of transcriptional networks. PLoS Comput Biol 1: e55. doi: [10.1371/journal.pcbi.0010055](https://doi.org/10.1371/journal.pcbi.0010055) PMID: [16261196](#)

38. Toyofuku M, Uchiyama H, Nomura N (2012) Social behaviours under anaerobic conditions in *pseudomonas aeruginosa*. *Int J Microbiol* 2012: 405191. doi: [10.1155/2012/405191](https://doi.org/10.1155/2012/405191) PMID: [22518142](#)
39. Murruigarra D, Veliz-Cuba A, Aguilar B, Arat S, Laubenbacher R (2012) Modeling stochasticity and variability in gene regulatory networks. *EURASIP J Bioinform Syst Biol* 2012: 5. doi: [10.1186/1687-4153-2012-5](https://doi.org/10.1186/1687-4153-2012-5) PMID: [22673395](#)
40. Veliz-Cuba A, Jarrah AS, Laubenbacher R (2010) Polynomial algebra of discrete models in systems biology. *Bioinformatics* 26: 1637–43. doi: [10.1093/bioinformatics/btq240](https://doi.org/10.1093/bioinformatics/btq240) PMID: [20448137](#)
41. Hinkelmann F, Brandon M, Guang B, McNeill R, Blekherman G, et al. (2011) Adam: Analysis of discrete models of biological systems using computer algebra. *BMC Bioinformatics* 12: 295. doi: [10.1186/1471-2105-12-295](https://doi.org/10.1186/1471-2105-12-295) PMID: [21774817](#)
42. Stanier RY, Palleroni NJ, Doudoroff M (1966) The aerobic pseudomonads: a taxonomic study. *J Gen Microbiol* 43: 159–271. doi: [10.1099/00221287-43-2-159](https://doi.org/10.1099/00221287-43-2-159)
43. Fujiwara T, Fukumori Y (1996) Cytochrome cb-type nitric oxide reductase with cytochrome c oxidase activity from paracoccus denitrificans atcc 35512. *J Bacteriol* 178: 1866–71. PMID: [8606159](#)
44. Hulse CL, Tiedje JM, Averill BA (1988) A spectrophotometric assay for dissimilatory nitrite reductases. *Anal Biochem* 172: 420–6. doi: [10.1016/0003-2697\(88\)90464-2](https://doi.org/10.1016/0003-2697(88)90464-2) PMID: [2847585](#)
45. Kristjansson JK, Hollocher TC (1980) First practical assay for soluble nitrous oxide reductase of denitrifying bacteria and a partial kinetic characterization. *J Biol Chem* 255: 704–7. PMID: [7356639](#)
46. US Environmental Protection Agency. <http://www.epa.gov/glindicators/water/phosphorusb.html>. Accessed 03/03/2014.
47. US Environmental Protection Agency. <http://www.epa.gov/grtlakes/monitoring/limnology/index.html>. Accessed 03/03/2014.
48. Kier LD, Weppelman RM, Ames BN (1979) Regulation of nonspecific acid phosphatase in salmonella: phon and phop genes. *J Bacteriol* 138: 155–61. PMID: [374361](#)
49. Groisman EA, Heffron F, Solomon F (1992) Molecular genetic analysis of the escherichia coli phop locus. *J Bacteriol* 174: 486–91. PMID: [1530848](#)