

Comparative Demography of Skates: Life-History Correlates of Productivity and Implications for Management

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Abstract

Age-structured demographic models were constructed based on empirical estimates of longevity and maturity for five deepwater Bering Sea skates to investigate how observed differences in life history parameters affect population growth rates. Monte Carlo simulations were used to incorporate parameter uncertainty. Estimated population growth rates ranged from 1.045 to 1.129 yr⁻¹ and were lower than those reported for other Alaskan skates and most chondrichthyans. Population growth rates of these and other high-latitude skates increased with relative reproductive lifespan, but displayed no significant relationship with body size or depth distribution, suggesting that assemblage shifts may be difficult to predict for data-poor taxa. Elasticity analyses indicated that juvenile and adult survival had greater per-unit effects on population growth rates than did egg-case survival or fecundity. Population growth rate was affected more by uncertainty in age at maturity than maximum age. The results of this study indicate that if skates are deemed to be a management concern, gear modifications or depth-specific effort controls may be effective.

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Introduction

Although it has long been recognized that many elasmobranchs have life histories that may hinder the sustainability of large-scale fisheries [1], variability in productivity within and among elasmobranch taxa has only begun to be rigorously quantified. Of these taxa, it could be argued that skates have received the least research attention. However, concern for the appropriate management of skates has increased in the last decade as declines of skate populations have been detected in response to both direct and indirect fishing pressure [2,3,4,5]. The emerging body of literature on skates indicates that species vary widely in life-history traits [6], which further complicates the assessment and management of mixed-species fisheries and highlights the need for species-specific data.

Eleven known species of the genus *Bathyraja* (Chondrichthyes: Rajiformes: Arhychobatidae) occur on the eastern Bering Sea (hereafter EBS) continental slope [7]. Although none of these species are currently targeted by fisheries, annual skate catches averaged between 56 and 75% of the 'other species' category taken as bycatch in Bering Sea and Aleutian Islands commercial bottom-trawl and longline fisheries from 1992 to 2009 [8]. Skates are considered a low-value catch, but since 1998 there has been a sharp increase in the market price of large-bodied skates, with a corresponding increase in the proportion of large-bodied skate species retained for processing [9]. Individuals that are not retained are not likely to survive because estimates of discard

mortality are high (\sim 41–45%) [10,11]. In addition, a pilot fishery targeting two large-bodied skate species, *Beringraja binoculata* and *Raja rhina*, has recently developed in the Gulf of Alaska [9], despite concerns that the available life-history data are insufficient to properly manage the fishery [12]. Given the current state of biological knowledge about the system, there is limited evidence to predict the population- and assemblage-level effects of fishing mortality on EBS skates.

Despite the lack of targeted take, it is plausible that current rates of indirect catch in commercial longline and trawl fisheries could affect the structure of the EBS skate assemblage. Recent research indicates that Alaskan skate species vary widely in longevity and age at maturity [13,14,15,16,17,18], indicating high potential for community-level shifts. Evidence of directed shifts in heavilyexploited rajid assemblages has been documented in several regions, with smaller and more productive skate species becoming more dominant as larger species are disproportionately removed from communities [2,3,19]. In the northeast Atlantic, taxonomically-aggregated skate catch trends from the last 40 years displayed an unexpected level of consistency that disguised considerable changes in species-specific abundance [3]. Similarly, changes in EBS assemblage structure may be difficult to detect; prior to recent concerted efforts to improve the species-specificity of bycatch data in Alaska's groundfish fisheries, approximately one-third of the observed skate catch (% by weight) reported from 2004 to 2008 was identified only as "Bathyraja sp." [9].

Until recently, there were few data available to estimate speciesspecific life-history parameters for skates in the EBS. In the past five years, researchers have produced age, growth, and maturity estimates for eight species within the assemblage [14,15,16,17,20,21,22,23]. Ebert et al. [14] used these parameter estimates to calculate population growth rates for four skate species that occur primarily on the shelf and upper slope in the EBS; however, until now, no estimates of population growth have been reported for inhabitants of the deeper portions of the slope. It is reasonable to expect that these deepwater species have lower population growth rates, given that this appears to be a general pattern within higher taxonomic groups of elasmobranchs [24.25] and that longevity seems to scale with depth among members of the Alaskan skate assemblage [22].

Here, we aim to detect among-species trends in productivity within the EBS skate assemblage. In addition, we investigate biological and ecological correlates to population growth rates to help assess the relative productivity of species currently lacking lifehistory data. Towards these objectives, we used Leslie matrix models to estimate population growth rates for five slope-dwelling members of the EBS skate assemblage for which age, growth, and reproductive data have recently been reported: Bathyraja lindbergi, B. maculata, B. taranetzi [23], B. minispinosa [20], and B. trachura [22]. To further explore taxonomic trends in productivity, we compare the population growth rates generated from these models to those of other EBS skates, skates from other regions, and less closely related elasmobranchs. These results will provide managers with some intuition of the likely direct and indirect effects of management actions by identifying which biological processes and phases of the life-cycle are most critical for skate population growth and persistence.

Methods

Model Structure and Simulation

A density-independent, probabilistic, age-structured matrix modeling approach was applied [26,27]. Vital rates were estimated from assessments of age, growth, and maturity as described below [16,20,22,23]. Field sampling approval was obtained from the Institutional Animal Care and Use Committee (IACUC #501) at San Jose State University. For each species, a population projection matrix (A) of age-specific survival probabilities (P_x , the probability of surviving from age x to x + 1) and fertilities (F_x) was created. The finite annual rate of population growth ($\lambda = e^{t}$, where r is the predicted instantaneous rate of population growth per capita; sensu [28]) was calculated as the dominant eigenvalue of each projection matrix. Reproductive value (v) and stable-age distribution (w) were calculated as the left and right eigenvectors associated with λ . Two measures of generation time were calculated: the mean age of parents as calculated from the total lifetime offspring production of a parental cohort (μ_I) [26] and the mean age of parents as calculated from the total offspring production of a population at the stable-age distribution (\bar{A}). \bar{A} was calculated from matrix elements following Cortés [27]. Other population parameters calculated were net reproductive rate (R_0) [26], and rate of increase per generation (rT) [29]. All models were female-only with an annual time step and assumed a closed population with birth-flow reproduction.

Elasticities, or proportional sensitivities [30,31], of λ were calculated to evaluate the proportional contribution of different aspects of the life cycle to λ following Caswell [26] as

$$e_{ij} = \frac{a_{ij}}{\lambda} \frac{\partial \lambda}{\partial a_{ij}},\tag{1}$$

where a_{ij} is the element in row i and column j of the matrix A. Fertility, juvenile survival, and adult survival elasticities were calculated as the sum of the corresponding matrix-element elasticities. Monte Carlo simulation was used to account for uncertainty in life-history parameter estimates that may have arisen due to measurement and process error [27]. Separate independent and identically distributed (iid) probability density functions (PDFs) were created for each species to represent uncertainty in longevity (ω) , age at maturity (α) , age-specific survival (S_x) , and age-specific fecundity (m_x) [27,32,33]. Simulations (n=5,000) were performed with values of each life history parameter chosen at random from the predetermined PDFs to create replicate matrices. Separate draws for S_x and S_x were taken for each individual age class within a given simulation step.

To determine the effect of assuming independence in vital rates among ages, an additional set of simulations was performed with only a single draw of $S_{\mathbf{x}}$ and $m_{\mathbf{x}}$ per simulation step for all age classes that shared the same PDF parameters. In other words, a scenario with completely uncorrelated (or iid) vital rates among ages was compared to a scenario with very high correlation among ages. Demographic output parameters were calculated from each replicate matrix, and summarized as means and 95% confidence intervals (defined as the range of values between the $2.5^{\rm th}$ and $97.5^{\rm th}$ percentiles).

Vital Rate Parameter Estimation

Mortality was expressed as age-specific annual survival rates $(S_x = e^{-M_x})$, where M_x is the total instantaneous rate of mortality of individuals of age x). Species-specific catch-at-age data were unavailable; therefore, M_r was estimated with six commonly-used indirect methods based on life-history parameters (Table 1, Table S1) [34,35,36,37]. Five of the methods used produce only a single mortality rate for the population, implying no difference among age classes [38,39]. The Chen and Watanabe [40] method estimates age-specific mortality rates based on von Bertalanffy growth function parameters. Because no prior information indicated differing accuracy of specific survival estimates, a uniform PDF was selected for S_x , with the range defined as the range of values obtained from all indirect estimation techniques (Table 1). Survival of the first two juvenile age classes was reduced by factors of 0.5 and 0.75, respectively, to represent additional mortality that these early-life ages are likely to incur [41,42].

Embryo survival was represented by empirical estimates of eggcase predation rates in the EBS [43,44]. Direct estimates were available for only two of the five species, *B. taranetzi* and *B. trachura*. Egg cases with similar morphologies likely experience similar levels of predation; therefore, estimates of *B. aleutica* embryo survival were used as a proxy for *B. lindbergi* and *B. maculata* [43]. Proportions of predated egg-cases across sites for five species in the EBS (*B. aleutica*, *B. interrupta*, *B. minispinosa*, *B. parmifera*, *B. trachura*) were used to bound the range of estimates for all species. Embryo survival was represented by triangular PDFs, with the most likely value for each species specified as the corresponding direct estimate or proxy (Table 2).

A skewed triangular PDF was selected for longevity (ω) , with the most probable value estimated to be the maximum observed age (T_{max}) . The lower range was defined as

$$\omega_L = T_{\text{max}}(1 - T_{\text{max}}E), \tag{2}$$

Table 1. Methods and indirect estimates of overall natural mortality (M) and annual survival (S_x) , where $S_x = \ln M$ used to parameterize demographic models in this study.

| Method | Bathyraja lindbergi | | Bathyraja maculata | | Bathyraja minispinosa | | Bathyraja taranetzi | | Bathyraja trachura | |
|---|---------------------|----------------|--------------------|----------------|-----------------------|----------------|---------------------|----------------|--------------------|----------------------|
| | М | S _x | М | S _x | М | S _x | М | S _x | М | <i>5_x</i> |
| 1. $\ln M = 1.46 - 1.010 \ln \omega$ [38] | 0.13 | 0.88 | 0.13 | 0.88 | 0.12 | 0.89 | 0.30 | 0.74 | 0.12 | 0.89 |
| 2. $\ln M = 1.44 - 0.982 \ln \omega$ [38] | 0.14 | 0.87 | 0.14 | 0.87 | 0.13 | 0.88 | 0.32 | 0.73 | 0.13 | 0.88 |
| 3. $M = 1.65/\alpha[39]$ | 0.09 | 0.91 | 0.07 | 0.93 | 0.07 | 0.93 | 0.18 | 0.84 | 0.06 | 0.94 |
| 4. $M = 1.50k[39]$ | 0.06 | 0.94 | 0.05 | 0.95 | 0.03 | 0.97 | 0.17 | 0.85 | 0.06 | 0.94 |
| 5. $M = 1.60k[39]$ | 0.07 | 0.94 | 0.05 | 0.95 | 0.03 | 0.97 | 0.18 | 0.84 | 0.06 | 0.94 |
| 6. $M_x = f(\alpha, k, t_o)$ [40] | 0.05-0.31 | 0.74-0.95 | 0.04-0.24 | 0.78-0.96 | 0.03-0.18 | 0.83-0.97 | 0.08-0.51 | 0.60-0.92 | 0.04-0.35 | 0.71-0.9 |

The parameters ω and α correspond to empirical estimates of longevity and the median age at maturity, respectively; k is the growth coefficient and t_0 the theoretical age when body size is zero estimated from the three-parameter von Bertalanffy growth function for each species [16,20,22,23]. See Chen and Watanabe [40] for the detailed form of their age-specific mortality functions. doi:10.1371/journal.pone.0065000.t001

where E is the index of average percent error (IAPE) [45] calculated using empirical data from the greatest 20% of age classes sampled by Ebert et al. [16], Maurer [23], Ainsley et al. [20], and Winton [22]. The upper range was defined as

$$\omega_U = T_{\text{max}}(c+E), \tag{3}$$

Where c is an arbitrary constant (here c = 1.4) to account for the likelihood that the absolute maximum age of each species was not

observed in the life history studies. Equation (3) was used because theoretical longevities calculated from von Bertalanffy growth function parameters are likely overestimates [16,20]. The method we have chosen for establishing a maximum bound on longevity also facilitates the comparability of demographic analysis results among species [27].

Age at maturity (α) was simplified to a knife-edge function, with all age classes younger than the age at 50% maturity considered immature and assigned a fecundity of zero. The fecundity value for the first mature age class was halved, to reflect that only 50% of

Table 2. Summary of parameter estimates and probability distribution functions (PDF) used to construct probabilistic agestructured matrix models for five deepwater Bering Sea skate species.

| Species | Parameter | PDF | Minimum | Maximum | Likeliest | 95% CI |
|-----------------------|----------------|------------|---------|---------|-----------|-----------|
| Bathyraja lindbergi | ω | triangular | 31 | 46 | 32 | |
| | α | logistic | 18 | 21 | 17.7 | 16.2-19.2 |
| | mx | uniform | 16 | 52 | | |
| | l_e | triangular | 0.65 | 0.99 | 0.98 | |
| Bathyraja maculata | ω | triangular | 31 | 46 | 32 | |
| | α | logistic | 18(21) | 30(24) | 22.5 | 21.7-23.3 |
| | mx | uniform | 16 | 52 | | |
| | l _e | triangular | 0.65 | 0.99 | 0.98 | |
| Bathyraja minispinosa | ω | triangular | 36 | 53 | 37 | |
| | α | logistic | 23 | 25 | 23.5 | 22.3-24.7 |
| | mx | uniform | 16 | 52 | | |
| | I_e | triangular | 0.65 | 0.98 | 0.87 | |
| Bathyraja taranetzi | Ω | triangular | 13 | 20 | 14 | |
| | A | logistic | 8 | 13(10) | 9.2 | 8.7-9.7 |
| | mx | uniform | 16 | 52 | | |
| | l_e | triangular | 0.65 | 0.98 | 0.81 | |
| Bathyraja trachura | Ω | triangular | 35 | 51 | 36 | |
| | A | logistic | 22(25) | 33(29) | 27.5 | 25.8-28.1 |
| | mx | uniform | 16 | 52 | | |
| | l _e | triangular | 0.64 | 0.98 | 0.65 | |

Values in parentheses indicate minimum or maximum values used when the range generated using 95% confidence intervals was more restricted than the range bound by first and 100% maturity. ω = longevity (yr); α = median age at maturity (yr); m_x = fecundity (yr⁻¹); l_e = egg case survival (yr⁻¹). doi:10.1371/journal.pone.0065000.t002

individuals would be reproducing during that year. For each species, a logistic PDF with a mean value equal to the age at 50% maturity was chosen to represent uncertainty in α , with lower and upper limits defined by age at first and 100% maturity [16,20,22,23]. To examine the potential impact of fisheries activity on population growth rates, values of size at maturity were plotted with species-specific length-frequency data from observed commercial fishery activity in the Bering Sea and Aleutian Islands management area from 2004 to 2012 (data provided by NMFS Alaska Fisheries Science Center).

Birth rates were expressed as age-specific fecundity values (m_x, m_y) where x = age in years). Skates are oviparous, so fecundity was defined as the number of eggs extruded female⁻¹ yr⁻¹. Fecundity estimates are notoriously difficult to obtain for oviparous elasmobranch species and were not available for the five species in this study; therefore, estimates from other skate species were used as a proxy. Only those fecundity estimates that were calculated with a robust method (captive parturition rate or ovarian fecundity estimates accounting for indeterminate spawning mode) were included. Mean fecundities for each skate species ranged from 32 to 104 (Table 3). Assuming a sex ratio at birth of 1:1 [46], these values were halved to represent the number of female offspring female⁻¹ year⁻¹. Because no prior information indicated that the fecundity of a given skate species is a better representative of EBS skate fecundity than any other, a uniform PDF (ranging from 16 to 52) was used to represent uncertainty in fecundity for all five species.

Many Alaskan skates have protracted and asynchronous eggcase deposition seasons [47], thus F_x and P_x were calculated for a birth-flow population, wherein individuals are constantly entering and leaving a given age class. To express this property in the matrix calculations, survival was approximated over the interval following Caswell [26] as

$$P_i = \frac{l(i) + l(i+1)}{l(i-1) + l(i)},\tag{4}$$

where survival from i-1 to i is

$$l(i) = l(i-1)S_i. (5)$$

Fertility coefficients were calculated following Caswell [26] as the product of averaged m_x values and the survival probability for half

Table 3. Skate fecundity data used to parameterize demographic models in this study.

| Species | Mean fecundity | References |
|--------------------|----------------|------------|
| Amblyraja radiata | 41 | [80] |
| Dipturus laevis | 90 | [80] |
| Leucoraja erinacea | 33 | [81] |
| Leucoraja naevus | 90 | [88] |
| Leucoraja ocellata | 48 | [80] |
| Okamejei kenojei | 86 | [84] |
| Raja brachyura | 65 | [2,83,87] |
| Raja clavata | 103 | [82,83,85] |
| Raja eglanteria | 60 | [86] |
| Raja montagui | 43 | [81,83,87] |

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of a time period $[l(0.5e)=l(e)\sqrt{l(0)}]$, where e is the egg-case age class]. In mathematical terms, F_i was calculated as

$$F_i = \frac{l(0.5e)(m_i + m_{i+1}P_i)}{2}.$$
 (6)

Correlates of Population Growth Rates

Tests of correlation between population growth rates and specific life-history traits were performed to identify indicators of productivity. Estimates of population growth rates for the five species investigated here were compared with other members of the Alaskan skate assemblage as well as other high-latitude skates. Estimates for other species were obtained from the literature (Table 4). Tests were based on a *t*-statistic calculated from Pearson's product moment correlation coefficient.

Results

Indirect estimates of M ranged from 0.03 to 0.51 yr $^{-1}$ (Table 1). Jensen's [39] methods generally produced the lowest estimates. The age-specific Chen and Watanabe [40] method generated higher M estimates for younger age classes than produced by age-independent methods. Corresponding S_x values ranged between 0.77–0.95 yr $^{-1}$ for B. lindbergi, 0.78–0.96 yr $^{-1}$ for B. maculata, 0.83–0.97 yr $^{-1}$ for B. minispinosa, 0.60–0.91 yr $^{-1}$ for B. taranetzi, and 0.71–0.96 yr $^{-1}$ for B. trachura (Table 1).

The projected finite rates of population increase (λ) were similar among the five species (Table 5). Mean λ ranged from 1.05 yr⁻¹ (*B. trachura*) to 1.12 yr⁻¹ (*B. taranetzi*), with lower confidence limits greater than 1.02 yr⁻¹ for all species. Corresponding estimates of \bar{A} were longest for *B. trachura* and shortest for *B. taranetzi*; mean μ_I values exhibited the same pattern. Means of rT and R_o were least for *B. taranetzi* and *B. trachura* but similar among other species.

Population growth rates were more sensitive to variation in age at maturity than any other parameter, with only minor changes in response to variation in longevity (Figure 1). Elasticity patterns were similar among all five species. Juvenile survival elasticities were the greatest for all species and ranged from a mean of 0.69 for B. taranetzi to 0.81 for B. trachura (Figure 2). Mean fertility elasticities were the least, ranging from 0.03 for B. minispinosa to 0.09 for B. taranetzi. Expressing these elasticities as ratios indicates that in order to compensate for as little as a 10% decrease in juvenile survival, an 81% (B. taranetzi) to 226% (B. minispinosa) increase in fertility or egg case survival would be required to return populations to their initial growth rates, depending on the species in question. Similarly, a 10% decrease in adult survival would require increases in fertility or egg case survival ranging from 21% (B. trachura) to 55% (B. minispinosa). Elasticity ratios of adult to juvenile survival were smaller; a 10% decrease in adult survival could be compensated for by increases in juvenile survival ranging from approximately 2% to 3% for all species.

Examination of length-frequency data from commercial fisheries in the EBS indicated that all five species are captured as both juveniles and adults (Figure 3). The smallest *B. lindbergi, B. trachura* and *B. minispinosa* observed in the catch were only slightly larger than the estimated size at hatching. In general, the dominant mode present in the catch was near the size at maturity. One exception to this pattern was found in the catch of *B. maculata*, which exhibited two modes at sizes slightly smaller and larger than the size at maturity, which were separated by a local minimum at the size at maturity itself.

Table 4. Estimated finite annual population growth rates (λ) and possible life-history correlates to λ for high-latitude skates.

| Species | 2 | a.w | TL_{max} (cm) | Depth (m) | References |
|-----------------------|-------|------|-----------------|-----------|---------------------------------|
| Rhinoraja interrupta | 1.360 | 0.58 | 83 | 695 | [7,14,46,90]; Ebert unpubl data |
| Beringraja binoculata | 1.334 | 0.19 | 244 | 402 | [13,14,92,93] |
| Bathyraja aleutica | 1.252 | 0.53 | 154 | 809 | [14,46,93] |
| Leucoraja erinacea | 1.234 | 0.50 | 57 | 192 | [48,89,94] |
| Dipturus laevis | 1.221 | 0.24 | 134 | 800 | [48,94,95,96,97] |
| Raja rhina | 1.202 | 0.35 | 204 | 320 | [7,13,14,93]; Ebert unpubl data |
| Amblyraja radiata | 1.197 | 0.58 | 101 | 859 | [5,95,98,99,100] |
| Leucoraja ocellata | 1.139 | 0.43 | 111 | 186 | [48,89,94] |
| Bathyraja taranetzi | 1.116 | 0.66 | 77 | 500 | [7]; this study |
| Bathyraja lindbergi | 1.110 | 0.55 | 97 | 540 | [7,93]; this study |
| Bathyraja minispinosa | 1.096 | 0.64 | 90 | 635 | [93]; this study |
| Bathyraja maculata | 1.079 | 0.70 | 120 | 564 | [7,93]; this study |
| Bathyraja trachura | 1.048 | 0.75 | 94 | 1169 | [101,102]; this study |
| Raja clavata | 0.930 | 0.71 | 104 | 289 | [2,91,103,104,105] |

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Effects of Correlated Vital Rate Elements

Simulations with correlation in vital rates among ages in a given year produced similar mean outcomes as the base case (Table 6), but substantially increased the variability of these estimates (95% confidence limits for λ were as low as 1.01 yr⁻¹ for *B. trachura* and as high as 1.22 yr⁻¹ for *B. taranetzi*; Table 6). Compared to the fully uncorrelated scenario, mean λ and rT were marginally decreased, but variability was substantially increased. Both mean

Table 5. Predicted means and 95% confidence intervals (range bounded by 2.5th and 97.5th percentiles) of finite annual rates of population growth (λ) , net reproductive rates (R_0) , rates of increase per generation (rT), mean ages of females generating offspring produced by a cohort during its lifetime (μ_1) , and mean ages of females generating offspring produced by a population at the stable-age distribution (\bar{A}) .

| Species | | λ | RO | rT | μ1 | Ā |
|-----------------------|-------|-------|-------|------|-------|-------|
| | 2.5% | 1.090 | 8.51 | 2.04 | 23.71 | 21.54 |
| Bathyraja lindbergi | mean | 1.112 | 13.18 | 2.47 | 25.30 | 22.61 |
| | 97.5% | 1.132 | 18.95 | 2.91 | 27.35 | 24.35 |
| | 2.5% | 1.063 | 5.45 | 1.57 | 26.64 | 24.71 |
| Bathyraja maculata | mean | 1.081 | 9.12 | 2.05 | 28.53 | 26.33 |
| | 97.5% | 1.100 | 13.99 | 2.54 | 30.79 | 28.08 |
| | 2.5% | 1.080 | 10.46 | 2.25 | 29.27 | 27.32 |
| Bathyraja minispinosa | mean | 1.097 | 16.82 | 2.73 | 31.33 | 28.80 |
| | 97.5% | 1.114 | 25.10 | 3.22 | 33.93 | 30.60 |
| | 2.5% | 1.063 | 2.13 | 0.60 | 11.30 | 9.31 |
| Bathyraja taranetzi | mean | 1.119 | 3.98 | 1.13 | 12.25 | 10.36 |
| | 97.5% | 1.170 | 6.33 | 1.60 | 13.47 | 11.52 |
| | 2.5% | 1.027 | 2.38 | 0.87 | 30.08 | 29.44 |
| Bathyraja trachura | mean | 1.048 | 4.86 | 1.53 | 32.54 | 31.81 |
| | 97.5% | 1.068 | 8.26 | 2.11 | 35.47 | 34.43 |

Values were estimated from 5,000 Monte Carlo simulations. doi:10.1371/journal.pone.0065000.t005

and variance of R_{θ} were substantially increased. Neither means nor variances of \bar{A} , μ_I or any of the elasticities (Figure S1) were appreciably changed.

Correlates of Population Growth Rates

Among high-latitude skates for which data were available (Table 4), population growth rates were negatively correlated with the ratio of age at maturity to longevity (Figure 4; Pearson's product-moment correlation, $r_p = -0.65$, n = 14, P = 0.01). In contrast, there was no significant correlation between population growth rate and maximum total length ($r_p = 0.37$, n = 14, P = 0.19) or the midpoint of the depth range ($r_p = 0.03$, n = 14, P = 0.93). Similarly, within the EBS skate assemblage population growth rates were negatively correlated with the ratio of age at maturity to longevity (Figure 5; $r_p = -0.69$, n = 9, P = 0.04) and were not significantly correlated with maximum total length ($r_p = 0.50$, n = 9, P=0.17) or the midpoint of the depth range $(r_p=-0.29, n=9,$ P = 0.44). Note that the ratio of age at maturity to longevity increases as the number of reproductive years becomes lower relative to the total lifespan; thus this metric is an inverse proxy for reproductive lifespan (i.e., higher values generally indicate a shorter window of lifetime reproduction).

Among high-latitude skates for which data were available (Table 4), population growth rates were negatively correlated with the proportion of the lifespan devoted to maturation, or the ratio of age at maturity to longevity (Figure 4; Pearson's productmoment correlation, $r_p = -0.65$, n = 14, P = 0.01). In contrast, there was no significant correlation between population growth rate and maximum total length ($r_p = 0.37$, n = 14, P = 0.19) or the midpoint of the depth range ($r_p = 0.03$, n = 14, P = 0.93). Similarly, within the EBS skate assemblage population growth rate was negatively correlated with the proportion of lifespan devoted to maturation (Figure 5; $r_p = -0.69$, n = 9, P = 0.04) and was not significantly correlated with maximum total length ($r_p = 0.50$, n = 9, P = 0.17) or the midpoint of the depth range $(r_p = -0.29, n = 9, P = 0.44)$. Note that the proportion of the lifespan devoted to maturation is equivalent to 1 – [relative reproductive lifespan], where the relative reproductive lifespan is reproductive lifespan scaled to longevity. Thus, the above results indicate that population growth rates increased with the relative reproductive lifespan.

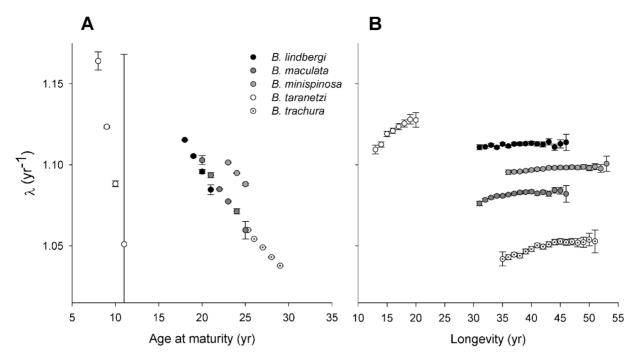


Figure 1. Influence of age at maturity and longevity on finite population growth rates (λ). Mean estimates were produced using probabilistic, age-structured matrix models based on empirical estimates of longevity and maturity for five Bering Sea skate species: Bathyraja lindbergi, B. maculata, B. minispinosa, B. taranetzi, and B. trachura. Variability in the estimated age at maturity (A) affected λ with greater per-unit magnitude than variation in longevity (B). Error bars indicate 95% confidence intervals. doi:10.1371/journal.pone.0065000.g001

Discussion

Elasmobranch life histories have been demonstrated to fall along a "fast-slow" continuum, with "slow" species defined as those with later age at maturity, longer generation times, and lower population growth rates than "fast" species [4,27]. The relative position of skates along this life-history continuum remains unclear because few demographic analyses have been performed on this group. Of the skate species analyzed herein, *B. taranetzi* was the most productive; this species had the highest predicted

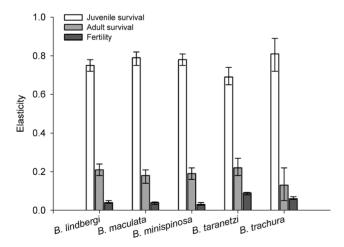


Figure 2. Elasticity analysis. Predicted means and 95% confidence intervals (range bounded by 2.5th and 97.5th percentiles) of elasticities for five Bering Sea skate species. Mean values were estimated from 5,000 Monte Carlo simulations assuming independent and identically distributed vital rates among age classes. doi:10.1371/journal.pone.0065000.q002

population growth rate and shorter generation time than the other species. *Bathyraja lindbergi* and *B. maculata* were the next most productive, with the former perhaps slightly more robust to perturbations in mortality because of a shorter generation time than that of *B. maculata. Bathyraja trachura* is perhaps the least productive of the species analyzed here, primarily because it has the longest generation time.

Among studies that have incorporated uncertainty in life-history parameters for other elasmobranchs, λ values estimated for the five species herein were substantially lower than the midpoint of the range (λ =1.25 yr⁻¹), but similar to the median (λ =1.08 yr⁻¹) [14,27,33,48,49,50,51,52,53,54]. These results are contrary to the results of prior cross-taxonomic comparisons, which placed the Rajiformes as the chondrichthyan order least vulnerable to extinction [24] and the most likely to recover from overexploitation [25]. Using the Musick [55] criteria based on productivity—as measured by population growth rate—all five skates examined in this study are classified as vulnerable to population depletion by fishing mortality.

In combination with estimated population growth rates, evaluation of other demographic parameter estimates also indicates that the species analyzed here have moderately "slow" life histories compared to other elasmobranchs and are substantially "slower" than other Alaskan skates. Mean generation times for all species except B. taranetzi were higher than other Alaskan skates [14]. Compared to other elasmobranchs, values of generation time for B. taranetzi were close to the median $(\bar{A}=11.20 \text{ yr})$, but less than the midpoint $(\bar{A}=29.25 \text{ yr})$ of the range [14,27,33,49,50,52]; for the other four species, values of \bar{A} were in the upper end of the range. Mean rates of increase per generation (rT) were lower than other Alaskan skates [14] but higher than those of many other elasmobranchs [33,56], with the exception of B. taranetzi. Values of R_0 were lower than those of

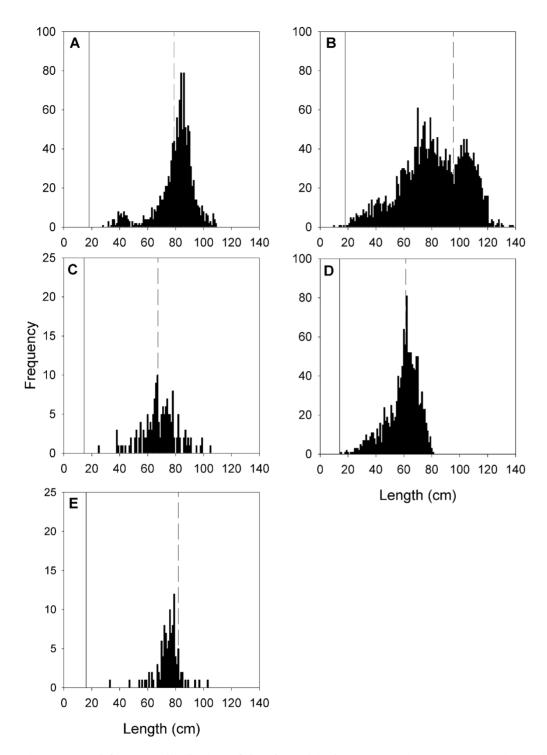


Figure 3. Length frequency distributions of skate bycatch in the eastern Bering Sea. Data were obtained from observers (National Marine Fisheries Service, Alaska Fisheries Science Center) during commercial fishery activity in the Bering Sea and Aleutian Islands management area from 2004 to 2012. Estimated size at birth (solid line) and size at maturity (dashed line) are indicated for each species: *Bathyraja lindbergi* (A), *B. maculata* (B), *B. minispinosa* (C), *B. taranetzi* (D), and *B. trachura* (E). doi:10.1371/journal.pone.0065000.g003

most other Alaskan skates [14], with some similarity between B. aleutica (mean $R_0 = 23.26$; [14]) and B. minispinosa (mean $R_0 = 16.82$; Table 5). In summary, skate life histories likely exhibit a similar continuum to that of sharks, with substantial interspecific variation present even within the North Pacific basin. These results further emphasize the idea that this type of fast-slow continuum

occurs within many lower taxonomic groups [57], and that it is often not appropriate to generalize a specific life-history strategy to an entire family.

Variability in α often affects λ with greater per-unit magnitude than other vital rates in elasmobranchs [4,27]; there is often enough uncertainty in this parameter to cause great uncertainty in

Table 6. Predicted means and 95% confidence intervals (range bounded by 2.5th and 97.5th percentiles) of finite annual rates of population growth (λ) , net reproductive rates (R_0) , rates of increase per generation (rT), mean ages of females generating offspring produced by a cohort during its lifetime (μ_1) , and mean ages of females generating offspring produced by a population at the stable-age distribution (\bar{A}) .

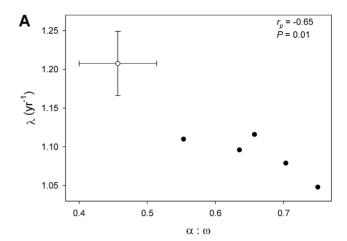
| Species | | λ | RO | rT | μ1 | Ā |
|-----------------------|-------|-------|-------|------|-------|-------|
| | 2.5% | 1.056 | 3.70 | 1.22 | 23.68 | 21.30 |
| Bathyraja lindbergi | mean | 1.110 | 15.01 | 2.44 | 25.30 | 22.62 |
| | 97.5% | 1.164 | 38.73 | 3.67 | 27.66 | 24.41 |
| | 2.5% | 1.022 | 1.82 | 0.55 | 26.65 | 24.25 |
| Bathyraja maculata | mean | 1.079 | 11.51 | 2.01 | 28.52 | 26.32 |
| | 97.5% | 1.137 | 36.76 | 3.53 | 31.03 | 28.53 |
| | 2.5% | 1.039 | 3.15 | 1.08 | 29.13 | 26.82 |
| Bathyraja minispinosa | mean | 1.096 | 22.47 | 2.70 | 31.37 | 28.81 |
| | 97.5% | 1.151 | 75.00 | 4.42 | 34.69 | 31.03 |
| | 2.5% | 1.013 | 1.17 | 0.12 | 11.36 | 9.24 |
| Bathyraja taranetzi | mean | 1.116 | 4.27 | 1.10 | 12.24 | 10.34 |
| | 97.5% | 1.218 | 10.41 | 2.03 | 13.43 | 11.57 |
| | 2.5% | 1.008 | 1.31 | 0.27 | 30.07 | 29.42 |
| Bathyraja trachura | mean | 1.046 | 5.11 | 1.46 | 32.52 | 31.81 |
| | 97.5% | 1.083 | 13.10 | 2.57 | 35.46 | 34.49 |

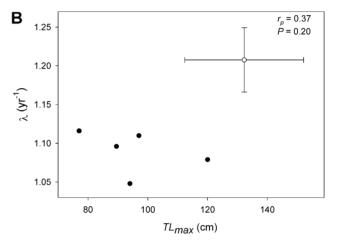
Values were estimated from 5,000 Monte Carlo simulations with perfect correlation in vital rates among age classes with the same probability distributions.

doi:10.1371/journal.pone.0065000.t006

estimates of λ , e.g. [58]. Population growth rates for the five species investigated herein were highly sensitive to this parameter estimate, as exemplified by the difference in the absolute change in λ across the range of parameter uncertainty in α relative to ω (Figure 1). Age at maturity was particularly influential in the case of the shorter-lived *B. taranetzi*. Similarly, the effect of differing ω on λ was greater in *B. taranetzi* than the other four species. These interspecific differences were likely caused by the presence of a higher relative proportion of individuals remaining in the reproductive and ultimate age classes in *B. taranetzi*.

Elasticity analyses indicated that population growth rates were more influenced by juvenile and adult survival than the combination of egg case survival and fecundity (encompassed within the fertility elasticity; Figure 2), supporting this as a general pattern for many long-lived vertebrates [4,27,59,60]. Therefore, all five species follow the expected pattern proposed by Cortés [27], who found that juvenile survival elasticities were the highest for all shark species with $\alpha \ge 5$ yr. Predicted rank order of elasticity values was similar to long-lived elasmobranchs [27] and sea turtles [59]. Juvenile survival elasticities estimated in this study represent some of the highest values estimated for any vertebrate, whereas fertility elasticities were on the extreme low end of this spectrum [59,61]. Concordantly, adult survival elasticities were quite low relative to other vertebrates. Juvenile survival and fertility elasticities were not as extreme for B. taranetzi as for the other four species. Adult survival elasticities were most extreme for B. trachura; values for the other four species were moderately low compared to all elasmobranchs [median = 0.31; range = (0.12-[0.51] [14,27,33,49,50,51,52,53].





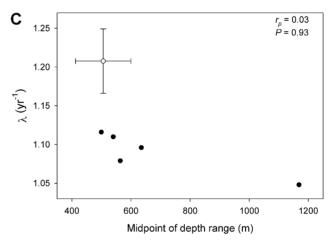
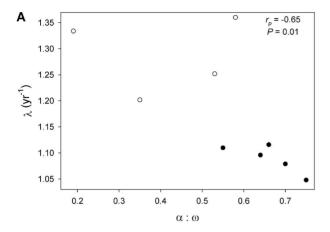
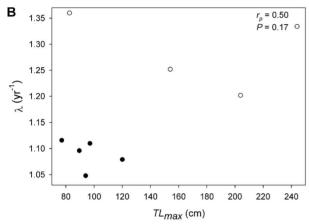


Figure 4. Correlates of population growth rates in high latitude skates. Mean finite rates of population growth (λ) predicted from probabilistic, age-structured matrix models in relation to the proportion of lifespan devoted to maturation [the ratio of age at maturity (α) to longevity (ω)] (A), maximum total length $(TL_{max}; B)$, and midpoint of depth range (C). Note that higher values of $\alpha:\omega$, reported as a proportion, reflect a shorter reproductive lifespan relative to the total lifespan. Data are shown for each of the five species analyzed in this study, in comparison to the mean and variance from nine other high-latitude skates (Table 5). Pearson's product-moment correlation coefficients and related P-values are shown for each relationship. Error bars indicate 95% confidence intervals (among-species). doi:10.1371/journal.pone.0065000.g004





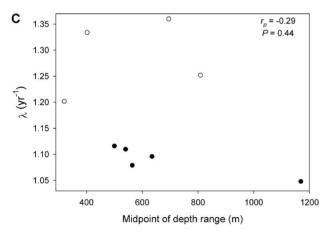


Figure 5. Correlates of population growth rates in Alaskan skates. Mean finite rates of population growth (λ) predicted from probabilistic, age-structured matrix models in relation to the proportion of lifespan devoted to maturation [the ratio of age at maturity (α) to longevity (α)] (A), maximum total length (TL_{max} , B), and midpoint of depth range (C). Note that higher values of α : α , reported as a proportion, reflect a shorter reproductive lifespan relative to the total lifespan. Data are shown for each of the five species analyzed in this study (filled circles), in comparison to the four other species of Alaskan skates for which data were available (open circles; Table 5). Pearson's product-moment correlation coefficients and related p-values are shown for each relationship. doi:10.1371/journal.pone.0065000.g005

Given the limited potential of elasmobranchs to increase adult survival or fecundity, exploitation compensation would most likely be exhibited through increases in egg case or juvenile survival, or through decreases in age at maturity over time [4,27]. The results reported herein indicate that density-dependent compensation in the form of increased egg case survival would not substantially increase λ , particularly for *B. lindbergi*, *B. maculata*, and *B. minispinosa*. Additionally, there is evidence that embryonic survival during the egg case stage is positively correlated with density in skates [44], making increased egg case survival an even less likely compensation mechanism. In contrast, ratios of adult to juvenile survival indicate that compensation is more likely to occur through increased juvenile survival due to reduced intraspecific competition at lower population densities.

Data Quality and the Status of the Best Available Science in Elasmobranch Demography

It is important to note that the demographic parameter estimates reported herein are a first approximation and will be improved as life-history data continue to be collected. The reliability of the results of demographic models is determined by the quality of the vital rate estimates incorporated; the more robust and precise the parameter estimates and the assumptions they rely upon, the more reliable any potential management strategy based upon the results. For example, because sample sizes used for age estimation of each species were likely not large enough to include the absolute oldest individuals in the population, maximum observed ages may underestimate the species' true longevity. However, maximum ages are likely not far off those empirically estimated; therefore, the application of a correction factor was used to produce longevity estimates within reasonable bounds. Additionally, age estimates were based on the assumption of annual band pair formation and have not been directly validated in these species [16,20,22,23]. Considering the importance of accurate age estimates for management, direct validation of age estimates is imperative. Although the models presented were developed using the best available information, the uncertainty in longevity and the lack of species-specific fecundity data dictates that these results should not be used for applied purposes without appropriate precaution (see Text S1 for additional caveats and empirical needs).

Although there have been recent calls to incorporate densitydependent effects into demographic models [28,62], the data necessary to parameterize these effects is sorely lacking. In addition, some of the techniques proposed assume that there is no fishing mortality on juveniles [63], which is violated in the case of EBS skates [8], as indicated by the length-frequency distribution of skates taken as bycatch in commercial fisheries in the EBS (Fig. 3). Compensation is expected to occur primarily through increased juvenile survival at low population sizes [64], yet creating a realistic representation of this effect is rarely possible given the current quality of chondrichthyan data. Models that have attempted to incorporate density-dependent effects have demonstrated a limited magnitude of potential compensatory effects [65]. Proper representation would require collection of repeated estimates of juvenile survival over time or space, coupled with concurrent density estimates.

Life-history Correlates of Population Growth Rate

Results of both comparisons among all high-latitude skates (Fig. 4) and comparisons within the EBS region (Figure 5) support neither the paradigm of a negative correlation of population growth rate with body size [66], nor the paradigm of correlation with depth [24,25,67]. These results indicate that the nature of

potential shifts in community structure of skates will not be easily predicted by body size, as has been argued for the North Atlantic skate assemblage [66]. For the eastern Scotian Shelf skate assemblage, McPhie and Campana [5] found that the species with the highest ratios of age at maturity to maximum age had the lowest predicted population growth rates, and that this was a better predictor of population growth rate than was body size. Including the species analyzed here, data for high-latitude skates supports the ratio between age at maturity and longevity as the best indicator of population growth rate (Figs. 4, 5). This does not necessarily invalidate the use of body size as an indicator of productivity or vulnerability when age data are not available, as these sets of metrics often covary. Body size may indeed explain different forms of variation in life histories than those encompassed by age metrics [68].

The greater importance of these age metrics relative to body size has also been identified at higher taxonomic scales among chondrichthyans [69], teleosts and mammals [70]. This result is intuitive; the earlier in life that a species begins reproducing and the more years it reproduces increases its expected lifetime reproductive output. This could be offset by interspecific differences in fecundity or patterns of natural mortality, but the analyses here and elsewhere suggest that this is not the case. Finally, it must be cautioned that although the direction of these relationships are likely robust, the specific form of these relationships should not be taken as gospel given the relatively low sample sizes in these correlation tests.

Across much broader taxonomic scales than those investigated here, theory and data have shown that there are invariant relationships between age at maturity and longevity (or reproductive lifespan) within related taxa [71]. Thus, this metric has been previously established as useful for characterizing life-history patterns [72,73]; it reflects the tradeoff between the mortality risk of delaying maturity and the reproductive benefits of increased fecundity with larger body size [74]. Our results add support to the importance of this quantity as a metric for synthesizing how suites of life history traits combine to determine population growth rate.

Management Implications

Since 2011, skates in the EBS have been managed as a separate aggregate complex rather than as a conglomeration with dissimilar taxa [8]. Though this is certainly an improvement over previous practices, skate assemblages in other regions have experienced alterations in abundance, size and age structure, and species composition as a result of both direct and indirect fishing effort over relatively short timescales [2,3,19]. To date, demographic analyses have been conducted on nine of the fifteen members of the Alaskan skate assemblage (this study; [14]). Resulting estimates of demographic parameters indicate that the five species investigated herein are among the least productive species in the assemblage and may experience more rapid declines than other species as the result of exploitation (Table 4). Among the five species analyzed here, it appears that commercial fisheries most often encounter B. maculata, with B. lindbergi and B. minispinosa a close second; B. taranetzi and B. trachura are caught in much lower numbers (Fig. 3). The combination of evidence indicates that B. maculata is the most likely to decline in abundance, given the relatively low predicted rate of population growth and high fisheries-encounter and retention rates [9]. Although considerable measures have been taken in recent years to record species-specific bycatch data in Alaska's groundfish fisheries, current predictions of potential assemblage-wide shifts would be speculative, given the sparseness of this dataset and limited knowledge of the differences in the size- and depth-selectivity of the region's fisheries.

The results of the present study provide valuable information to inform decisions regarding management of the skate assemblage in the EBS. Length-frequency data collected by observers in the EBS and Aleutian Islands indicates that both juveniles and adults of all five species are taken as bycatch in the regions' fisheries and that skates are susceptible to commercial gear at extremely young ages (Figure 3). Given the relative similarity of hatching size among the five species of interest, the presence of newly-hatched B. maculata and B. taranetzi in the catch indicates that the entire assemblage is likely susceptible to gear from the time of hatching. There is also evidence that egg-cases are occasionally caught in large quantities in Bering Sea groundfishery bottom trawls (North Pacific Fisheries Groundfish Observer Program unpubl data, as cited in [47]); therefore, these five species are likely exploited at all stages of their lifecycle. In addition to the potentially negative effects of early age at first recruitment on sustainable catch rates [65], elasticity analyses indicate that population growth rates of these five species are more affected on a per-unit basis by changes in juvenile survival rather than adult or egg case survival. Thus, efforts directed at regulating juvenile mortality would likely be the most effective approach if mitigation of population depletion becomes a management concern.

Although elasticities do not provide a sufficient basis for management on their own [27,59], they do facilitate prioritization of management efforts. Trawl survey data indicate that several areas along the EBS shelf serve as nursery sites for bathyrajid species [47]. Although elasticity analyses indicate that changes in embryonic and adult survival do not have as large of a per-unit influence on population growth rates as does juvenile survival, protection of nursery sites where egg cases and adults are aggregated [47] is nevertheless likely to be important for population persistence, particularly when considering the sheer magnitude of potential effects of trawl gear on such sites. As such, it would benefit fisheries managers to consider how management actions directed at one life stage affect vital rates at other life stages [27].

If managers determine that fishery impacts are imposing unacceptably negative effects on skate populations in the region, depth closures could be effective. Rajids are often distributed bathymetrically according to size, e.g. [75]; thus, permanent or temporary depth/area closures could be implemented to reduce juvenile mortality and stabilize age at maturity schedules. For skates, it is often assumed that all ages are potentially vulnerable to fishery mortality [8], with a slight increase in selectivity with size [76]. The bycatch data presented here indicate that selectivity for most EBS skates (with the exception of *B. maculata*) reaches a maximum near the size at maturity and perhaps declines at larger sizes (Figure 3). This pattern is troubling because it is indicative of directional selection on size and age at maturity. The results of the elasticity analysis presented here suggest this form of selection on maturation may lead to dramatic changes in population growth rates.

Within the EBS skate assemblage, body size is negatively correlated with depth ([47]; Ebert unpubl data). Therefore, restricting fishing at mid-depths and concentrating fishing at shallower and deeper depths could create a situation that approximates a *de facto* slot fishery, thereby reducing artificial selection for change in age at maturity. Currently, the skates with the deepest distributions off of Alaska have quite low bycatch rates [9], but if the fleets shift their effort and begin fishing more intensively in the deeper depth strata, these species would benefit greatly from depth restrictions, particularly in areas of high eggcase densities [47]. Based on the available length-frequency data of skate bycatch in the region, juvenile mortality from trawl gear

could be substantial if fishing occurs in or around nursery sites. With innovation it is possible that bycatch reduction devices could be used to accomplish the same objectives of size-selective harvest [77,78]. These fisheries management recommendations are more or less in line with previous suggestions for chondrichthyans [60,64], including those provided through density-dependent analyses [65,79].

Supporting Information

Figure S1 Elasticity analysis for simulations incorporating correlation in vital rates. Predicted means and 95% confidence intervals (range bounded by 2.5th and 97.5th percentiles) of elasticities for five Bering Sea skate species. Mean values were estimated from 5,000 Monte Carlo simulations assuming perfect correlation in vital rates among age classes with the same probability distributions. (DOCX)

Table S1 Additional natural mortality estimates. Agespecific estimates of natural mortality (M_x) and annual survival (S_x) , where $S_x = \ln M_x$) used to parameterize demographic models in this study. Calculations follow Chen and Watanabe [41]. (DOCX)

References

- Holden MJ (1973) Are long-term sustainable fisheries for elasmobranchs possible? Rapp P-v Réun Cons Int Explor Mer 164: 360–367.
- Walker PA, Hislop JRG (1998) Sensitive skates or resilient rays? Spatial and temporal shifts in ray species composition in the central and north-western North Sea between 1930 and the present day. ICES J Mar Sci 55: 392–402.
- Dulvy NK, Metcalfe JD, Glanville J, Pawson MG, Reynolds JD (2000) Fishery stability, local extinctions, and shifts in community structure in skates. Conserv biol 14: 283–293.
- Frisk MG, Miller TJ, Dulvy NK (2005) Life histories and vulnerability to exploitation of elasmobranchs: inferences from elasticity, perturbation and phylogenetic analyses. J Northwest Atl Fish Sci 35: 27–45.
- McPhie RP, Campana SE (2009) Reproductive characteristics and population decline of four species of skate (Rajidae) off the eastern coast of Canada. J Fish Biol 75: 223–246.
- Frisk M (2010) Life history strategies of batoids. In: Carrier JC, Musick JA, Heithaus MR, editors. Sharks and their relatives II. Boca Raton: CRC Press. 283–316
- Stevenson DE, Orr JW, Hoff GR, McEachran JD (2008) Emerging patterns of species richness, diversity, population density, and distribution in the skates (Rajidae) of Alaska. Fish Bull 106: 24–39.
- 8. Ormseth O, Matta B, Hoff J (2010) Bering Sea and Aleutian Islands skates. In: K. Aydin SB, D Clausen, M.E Conners, C Conrath, M Dalton, J DiCosimo, K. Echave SG, K Goldman, D Hanselman, J Heifetz, J Hoff, T Honkalehto, J Ianelli, S Kotwicki, R Lauth, S Lowe, C Lunsford, D McKelvey, B Matta, D Nichol, O.A Ormseth, C.J Rodgveller, C.N Rooper, P Spencer, C Spital, I Spies, W Stockhausen, T TenBrink, G Thompson, C Tribuzio, T Wilderbuer, M Wilkins, editors. Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions. Anchorage: North Pacific Fishery Management Council. 1365–1450.
- Stevenson DE, Lewis KA (2010) Observer-reported skate bycatch in the commercial groundfish fisheries of Alaska. Fish Bull 1082: 208–217.
- Laptikhovsky VV (2004) Survival rates for rays discarded by the bottom trawl squid fishery off the Falkland Islands. Fish Bull 102: 757–759.
- Enever R, Catchpole TL, Ellis JR, Grant A (2009) The survival of skates (Rajidae) caught by demersal trawlers fishing in UK waters. Fish Res 97: 72–76.
- 12. Ormseth O, Matta B (2007) Bering Sea and Aleutian Islands skates. In: K. Aydin SB, D Clausen, M.E Conners, D Courtney, J DiCosimo, M Dorn, J Fujioka, S Gaichas, K Goldman, D Hanselman, J Heifetz, G Hoff, T Honkalehto, J Ianelli, E Jorgenson, S Kotwicki, R Lauth, S Lowe, C Lunsford, B Matta, D Nichol, O.A Ormseth, R Reuter, C.J Rodgveller, P Spencer, I Spies, W Stockhausen, T TenBrink, G Thompson, C Tribuzio, T Wilderbuer, M Wilkins, G Williams, and N Williamson editors. Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands region. Anchorage: North Pacific Fishery Management Council. 909–930.
- McFarlane GA, King JR (2006) Age and growth of big skate (*Raja binoculata*) and longnose skate (*Raja rhina*) in British Columbia waters. Fish Res 78: 169– 178

Text S1 Additional caveats and empirical needs. Additional text pertaining to the discussion section *Data quality and the status of the best available science in elasmobranch demography*. (DOCX)

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Author Contributions

Conceived and designed the experiments: DAE GMC LAKB MVW. Performed the experiments: LAKB MVW. Analyzed the data: LAKB MVW. Wrote the paper: LAKB MVW SMA.

- Ebert DA, Smith WD, Haas DL, Ainsley SM, Cailliet GM (2007) Life history and population dynamics of Alaskan skates: providing essential biological information for effective management of bycatch and target species. North Pacific Research Board Final Report 510.
- Matta M, Gunderson D (2007) Age, growth, maturity, and mortality of the Alaska skate, *Bathyraja parmifera*, in the eastern Bering Sea. Environ Biol Fishes 80: 309–323.
- Ebert DA, Maurer JR, Ainsley SM, Barnett LAK, Cailliet GM (2009) Life history and population dynamics of four endemic Alaskan skates: determining essential biological information for effective management of bycatch and target species. North Pacific Research Board Final Report 715.
- Ebert DA, Smith WD, Cailliet GM (2008) Reproductive biology of two commercially exploited skates, *Raja binoculata* and *R. rhina*, in the western Gulf of Alaska. Fish Res 94: 48–57.
- Gburski C, Gaichas S, Kimura D (2007) Age and growth of big skate (Raja binoculata) and longnose skate (R. rhina) in the Gulf of Alaska. Environ Biol Fishes 80: 337–349.
- Agnew DJ, Nolan CP, Beddington JR, Baranowski R (2000) Approaches to the assessment and management of multispecies skate and ray fisheries using the Falkland Islands fishery as an example. Can J Fish Aquat Sci 57: 429–440.
- Ainsley SM, Ebert DA, Cailliet GM (2011) Age, growth, and maturity of the whitebrow skate, *Ballyraja minispinosa*, from the eastern Bering Sea. ICES J Mar Sci 68: 1426–1434.
- Ainsley SM, Ebert DA, Cailliet GM (2011) A comparison of reproductive parameters of the Bering skate, *Ballyraja interrupta*, from two Alaskan large marine ecosystems. Mar Freshw Res 62: 557–566.
- Winton MV (2011) Age, growth, and demography of the roughtail skate, Bathyraja trachura (Gilbert, 1892), from the eastern Bering Sea [Master thesis, Moss Landing Marine Labs, California State University Monterey Bay].
- Maurer JR (2009) Life history of two bering sea slope skates: Bathyraja lindbergi and B. maculata [Master thesis, Moss Landing Marine Labs, California State University Monterey Bay].
- García VB, Lucifora LO, Myers RA (2008) The importance of habitat and life history to extinction risk in sharks, skates, rays and chimaeras. Proc R Soc B 275: 83–89.
- Simpfendorfer CA, Kyne PM (2009) Limited potential to recover from overfishing raises concerns for deep-sea sharks, rays and chimaeras. Environ Conserv 36: 97–103.
- 26. Caswell H (2001) Matrix population models: construction, analysis, and interpretation. Sunderland, USA: Sinauer Associates.
- Cortés E (2002) Incorporating uncertainty into demographic modeling: application to shark populations and their convervation. Conserv biol 16: 1048–1069
- Gedamke T, Hoenig JM, Musick JA, DuPaul WD (2007) Using demographic models to determine intrinsic rate of increase and sustainable fishing for elasmobranchs: pitfalls, advances, and applications. N Am J Fish Manage 27: 665–618
- Fowler C (1988) Population dynamics as related to rate of increase per generation. Evol Ecol 2: 197–204.

- de Kroon H, Plaisier A, van Groenendael J, Caswell H (1986) Elasticity: the relative contribution of demographic parameters to population growth rate. Ecology 67: 1427–1431.
- Caswell H, Naiman RJ, Morin R (1984) Evaluating the consequences of reproduction in complex salmonid life cycles. Aquaculture 43: 123–134.
- Tuljapurkar S (1997) Stochastic matrix models. In: Tuljapurkar S, Caswell H, editors. Structured population models in marine, freshwater and terrestrial ecosystems. New York: Chapman and Hall. 59–88.
- Smith WD, Cailliet GM, Cortés E (2008) Demography and elasticity of the diamond stingray, *Dasyatis dipterura*: parameter uncertainty and resilience to fishing pressure. Mar Freshw Res 59: 575–586.
- 34. Cortés E (2000) Life history patterns and correlations in sharks. Rev Fish Sci 8: 299–344.
- Cortés E (1998) Demographic Analysis as an Aid in Shark Stock Assessment and Management. Fish Res 39: 199–208.
- Simpfendorfer CA (1999) Mortality Estimates and Demographic Analysis for the Australian Sharpnose Shark, *Rhizoprionodon taylori*, from Northern Australia. Fish Bull 97: 978–986.
- Heupel MR, Carlson JK, Simpfendorfer CA (2007) Shark nursery areas: concepts, definition, characterization and assumptions. Mar Ecol Prog Ser 337: 287–297.
- Hoenig JM (1983) Empirical use of longevity data to estimate mortality rates. Fish Bull 82: 898–903.
- Jensen AL (1996) Beverton and Holt life history invariants result from optimal trade-off of reproduction and survival. Can J Fish Aquat Sci 53: 820–822.
- Chen S, Watanabe S (1989) Age dependence of natural mortality coefficient in fish population dynamics. Bull Jpn Soc Sci Fish 55: 205–208.
- 41. Hoenig JM, Gruber SH (1990) Life history patterns in the elasmobranchs: implications for fisheries management. In: Pratt Jr HL, Gruber SH, Taniuchi T, editors. Elasmobranchs as living resources: advances in the biology, ecology, systematics, and the status of the fisheries: NOAA Technical Report 90. 1–16.
- Heupel MR, Simpfendorfer CA (2002) Estimation of mortality of juvenile blacktip sharks, *Carcharhinus limbatus*, within a nursery area using telemetry data. Can J Fish Aquat Sci 59: 624–632.
- Hoff GR (2007) Reproductive biology of the Alaska skate Bathyraja parmifera, with regard to nursery sites, embryo development and predation [Dissertation, University of Washington].
- Hoff GR (2009) Skate Bathyraja spp. egg predation in the eastern Bering Sea. J Fish Biol 74: 250–269.
- Beamish RJ, Fournier DA (1981) A method for comparing the precision of a set of age determinations. Can J Fish Aquat Sci 38: 982–983.
- Ebert DA (2005) Reproductive biology of skates, Bathyraja (Ishiyama), along the eastern Bering Sea continental slope. J Fish Biol 66: 618–649.
- Hoff G (2010) Identification of skate nursery habitat in the eastern Bering Sea. Mar Ecol Prog Ser 403: 243–254.
- Frisk MG, Miller TJ, Fogarty MJ (2002) The population dynamics of little skate Leucoraja erinacea, winter skate Leucoraja ocellata, and barndoor skate Dipturus laevis: predicting exploitation limits using matrix analyses. ICES J Mar Sci 59: 576–586.
- Goldman KJ (2002) Aspects of age, growth, demographics and thermal biology of two lamniform shark species [Dissertation, Virginia Institute of Marine Science]: Virginia Institute of Marine Science. 220 p.
- Romine JG, Musick JA, Burgess GH (2009) Demographic analyses of the dusky shark, Carcharhinus obscurus, in the Northwest Atlantic incorporating hooking mortality estimates and revised reproductive parameters. Environ Biol Fishes 84: 277–289.
- Beerkircher L, Shivji M, Cortés E (2003) A Monte Carlo demographic analysis
 of the silky shark (*Carcharhinus falciformis*): implications of gear selectivity. Fish
 Bull 101: 168–174.
- Carlson JK, Cortes E, Bethea DM (2003) Life history and population dynamics of the finetooth shark (*Carcharhinus isodon*) in the northeastern Gulf of Mexico. Fish Bull 101: 281–292.
- Aires-da-Silva AM, Gallucci VF (2007) Demographic and risk analyses applied to management and conservation of the blue shark (*Prionace glauca*) in the North Atlantic Ocean. Mar Freshw Res 58: 570–580.
- McAuley RB, Simpfendorfer CA, Hall NG (2007) A method for evaluating the impacts of fishing mortality and stochastic influences on the demography of two long-lived shark stocks. ICES J Mar Sci 64: 1710–1722.
- Musick JA (1999) Criteria to define extinction risk in marine fishes. Fisheries 24: 6–14.
- Mollet HF, Cailliet GM (2002) Comparative population demography of elasmobranchs using life history tables, Leslie matrices and stage-based matrix models. Mar Freshw Res 53: 503–516.
- Charnov EL, Berrigan D (1991) Evolution of life history parameters in animals with indeterminate growth, particularly fish. Evol Ecol 5: 63–68.
- Wiegand J, Hunter E, Dulvy NK (2011) Are spatial closures better than size limits for halting the decline of the North Sea thornback ray, *Raja clavata*? Mar Freshw Res 62: 722–733.
- Heppell SS, Crowder LB, Menzel TR (1999) Life table analysis of long-lived marine species with implications for conservation and management. Am Fish Soc Symp 23: 137–148.
- Kinney MJ, Simpfendorfer CA (2009) Reassessing the value of nursery areas to shark conservation and management. Conserv Lett 2: 53–60.

- Heppell SS, Caswell H, Crowder LB (2000) Life histories and elasticity patterns: perturbation analysis for species with minimal demographic data. Ecology 81: 654–665.
- 62. Cortés E (2007) Chondrichthyan demographic modelling: an essay on its use, abuse and future. Mar Freshw Res 58: 4–6.
- Smith SE, Au DW, Show C (1998) Intrinsic rebound potentials of 26 species of Pacific sharks. Mar Freshw Res 49: 663

 –678.
- 64. Dulvy N, Forrest R (2010) Life histories, population dynamics, and extinction risks in chondrichthyans. In: Carrier JC, Musick JA, Heithaus MR, editors. Sharks and Their Relatives II: biodiversity, adaptive physiology, and conservation. Boca Raton: CRC Press. 639–679.
- Forrest RE, Walters CJ (2009) Estimating thresholds to optimal harvest rate for long-lived, low-fecundity sharks accounting for selectivity and density dependence in recruitment. Can J Fish Aquat Sci 66: 2062–2080.
- Dulvy NK, Reynolds JD (2002) Predicting extinction vulnerability in skates. Conserv biol 16: 440–450.
- Morato T, Watson R, Pitcher TJ, Pauly D (2006) Fishing down the deep. Fish Fish 7: 24–34.
- Juan-Jordá MJ, Mosqueira I, Freire J, Dulvy NK (2012) Life in 3-D: life history strategies in tunas, mackerels and bonitos. Rev Fish Biol Fish: 1–21.
- Stevens JD, Bonfil R, Dulvy NK, Walker PA (2000) The effects of fishing on sharks, rays, and chimeras (chondrichthyans), and the implications for marine ecosystems. ICES J Mar Sci 57: 476

 –494.
- Hutchings JA, Myers RA, García VB, Lucifora LO, Kuparinen A (2012) Lifehistory correlates of extinction risk and recovery potential. Ecol Appl 22: 1061– 1067.
- Charnov EL, D Berrigan (1990) Dimensionless numbers and life history evolution: age of maturity versus the adult lifespan. Evol Ecol 4: 273–275.
- Charnov EL (2002) Reproductive effort, offspring size and benefit—cost ratios in the classification of life histories. Evol Ecol Research 4: 749–758.
- 73. Charnov EL, Gislason H, Pope JG (2012) Evolutionary assembly rules for fish life histories. Fish Fish.
- Gemmill Skorping, Read (1999) Optimal timing of first reproduction in parasitic nematodes. J Evol Biol 12: 1148–1156.
- Brickle P, Laptikhovsky V, Pompert J, Bishop A (2003) Ontogenetic changes in the feeding habits and dietary overlap between three abundant rajid species on the Falkland Islands' shelf. J Mar Biol Assoc U K 83: 1119–1125.
- Kotwicki S, Weinberg KL (2005) Estimating capture probability of a survey bottom trawl for Bering Sea skates (*Bathyraja* spp.) and other fish. Alsk Fish Res Bull 11: 135–145.
- Price B, Gearhart J (2011) Evaluations of turtle excluder device (TED) performance in U.S. southeast Atlantic and Gulf of Mexico skimmer trawl fisheries. US Department of Commerce, NOAA Technical Memorandum. NMFS-SEFSC-615.
- Brewer D, Rawlinson N, Eayrs S, Burridge C (1998) An assessment of bycatch reduction devices in a tropical Australian prawn trawl fishery. Fish Res 36: 195–215.
- Au DW, Smith SE, Show C (2008) Shark productivity and reproductive protection, and a comparison with teleosts. In: Camhi MD, Pikitch EK, Babcock EA, editors. Sharks of the open ocean: biology, fisheries and conservation. Oxford, UK: Wiley-Blackwell. 298–308.
- Parent S, Pépin S, Genet J-P, Misserey L, Rojas S (2008) Captive breeding of the barndoor skate (*Dipturus laevis*) at the Montreal Biodome, with comparison notes on two other captive-bred skate species. Zoo Biol 27: 145–153.
- Holden MJ (1974) Problems in the rational exploitation of elasmobranch populations and some suggested solutions. In: Harden-Jones FR, editor. Sea Fish Res: Logos Press, London. 117–138.
- 82. Holden MJ (1975) The fecundity of *Raja clavata* in British waters. J Cons 36: 110–118.
- Holden MJ, Rout DW, Humphreys CN (1971) The rate of egg laying by three species of ray. J Cons 33: 335–339.
- Ishihara H, Mochizuki T, Homma K, Taniuchi T (1997) Reproductive strategy
 of the Japanese common skate (spiny rasp skate) Okamejei kenojei. In: Fowler SL,
 Reed TM, Dipper FA, editors; Sabah, Malaysia. IUCN SSC Shark Specialist
 Group. IUCN. 236–240.
- Ryland JS, Ajayi TO (1984) Growth and population dynamics of three Raja species (Batoidei) in Carmarthen Bay, British Isles. J Cons 41: 111–120.
- Luer CA, Gilbert PW (1985) Mating behavior, egg deposition, incubation period, and hatching in the clearnose skate, *Raja eglanteria*. Environ Biol Fishes 13: 161–171.
- Holden MJ (1972) The growth rates of Raja brachyura, R. clavata and R. montagui as determined from tagging data. J Cons 34: 161–168.
- 88. du Buit MH (1976) The ovarian cycle of the cuckoo ray, *Raja naevus* (Muller and Henle), in the Celtic Sea. J Fish Biol 8: 199–207.
- Frisk MG, Miller TJ (2006) Age, growth, and latitudinal patterns of two Rajidae species in the northwestern Atlantic: little skate (*Leucoraja erinacea*) and winter skate (*Leucoraja ocellata*). Can J Fish Aquat Sci 63: 1078–1091.
- Miller DJ, Lea RN (1972) Guide to the coastal marine fishes of California. Calif Dep Fish Game Fish Bull 157.
- Gallagher MJ, Nolan CP, Jeal F (2005) Age, growth and maturity of the commercial ray species from the Irish Sea. J Northwest Atl Fish Sci 35: 47–66.
- Eschmeyer WN, Herald ES (1983) A field guide to Pacific Coast fishes of North America: from the Gulf of Alaska to Baja, California. Boston: Houghton Mifflin.

- Mecklenburg CW, Mecklenburg TA, Thorsteinson LK (2002) The fishes of Alaska. Bethesda: American Fisheries Society.
- McEachran JD, Musick JA (1975) Distribution and relative abundance of seven species of skates (Pisces: Rajidae) which occur between Nova Scotia and Cape Hatteras. Fish Bull 73: 110–136.
- 95. Bigelow HB, Schroeder WC (1953) Sawfishes, guitarfishes, skates and rays. In: Fishes of the western North Atlantic. 1–514.
- 96. Kulka DW, Frank KT, Simon JE (2002) Barndoor skate in the northwest Atlantic off Canada: distribution in relation to temperature and depth based on commercial fisheries data. Department of Fisheries and Oceans, Canadian Science Advisory Secretariat Research Document 2002/073.
- Gedamke T, DuPaul WD, Musick JA (2005) Observations on the life history of the barndoor skate, *Dipturus laevis*, on Georges Bank (western North Atlantic). J Northwest Atl Fish Sci 35: 67–78.
- Kulka DW, Miri CM (2003) The status of thorny skate (Amblyraja radiata Donovan, 1808) in NAFO divisions 3L, 3N, 3O, and subdivision 3Ps. NAFO SCR. 03/57(N4875). 100 p.
- Sulikowski JA, Kneebone J, Elzey S, Jurek J, Danley PD, et al. (2005) Age and growth estimates of the thorny skate (Amblyraja radiala) in the western Gulf of Maine. Fish Bull 103: 161–168.

- 100. Sulikowski JA, Kneebone J, Elzey S, Jurek J, Howell WH, et al. (2006) Using the composite variables of reproductive morphology, histology and steroid hormones to determine age and size at sexual maturity for the thorny skate Amblyraja radiata in the western Gulf of Maine. J Fish Biol 69: 1449–1465.
- 101. Ishihara H, Ishiyama R (1985) Two new North Pacific skates (Rajidae) and a revised key to Bathyraja in the area. Jpn J Ichthyol 32: 143–179.
- Ebert DA (2003) Sharks, rays and chimaeras of California. Berkeley and Los Angeles: University of California Press. 284 p.
- Hunter E, Buckley AA, Stewart C, Metcalfe JD (2005) Migratory behaviour of the thornback ray, *Raja clavata*, in the southern North Sea. J Mar Biol Assoc U K 85: 1095–1105.
- Mytilineou C, Politou C-Y, Papaconstantinou C, Kavadas S, D'Onghia G, et al. (2005) Deep-water fish fauna in the Eastern Ionian Sea. Belg J Zool 135: 229–233.
- 105. Chevolot M, Ellis JR, Hoarau G, Rijnsdorp AD, Stam WT, et al. (2006) Population structure of the thornback ray (*Raja clavata* L.) in British waters. J Sea Res 56: 305–316.